# Sarasota Bay Juvenile Fisheries Habitat Assessment 



Protecting Our Water Heritage

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## EXECUTIVE SUMMARY

Critical habitat for juvenile estuarine fishery species within the Sarasota Bay system was characterized in relation to a bay segmentation scheme using an integrated approach for mapping of shore type, emergent wetland and submerged habitat features. Habitat feature classes were tabulated in order to derive length or area-based indices (ratios) which summarize the proportion of individual habitat features as a function of total shoreline length or water area within each bay segment. These critical habitat indices provide a means to rank the spatial distribution of critical habitats and to evaluate restoration strategies in relation to goals for establishing a complex of adjacent complementary habitats (i.e., habitat mosaics) throughout the Sarasota Bay system. Knowledge of the proportion of critical habitats can be used to prioritize future restoration projects in order to maximize the multiple habitat requirements of fishery species.

The shoreline mapping revealed that bulkheads presently represent 185.8 linear miles $(45.3 \%)$ of the shoreline of the SBNEP study area, with riprap representing 38.5 miles $(9.4 \%)$. These generally unvegetated shoreline areas offer great opportunities for future shoreline enhancement. A total of approximately 7.6 miles ( $1.8 \%$ ) of the shoreline is also dominated by nuisance and/or exotic species. Again, these areas offer easy opportunities for enhancement.

The Bays \& Estuaries habitat type was the dominant habitat type within the SBNEP study area, comprising approximately $91 \%$ of the area. The Total Intertidal Area represents those habitat types within the bay that are not considered Bays and Estuaries, and includes Beaches Other than Swimming, Mangrove Swamps, Oyster Bars, Salt Barren, Saltwater Marshes and Unvegetated Bottoms. The Total Intertidal Area varied from a low of $0.5 \%$ of a bay segment to a high of $32.5 \%$ of a bay segment. Bay segments 11 ( $0.5 \%$ ), 8 ( $1.3 \%$ ), 7 (2.9\%) and 10 ( $2.9 \%$ ) represent those bay segments with the least Total Intertidal Area. Conversely, bay segment 2 ( $32.5 \%$ ), 15 ( $32.4 \%$ ), and 9 ( $25.1 \%$ ) are the bay segments with the highest percentage of Total Intertidal Area. If possible, habitat restoration activities
should be directed toward those bay segments with the lowest percentage of Total Intertidal Area. However, the Total Intertidal Area ratio should not be the most significant factor considered when prioritizing restoration areas since it is partially influenced by the bay segmentation scheme itself.

In addition to structural (stationary) habitat, conditional (dynamic) habitat was characterized according to its spatial distribution across the Sarasota Bay system. Salinity regime within bay segments and tributary streams was tabulated and used as an indicator to identify those areas classified as oligohaline or mesohaline ( $<19 \mathrm{ppt}$ ) zones which are utilized by many commercially or recreationally important fishery species during portions of their early life history. Zones exhibiting a oligohaline-mesohaline salinity regime are restricted to tidally inundated portions in lower reaches of tributaries many of which are urbanized and devoid of critical habitat. The priority for future habitat restoration should focus on establishing critical habitats within these salinity zones, considered optimal as fishery nursery habitat.

Fishery communities found within these stationary and dynamic habitats throughout the Sarasota Bay system were characterized during summer and winter seasons in order to identify relationships between species abundance and occurrence within natural versus restored habitats and within microhabitats like marsh edges, channels, subtidal pools or seagrass beds. Knowledge of the species or groups (i.e., guilds) which utilize natural and restored sites and their affinities for microhabitat features can be used to enhance the design of future restoration projects in order to maximize fishery productivity.

Fishery sampling during 2002 revealed a fishery community exhibiting high dominance of bait and forage fishes which comprise a group, or guild, of approximately eleven resident species. Selected taxa, those species considered commercially or recreationally important, exhibited low abundance and typically comprised less than 10 percent of the total catch when averaged across all collections. Two species (i.e., Spot, Leiostomus xanthurus, and Striped Mullet, Mugil cephalus) dominated the catch of selected taxa and were frequently
caught while another seventeen selected taxa were collected in smaller numbers and occurred less frequently in collections.

Overall, the selected taxa were evenly represented at both natural and restored sites. When reviewing the selected taxa data by guilds, the natural sites were observed to have a higher mean density for both resident and transient species, however the restored sites had a slightly higher mean density of nursery species. The monitoring results indicate that the restored microhabitat types are faring fairly well when compared to that of the natural habitat types. Other than the density of transient species, the species composition and density of the restored sites is comparable to and in some cases better than that for the natural sites. However, the natural sites do appear to be more diverse and there may be a lag affect related to species richness. Based upon the varied locations and microhabitat types where selected species were captured, it is recommended that the restoration sites continue to be created as mosaics with various microhabitat types.

Three artificial reef sites were also characterized during the summer and winter seasons in order to identify nursery habitat provided by the artificial reefs. A total of 59,027 individuals represented by 27 different species or species groups were observed during the sampling events. Early juveniles of sardines (Clupeidae spp.), tomtate (Haemulon aurolineatum) and pinfish (Lagodon rhomboides) were the dominant species observed and exhibited seasonality in their occurrence. Many of the reef dependent species such as tomtate, gray snapper (Lutjanus griseus) and gag (Mycteroperca microlepis) were observed in much higher densities on the artificial reefs than in the seine samples on other habitat types. This also held true for the commercially important stone crab (Menippe mercenaria). The complexity of the material as well as the length of time since deployment appeared to be factors in the diversity of species that were observed. The artificial reef materials in the SBNEP study area provide a unique habitat type for a variety of reef dependent species and continued deployment of reef material is recommended. Enhancement of existing reefs and the placement of additional material to improve complexity would also enhance juvenile fish habitat qualities.

The habitat restoration working group of the SBNEP recommended the creation of a fiveyear restoration plan document to provide clear direction to the SBENP program in terms of habitat restoration. This document would be used to develop and improve restoration targets and may assist in obtaining additional funding. As a precursor to the development of this plan, a summary report of recommendations from the Sarasota Bay technical community would be invaluable.

### 1.0 INTRODUCTION

Sarasota Bay (Bay) was named in the Water Quality Act of 1987 as an estuary of national significance. Following a nomination process, the Sarasota Bay National Estuary Program (SBNEP) was formally established in June 1989 by the U.S. Environmental Protection Agency. The legislation required that federally appropriated funds be used for technical projects to characterize Bay quality for the purpose of formulating specific action plans to restore and protect the Bay. In January 1993, the results of technical investigations and policy deliberations were formulated into a Framework for Action report. After two years of debate, the Comprehensive Conservation and Management Plan (CCMP) evolved and included six action plans related to wastewater, stormwater, wetlands, fisheries, recreation and governance. Additional research and education projects to refine the CCMP and promote implementation have also occurred.

The CCMP includes a goal to restore 18 acres of tidal wetlands and 11 acres of freshwater wetlands annually. Since 1990, the community has restored over 170 acres of estuarine tidal wetlands with continuing projects designed to restore additional wetland acreage. The CCMP also promoted increasing fishery productivity through increased seagrass coverage (from nitrogen reduction and other water quality improvements) and artificial reefs. In 1997, an artificial reef master plan was approved to increase fish habitat using "reef balls" or other artificial reef modules. The CCMP also recommended that a technical project be conducted to monitor and assess the results of artificial reef construction related to potential benefits for fish and shellfish productivity within the Bay. The present study was conceived to address the status of fishery utilization of natural and restored wetland habitats and artificial reefs, as well as to inventory the critical nursery habitats that are important for sustaining adult fish populations. Ultimately, the results of this investigation can be used to guide decisions for improving local habitat restoration efforts.

Previously, the SBNEP conducted characterization studies about Sarasota Bay fish and fisheries (Edwards, 1992a). Similarly, characterization studies were focused on tidal
wetlands (Estevez, 1992) and bay bottom habitats (e.g., seagrasses) (Culter, 1992). Emerging from these investigations, various research needs were identified, including:

- How do tidal marshes compare to mangroves in terms of SBNEP goals for water quality and habitat?
- Should wetlands created for fish and invertebrates be designed differently from wetlands created for birds, or for other wetland values?
- By focusing on habitat types that are extensive and common throughout the Bay, ecological relationships can be researched as a priority concern.
- Is the classification of critical habitats according to location, form and species composition paralleled by distinct feeding preferences of fish within these habitats?
- What faunal guilds should be targeted for guiding habitat restoration, preservation and management decisions for the Bay?

As reported by Paperno, Mille and Kadison (2001), the highly productive nature of estuarine habitats and their importance to fish in many life history stages are well documented (Miller et al., 1985; Szedlmayer and Able, 1996; Wantiez et al., 1996; Able and Fahay, 1998; and others). Edwards and Hueter (1992) and Edwards (1992b) observed that many important estuarine fishes have a life history starting with spawning in open coastal waters, followed by movement of planktonic larval stages to very shallow estuarine habitats wherein larvae reach metamorphosis at an age of 3 to 4 weeks. In order for these newly-metamorphosed juveniles (early-juvenile) to survive, they must find specific nursery habitats. For such species, the entire recruitment or year-class strength may be determined by availability of suitable early-juvenile habitat (Edwards, 1989).

In this context, the nature of suitable habitat may be exceedingly complex, encompassing both structural (stationary) components (e.g., vegetative form, bank slope, sediment composition, etc.) and conditional (dynamic) components (e.g., freshwater flow, salinity gradient, etc.). Browder and Moore (1981) developed this idea into a conceptual model in which production (recruitment) occurs at the intersection of stationary and dynamic habitat.

In other words, only that portion of structural habitat located within a suitable salinity range is highly productive in terms of fishery recruitment. As Edwards (1992b) has observed, if sufficient nursery habitat is not maintained, a biological bottleneck will exist; and management actions may not be effective in maintaining fishery resources. Therefore, the characterization of critical nursery habitat should be given high priority in improvement and management programs for the Bay.

Artificial reefs are constructed primarily for the purpose of providing recreational sports fishing opportunities but also serve to increase hard bottom habitat and the abundance of reef dependent species. There are presently 14 artificial reefs within the SBNEP study area, including five sites that have been constructed within the last year. Artificial reef construction includes material opportunity such as concrete blocks, piling, ceramic insulators and limestone boulders, as well as prefabricated reef ball modules and PVC units. Due to permitting requirements, the artificial reef sites are presently located within the northern two-thirds of the study area, from Roberts Bay in the south to Emerson Point to the north. Artificial reef construction continues to be well-funded and future deployment of new reefs and enhancement of existing reefs is expected to continue.

To date, there has been no quantitative assessment of the artificial reefs in the study area or an analysis of the fish communities and differences between these reefs. As reef deployment continues, opportunities to improve the value of these reefs to various life stages of commercially and recreational important species (selected taxa) are available. The objective of this study is to determine:

- What species, at what life stages and at what density presently use the artificial reef system within the SBNEP study area?
- What characteristics are the most productive, (i.e. what material types, complexity, location, etc.)?
- What are the temporal and spatial differences of selected taxa use of these artificial reefs?
- Can the productivity of existing reefs be enhanced?

The overall objective is to provide recommendations for the proper siting and constructing of artificial reefs, or enhancement of the existing reefs, to improve the nursery habitat value for commercially and recreational important fish species.

### 2.0 METHODS

### 2.1 Study Area

Located in the east-central region of the Gulf of Mexico, the Sarasota Bay study area spans approximately 52 square miles of open water in Manatee and Sarasota Counties along the west-central coast of peninsular Florida. It encompasses subtropical, tidal lagoons and a diverse assemblage of emergent and submerged estuarine habitats, including: mangrove forests, salt marshes, salterns, seagrass beds, unvegetated tidal flats, oyster bars and hard bottom, which are separated from gulf waters by a chain of barrier island landforms and tidal inlet channels. Mangrove forests are composed primarily of red (Rhizophora mangle), black (Avicennia germinans), and white (Laguncularia racemosa) mangroves with buttonwood (Conocarpus erectus) as a common associate. Salt marshes are dominated by cordgrass (Spartina alterniflora and Spartina patens) and black needlerush (Juncus roemerianus) together with leather fern (Acrostichum spp.). Salterns, or salt barrens, occur as high marsh environments with irregular tidal inundation and hypersaline soil conditions sparsely vegetated by saltwort (Batis maritima) and glasswort (Salicornia spp.) together with saltgrass (Distichlis spicata) and jointgrass (Paspalum vaginatum). Invasive, nuisance vegetation comprised primarily of Brazilian pepper (Schinus terebinthifolius) and Australian pine (Casuarina equisetifolia) displace emergent wetland vegetation along upland fringes. Seagrass meadows are dominated by shoal grass (Halodule wrightii), turtle grass (Thalassia testudinum), manatee grass (Syringodium filiforme) together with a diverse drift algae community. Oyster reefs are common along shoreline fringes and coquina rock outcrops occur sporadically.

Geographically, the Sarasota Bay study area includes several named subregions, including Anna Maria Sound, Palma Sola Bay, Sarasota Bay, Roberts Bay, Little Sarasota Bay, Dryman Bay and Blackburn Bay, together with several tidal creek basins that occur between the mouth of Tampa Bay on the north and Venice Inlet on the south. The Sarasota Bay study area is urbanized. Since the early 1900s, dredge and fill activities have resulted in significant losses of the ecologically important habitats, and historically semi-isolated tidal current regimes within embayments have been joined through the construction of an Intra3coastal Waterway (ICW) channel that hydrologically links all embayments. Mosquito ditching further exacerbated the problem of wetland habitat loss. Salinity of bay waters is determined by tides and runoff, salinity ranges predominate in euhaline ( $23-35 \mathrm{ppt}$ ) or polyhaline ( $15-28 \mathrm{ppt}$ ) classes (Bulger et al. 1990). The drainage basin of the Sarasota Bay study area encompasses a land area of approximately 280 square miles and the headwaters of several tidal creeks which flow relatively short distances into the Bay exhibit significant channelization which has altered historical patterns of surface water runoff to the Bay.

A variety of tidal wetland restoration projects have been constructed since the 1980s throughout the Bay. Most projects exist adjacent to public-owned uplands and are dispersed throughout the Sarasota Bay study area where mangrove-marsh platforms linked by tidal channels and subtidal pools have been built. In addition, submerged lands have been enhanced through construction of artificial reefs using a variety of recycled materials and constructed reef modules.

### 2.2 Habitat Mapping \& GIS Database

Separate protocols were developed for mapping and GIS database development for a) shore type features, b) tidal wetlands features, and c) seagrass features within the study area according to a bay segmentation scheme (Figure 2.2-1) previously adopted by the SBNEP for its technical projects. A detailed photointerpretation key (Sauers, 2001) was assembled to guide the work process of mapping and GIS database development. Dates of aerial photography encompassed 1998, 1999 and 2001. Imagery was derived from source photography at scales between $1: 8,000$ and $1: 24,000$ and included true color diapositives,

digital color-infrared orthographic quarter quadrangles, and digital black and white panchromatic sectional images as obtained from the Southwest Florida Water Management District (SWFWMD), Manatee and Sarasota County Governments. Digital mosaics according to Section-Township-Range boundaries of color-infrared imagery were compiled by Sarasota County Government from digital orthographic quarter quadrangles obtained from SWFWMD with 1-meter pixel resolution. Black and white imagery as digital mosaics according to Section-Township-Range boundaries were compiled by and obtained from Manatee and Sarasota County Governments with 1-foot pixel resolution.

Shore type features were mapped along an arc representing the apparent shoreline (i.e., landwater interface) via photointerpretation and groundtruthing according to a cover classification system including geomorphic, shore morphology and intertidal vegetation tiers, with from six to ten feature types (classes) in each tier. A custom extension in the ArcView v.3.2a (ESRI 1996) software environment was developed by VHB to implement route system and dynamic segmentation functionality and to allow each distinct shore type feature class to be mapped as an "event" along a shoreline arc or "route". Technicians simultaneously viewed digital imagery and mapped shore type features in the ArcView software environment. Polyline shoreline features larger than 5-meters in length were mapped. Delineation and coding corrections or addition of new polyline features were made on the basis of field observations during groundtruthing. Query analyses were performed to tabulate and chart the proportion of feature classes within the entire study area and within each bay segment zone.

Tidal flats features were mapped via photointerpretation and groundtruthing according to a project-specific, refined Florida Land Use, Cover and Forms Classification SystemFLUCFCS (Florida Department of Transportation, 1999) adapted from SWFWMD feature classes used in its 1999 land cover map and GIS database. Unique FLUCFCS codes for salt barrens, oyster bars and unvegetated bottoms were added to existing SWFWMD tidal wetlands feature classes and mapped according to the SBNEP bay segmentation scheme using a geometric intersect operation in the ARC/INFO (ESRI 1999) software environment.

Technicians simultaneously viewed digital imagery and mapped tidal flats features in the ArcView software environment. Polygon tidal flats features larger than 20-meters in length of major axis were mapped. Delineation and coding corrections or addition of new polygon features were made on the basis of field observations during groundtruthing. Query analyses were performed to tabulate and chart the proportion of feature classes within the entire study area and within each bay segment zone.

Seagrass features were derived from the 1999 SWFWMD feature classes used in its 1999 seagrasses land cover map and GIS database. No additional photointerpretation or groundtruthing was applied to the map. The existing SWFWMD seagrass feature classes were mapped according to the SBNEP bay segmentation scheme using a geometric intersect operation in the ArcINFO (ESRI 1999) software environment. According to methodology adopted by SWFWMD, polygon seagrass features greater than 0.5 acre were mapped. No delineation or coding corrections or were made to seagrass or tidal flats feature classes; however, new polygon features were added in tributary areas (i.e., bays and estuaries feature class only) to delimit the same study area as provided for the tidal wetlands GIS database. Query analyses were performed to tabulate and chart the proportion of feature classes within the entire study area and within each bay segment zone.

Vanasse Hangen Brustlin, Inc. produced final GIS products for the SBNEP. These products included ArcView data files and selected paper map printouts at small scale sufficient to reveal the bay segmentation scheme and interpret summary Tables and charts. Electronic files, tabular and chart data repose with the City of Sarasota, Sarasota Bay National Estuary Program.

### 2.3 Critical Natural \& Restored Habitat Assessment

## Sampling

Fishery utilization of natural and restored shore-fringing habitats by juvenile fishery species within the SBNEP study area was characterized by using a fixed-station sampling strategy. Sampling stations were chosen with two goals in mind: 1) to optimize geographic coverage


across SBNEP-sponsored habitat restoration sites and 2) to detect differences in faunal assemblages that are related to local habitat characteristics. A preliminary inventory was made in August 2001 of microhabitat features at ten completed habitat restoration sites located within 5 of 16 bay segments throughout the study area. A similar inventory was made of microhabitat features at existing natural habitat sites in the study area. Multiple sampling stations were selected within sites for distinct geographic locations, and natural site stations were chosen which mimic the habitats occurring among the restoration site stations in order to serve as "controls" for those samples taken from restored sites. Sample stations were located within or directly adjacent to the targeted microhabitat features observed within all sites. Figures 2.3-1 and 2.3-2 depict the sampling station locations and Table 2.3-1 generally describes the sampling stations by location and habitat features.

A total of forty fixed seine stations were sampled by collecting two consecutive hauls during a five-day period during each of two seasonal periods (February and June) corresponding to peak recruitment of selected taxa considered commercially or recreationally important. Eighteen fixed seine stations were established in restored sites, and twenty-two fixed seine stations were established in natural sites (see Figures 2.3-1 and 2.3-2 and Table 2.3-1). Both 21.3 -meter and 6.1 -meter center bag seines with 3.2 mm delta mesh were deployed in various techniques along shorelines where water depths were less than 1.8-meters. A 21.3meter seine was deployed in one of two different ways: a) seine gently hauled adjacent to or away from the shoreline with wings held 15.5 -meters apart for a distance of 9.1 -meters across an area of about $140 \mathrm{~m}^{2}$ (standard-offshore sets); b) seine gently deployed from boat along steeply-sloped shorelines in a semicircular pattern (arc with radius 6.7-meters) and the area swept was $68 \mathrm{~m}^{2}$. A smaller 6.1 -meter seine was deployed in one of three different ways: a) seine gently hauled adjacent to or away from the shoreline with wings held 4.3meters apart for a distance of 9.1-meters across an area of $39 \mathrm{~m}^{2}$; b) seine gently deployed along shoreline in a semicircular pattern (arc with radius 2.0 -meters) and the area swept was $3 \mathrm{~m}^{2}$; and c) seine gently hauled along full width of narrow channel for some fixed distance and the area swept varied among three stations between $23 \mathrm{~m}^{2}$ and $36 \mathrm{~m}^{2}$.

Table 2.3-1. Index of fishery sampling stations and sample groups in Sarasota Bay for SBNEP fishery habitat assessment during February and June 2002. Target habitats reflect features located within and adjacent to seine stations. Microhabitats reflect features directly sampled by gear deployment techniques.

| NATURAL HABITAT FISHERY SAMPLING STATIONS |  |  | SAMPLED |
| :---: | :---: | :---: | :---: |
| STATIONS | LOCATION CODES | TARGET HABITATS |  |
| C1 | Tidy Island--TI | Tidal Channel, Salt Flat | Seagrass |
| C2 | Tidy Island--TI | Vegetated Low Marsh (Spartina) | Marsh Edge |
| C3 | Tidy Island--TI | Beach, Mudflat, Seagrass Flat | Marsh Edge |
| C4 | Leffis Key--SLK | Beach, Seagrass Flat (Thalassia) | Seagrass |
| C5 | Sister Keys--SLK | Subtidal Pool, Mangrove Fringe | Subtidal Pool |
| C6 | Sister Keys--SLK | Tidal Lagoon, Mangrove Fringe | Channel |
| C7 | Sister Keys--SLK | Seagrass Flat (Halodule) | Seagrass |
| C8 | El Conquistador--EBC | Vegetated Low Marsh (Juncus) | Channel |
| C9 | El Conquistador--EBC | Tidal Lagoon (Mangrove), Oyster Bar | Seagrass |
| C10 | Bowlees Creek--EBC | Oyster Bar, Seagrass Flat | Subtidal Pool |
| C11 | Bowlees Creek--EBC | Tidal Lagoon, Mangrove Fringe | Channel |
| C12 | Bowlees Creek--EBC | Beach, Seagrass Flat | Seagrass |
| C13 | South Lido Park--SLP | Seagrass Flat (Thalassia), Beach | Seagrass |
| C14 | South Lido Park--SLP | Tidal Lagoon (Mangrove), Mudflat | Seagrass |
| C15 | Phillippi Creek--PC | Riverine Channel, Low Marsh (Typha) | Marsh Edge |
| C16 | Phillippi Creek--PC | Oyster Bar, Subtidal Pool | Subtidal Pool |
| C17 | Heron Lagoon--HL | Tidal Lagoon, Subtidal Pool | Marsh Edge |
| C18 | Heron Lagoon--HL | Tidal Lagoon, Mangrove Fringe | Seagrass |
| C19 | North/Catfish Creeks--MNP | Tidal Lagoon (Mangrove), Oyster Bar | Subtidal Pool |
| C20 | MNP/ Bird Keys--MNP | Seagrass Flat, High Marsh | Seagrass |
| C21 | South Creek--OSP | Riverine Channel (Leather Fern) | Channel |
| C22 | South Creek--OSP | Vegetated Low Marsh (Juncus) | Marsh Edge |
| RESTORED HABITAT FISHERY SAMPLING STATIONS |  |  | SAMPLED <br> MICROHABITATS |
| STATIONS | LOCATION CODES | TARGET HABITATS |  |
| R1 | Leffis Key--LK | Seagrass Flat, Tidal Lagoon | Seagrass |
| R2 | Leffis Key--LK | Tidal Channel, Mudflat | Channel |
| R3 | Leffis Key--LK | Subtidal Pool, Mangrove Fringe | Subtidal Pool |
| R4 | Durante Park--DP | Vegetated Low Marsh (Juncus) | Marsh Edge |
| R5 | Durante Park--DP | Tidal Channel, Mudflat | Channel |
| R6 | Durante Park--DP | Seagrass Flat, Beach | Seagrass |
| R7 | Sarasota Baywalk--SBW | Salt Flat, Transitional High Marsh | Subtidal Pool |
| R8 | Sarasota Baywalk--SBW | Vegetated Low Marsh, Subtidal Pool | Marsh Edge |
| R9 | Sarasota Baywalk--SBW | Mudflat, Mangrove Fringe | Seagrass |
| R10 | Sixth Street Canal--SSC | Mudflat, Mangrove Fringe, Oyster Bar | Subtidal Pool |
| R11 | Sixth Street Canal--SSC | Tidal Lagoon, Seagrass Flat | Marsh Edge |
| R12 | Hog Creek--HC | Low Marsh (Typha), Subtidal Pool | Marsh Edge |
| R13 | Hog Creek--HC | Tidal Lagoon, Mangrove Fringe | Channel |
| R14 | Selby Gardens Pond--SGP | Vegetated Low Marsh (Juncus) | Subtidal Pool |
| R15 | Selby Shoreline--SS | Beach, Seagrass Flat | Seagrass |
| R16 | Palmer Point Park--LSB | Transitional High Marsh, Salt Flat | Marsh Edge |
| R17 | MNP/Bird Keys--LSB | Salt Flat, Seagrass Flat | Seagrass |
| R18 | Hidden Bay Condos--LSB | Vegetated Low Marsh, Tidal Lagoon | Channel |

A standard-offshore technique was employed with both seines in open waters or near shoreline margins. A boat-river technique was employed with both seines along steeply sloped shorelines or shorelines with extensive overhanging vegetation. In narrow channels or small subtidal pools, a non-standard technique of a straight line tow of variable distance was employed. A detailed description of each sampling technique is provided in the FMRI FIM protocols.

All samples were collected during daylight hours, some near dawn or dusk periods, and were processed in the field. All fish and selected shellfish or invertebrates captured in the gear were identified in the field to the lowest practical taxon, measured to the nearest millimeter and enumerated; those fish positively identified were then released. From each sample, up to 10 individuals of each fish and invertebrate taxon were randomly culled, measured and their lengths recorded, except up to 40 fish of selected taxa (commercially or recreationally important) were measured and remaining individuals counted. Unidentified specimens or other representative specimens were retained for laboratory verification of taxon identification or confirmation of field identification. Some taxa were not identified to species (Brevoortia spp., Menidia spp.) because of the possibility of hybridization in the study area or because they were not morphologically or meristically distinguishable at small juvenile sizes (Eucinostomus spp. $<40 \mathrm{~mm}$ SL, Gobiosoma spp. $<20 \mathrm{~mm}$ SL, or Strongylura spp. [marina and timиси only] < 100 mm SL). Very abundant collections were subsampled using a Motoda Box and processed after large, rare or selected taxa were sorted and separately processed from the sample. Physical data including water temperature ( ${ }^{\circ} \mathrm{C}$ ), salinity (\%), and conductivity (mhos) were recorded with each sample by using an YSI SCT Meter or YSI 610D/600 Sonde.

Species were assigned to one of three guild categories (i.e., nursery, transient, or resident) based on similar exploitation of resources available within sampled habitats or to group species that demonstrate similar functional use of habitats as documented in scientific literature of life history profiles. Nursery species were grouped due to their cyclical or seasonal use of habitats as juveniles before migrating offshore as adults, and selected species
(i.e., commercial or recreational importance) were given priority for assignment to this guild. Transient species were grouped due to their regular migrations between bay habitats on a diel or seasonal basis, occurring as occasional or accidental visitors to the estuary, and forage or spawn in lower reaches of the estuary. Resident species were grouped due to their restricted distribution to one or few habitats and spend their entire life cycle, through reproduction and growth to maturity, within the confines of the estuary and seldom, if ever, leave the estuary. Guild assignments were used in the analysis of community structure.

Various characteristics of microhabitat including bottom type, bottom vegetation type, shore type, inundation or overhang condition of shore type, bycatch type and quantity were recorded with each sample together with percent cover of the type classes. Additional characteristics of the bank slope and its distance from the sample were also recorded with each sample. Characteristics of the gear, its wing depth, the seine center-bag depth and the direction towed in relation to current or wind was also recorded with each sample. The coding of gear deployment technique, sample stratum, bottom vegetation and habitat characteristics were used to assign samples to specific microhabitat groups for data analysis and statistical treatment.

## Data Analysis

All records from field datasheets were entered into a relational database customized by FMRI-FIM in the MSAccess software environment. The database was extracted to appropriate grid files in MSExcel software environment format and statistical and community analysis software programs were used to compute parameters.

Fish size was measured and reported as standard length (SL) in millimeters. In addition to the up to 40 randomly culled specimens of selected taxa (up to 10 specimens for nonselected taxa), unmeasured specimens were proportionally adjusted for the length-frequency plots to reflect the size structure of the entire sample(s).

Due to variable catch efficiencies and selectivity of the gears used in this study, raw abundance data for each species were tabulated separately for each gear. In addition, raw abundance data for each species were converted to faunal density on the basis of catch per unit of $100 \mathrm{~m}^{2}$ area (CPUA) for each sample collected. Faunal density-CPUA values for each species were then used as individual count data for pooled seasonal or sample groups density computations. In order to help normalize faunal density data from contagiously distributed organisms, all individual CPUA data were $\log$-transformed $\left(\log _{e}(\mathrm{x}+1)\right)$ prior to similarity calculations, community structure analyses, and during various other statistical treatments.

Unidentified species complexes (i.e., Menidia spp., Anchoa spp., Clupeidae spp., Eucinostomus spp., Fundulus spp., Gobiosoma spp., Brevoortia spp., Mugil spp., and Strongylura spp.) were treated as a separate, single species for all analyses. With the above exceptions, all species were included in the calculation of descriptive indices of community structure recommended by Livingston (1976) or Florida Marine Research Institute (2000). These included species composition, number of individuals, frequency, percentage occurrence in samples, relative abundance (mean number of individuals per haul), faunal density (mean number of individuals, CPUA, per haul), species richness (mean number of species per haul), Shannon-Wiener diversity (rare species were not excluded). Index calculations were based on pooled seasonal data.

Station-related and location-related density differences were determined by all taxa, by selected taxa only, or by species groups according to guild assignment (nursery, resident, transient). Additional microhabitat type-related density differences between natural versus restored sites were determined by all taxa, by selected taxa only, or by species groups according to guild assignment (nursery, resident, transient).

Physicochemical parameters of surface water (i.e., temperature, conductivity and salinity) were measured and recorded during each trip to fishery sampling stations (i.e., 4 trips to each station over 2 seasonal periods). However, equipment malfunction resulted in missing
data values for some trips and stations during each seasonal period. Therefore, statistical analysis of water temperature $(\mathrm{n}=143)$ and salinity $(\mathrm{n}=130)$ differences at stations was made for pooled seasonal data. Data for physicochemical parameters were tabulated and analyzed using MSExcel2000 with the Analyse-it (v. 1.63) add-on software application. A non-parametric Kruskal-Wallis test was used for comparisons of temperature and salinity between stations. Null hypotheses were tested separately for temperature and salinity, as follows, $H_{0}$ : measured parameter values at all stations are the same; $H_{A}$ : measured parameter values at all stations are not the same. Statistical significance tests were made at the $5 \%$ significance level $(\alpha=0.05)$ by comparing the Kruskal-Wallis test statistic $\left(H_{C}\right)$ to the critical value of the Chi-square test statistic $\left(\chi^{2}\right)$ and for determining whether to reject, or accept, the null hypothesis, $H_{0}$.

### 2.4 Artificial Reef Habitat Assessment

The initial task under the artificial reef assessment was to review all existing information available on the construction of the artificial reefs. Existing data reviewed was from Sarasota County files and included information on the dates of material placement, the material type, quantity of material, any historic mapping and monitoring data, or any anecdotal evidence on the success of the reef deployment. In addition, available GPS coordinate data were obtained to locate all the reef material.

## Reef Characterization

Prior to sampling station selection, all of the existing reef sites were characterized. Reef characterization involved a reconnaissance-level survey of each reef site using SCUBA gear. The reef was characterized according to its structure, configuration and location. Survey data also identified material type (e.g., concrete blocks, reef balls, ceramic insulators, limestone boulders, PVC structures, etc.), approximate areal extent, height (or relief) of material and complexity. Complexity is a very important factor and is a measure of the number and variety of niches available on the reef for finfish. A scale of 1 to 5 was used to characterize complexity. A rank of one is lowest (representing a flat surface flush with a substrate), while a five is the highest rank (indicative of a reef with a wide variety of holes,
shapes, and sizes and vertical relief). In addition to the physical data collected from each reef site, the proximity of the site to tidal passes, tidal creeks and other estuarine habitats was measured from navigational charts.

The reconnaissance survey occurred during September of 2001 and coincided with an extreme red tide event in the bay system. This red tide devastated several of the reef sites, resulting in mortality of both invertebrate and finfish species. However, some sites were less impacted, with live specimens being observed. Due to the timing of this extreme red tide event, monitoring was postponed until February for the initial sampling period.

## Station Selection

Based upon the results of the reconnaissance-level survey, the Gerkin, Hart and Walker reef sites (Figure 2.3-1) were selected for monitoring. These sites were selected due to the variety of material types, the different ages of the deployed material, the variable proximity to the tidal passes, and because they all contained prefabricated reef ball units. These three reefs encompassed all of the material types that were deployed within the Sarasota Bay National Estuary Program study area.

## Survey Methodology

Each selected reef site was monitoring by point-census and linear transect methods. The point-census stations focused on small specific areas on each reef that contained a unique material type (e.g., reef ball, PVC reef, concrete block piles, etc.) Conversely, linear transects were of variable lengths and were established to repetitively survey large areas of specific material types.

Reef sites were specifically located and marked to allow subsequent repetitive sampling. This involved the selection and marking of each station location. Concrete blocks coated with white anti-fouling paint and black-stenciled site identification numbers marked the location of each point-census station and the end of each linear transect. The substrate surface area occupied by the reef material being surveyed was also measured using a
fiberglass tape to calculate fish density. Accurate navigational data for each station location was obtained using a GPS differential unit.

Three-point-census locations were established on the Hart and Walker reefs and four were established at the Gerkin site (a total of ten point-census locations between the three reef sites). The point-census survey involved a five-minute count of all species observed within one-meter of the point-census material during each event. Point-census locations also provided for replicate sampling of the same material type (e.g., reef ball) at each site during the same sampling event, as well as subsequent successive sampling events. The SCUBA diver surveyed the material for a distance of one-meter for one minute in each of the four quadrants around the material. During the final minute of observation, the observer fully investigated the material from close range to record cryptic species.

At least one linear transect was monitored at each reef site. A total of four linear transects locations were surveyed (two at the Walker site). A SCUBA diver slowly swam along the linear transect while recording the number and species of all individuals observed within one meter of the reef structure. The emphasis of this survey method was to provide counts of all selected taxa. Therefore, the diver thoroughly inspected all nooks and crannies of the reef structure while proceeding along the transect.

Survey results were recorded using an underwater slate. The slate was pre-labeled with common reef fish species names to allow for quick recording and to maximize the direct visual observation time. The slate results were subsequently transcribed to field data sheets. Observed individuals were placed into one of four size classes, early juvenile (EJ), juvenile (JUV), sub-adult (SA), and adult (A). All individuals were identified to the lowest practical taxonomic level and some of the early juveniles were captured and returned to the laboratory for positive identification. Some fish species were too numerous to count (Clupeidae spp. and larval fish), so the numbers of these species were visually estimated.

All point-census and linear transect surveys generally occurred over a five-day period during February and June 2002. Surveys were not conducted during the same day as survey plot establishment. Two replicate surveys were conducted during each of the sampling seasons (total of four surveys at each station). Replicate surveys within each season occurred at least one day apart. Within the sampling window, the survey dates were scheduled to maximize visibility in the water column (i.e., during calm periods).

## Sampling Stations

Monitoring station summary information is provided in Table 2.4-1. At the Gerkin site, two point-census locations occurred at one or two reef ball units. Additional point-census locations occurred at PVC ribcage design and a PVC web design. The PVC ribcage design was roughly $1.2 \mathrm{~m} \times 2.4 \mathrm{~m}$ at the base, with curved 5 cm PVC pipe arching to provide about 1.2 m of vertical relief. The PVC web unit is approximately 1.5 m high and has a geoweb plastic material in a roughly triangular shape. The linear transect at the Gerkin site occurred along a string of four PVC ribcage units.

At the Hart site, two of the point-census locations were individual reef ball units. The third point-census location was a small pile of concrete block. The linear transect at this location included three reef ball units on top of a large concrete block pile covering approximately $36 \mathrm{~m}^{2}$.

Reef ball units (both single modules and pairs) comprised all of the point-census locations at the Walker reef site. Two linear transects were monitored at the Walker site, with one being along a staggered string of 16 reef balls. The second linear transect at this site included four sets of reef balls in concrete units (each unit contains four reef balls imbedded in a concrete slab).

The depth of the water at the reef sampling stations varied from 2.4 m to 4.6 m . The Gerkin site was $3,800 \mathrm{~m}$ from a tidal pass, while the Walker site was most distant at $6,100 \mathrm{~m}$ from a pass). The reef ball units and concrete block piles had similar complexity with a rank of

Table 2.4-1 Physical parameters of artificial reef assessment stations.

| Reef | Site | Station <br> Type | Distance <br> to Pass $(\mathbf{m})$ | Depth <br> $(\mathbf{m})$ | Material <br> Type ${ }^{\mathbf{1}}$ | Complexity | Relief <br> $(\mathbf{m})$ | Surface <br> Area (m-2) |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | ---: | ---: |
| Gerkin | GL1 | Linear | 3800 | 4.6 | PVC RC (4) | 3 | 1.2 | 23.44 |
| Gerkin | GP1 | Point | 3800 | 2.4 | RB (2) | 2 | 0.8 | 1.31 |
| Gerkin | GP2 | Point | 3800 | 2.4 | RB | 2 | 0.8 | 0.66 |
| Gerkin | GP3 | Point | 3800 | 4.3 | PVC RC | 3 | 1.2 | 5.86 |
| Gerkin | GP4 | Point | 3800 | 4.3 | PVC Web | 3 | 1.5 | 2.75 |
| Hart | HL1 | Linear | 4800 | 3.1 | CB, RB (3) | 4 | 1.2 | 35.72 |
| Hart | HP1 | Point | 4800 | 3.1 | RB | 2 | 0.8 | 0.66 |
| Hart | HP2 | Point | 4800 | 3.1 | CB | 2 | 0.3 | 3.91 |
| Hart | HP3 | Point | 4800 | 3.1 | RB | 2 | 0.8 | 0.66 |
| Walker | WL1 | Linear | 6100 | 3.7 | RB in concrete $(4)$ | 3 | 0.9 | 20.00 |
| Walker | WL2 | Linear | 6100 | 3.1 | RB (16) | 3 | 0.8 | 10.48 |
| Walker | WP1 | Point | 6100 | 3.1 | RB | 2 | 0.8 | 0.66 |
| Walker | WP2 | Point | 6100 | 3.1 | RB | 2 | 0.8 | 0.66 |
| Walker | WP3 | Point | 6100 | 3.1 | RB (2) | 2 | 0.8 | 1.31 |

[^0]two; while the PVC ribcage, PVC web, reef ball in concrete, and multiple reef ball units had a complexity of three. The linear transect at the Hart site, which included both concrete blocks and reef balls, had the highest complexity rank at four. The relief of the reef material (height above the bottom) varied from a low of 0.3 m for the concrete block pile at the Hart site to a high of 1.5 m for the PVC web unit at the Gerkin site. The surface area of the bottom encompassed by the artificial reef surveyed varied from a low of $0.66 \mathrm{~m}^{2}$ for an individual reef ball unit to a high of roughly $36 \mathrm{~m}^{2}$ for the Hart linear transects area.

## Data Analysis

All data was entered into a MS Access database. This database was extracted and analyzed using MS Excel and SPSS software. The substrate area of the material was used to convert the raw abundance data to density to calculate the catch-per-unit-area (CPUA) measure based upon $100 \mathrm{~m}^{2}$.

Species not identifiable down to species level were pooled into multi-species complexes and treated as individual species for all subsequent data analyses. Included in these species complexes were larval fish, Gobiidae spp., Scorpaenidae spp., Prionotus spp., Blennidae spp., Balistidae spp., Clupeidae spp., Calamus spp., and Eucinostomus spp. Summary statistics were developed from the raw abundance numbers by both dominant taxa and select taxa after pooling all sites and sampling seasons. Summary statistics include the number of individuals for that size class, the relative abundance for each taxon, the percent of sample for which the taxon was observed and totals for each size class.

One species group (Clupeidae spp.) was removed from the database after the initial summary statistics discussed above. This species was excluded due to the high numbers observed within the some samples, and transient nature, which would unduly sway the data analyses. Larval fish were not excluded, despite occasional high numbers, however, because nursery habitat value was the focus of the study. Subsequent analyses were based upon faunal density, following the removal of Clupeidae spp. as mentioned above.

Between reef site-related density differences, within reef site-related density differences and artificial reef material type-related density differences were determined by all taxa (except Clupeidae spp.), select taxa and size class.

Untransformed data were tested for normality by reviewing box plots and normality plots, identifying the five highest and five lowest mean densities by analysis factor and testing using the Kolmogorov-Smirnov and Shapiro-Wilk statistics. Data were generally not normally distributed. Data were transformed $(\log 10)$ and retested. Data were slightly more normally distributed after transformation and $\log$ transformed mean densities were used in subsequent statistical tests. Homogeneity of variance was variable for means and medians depending on the analysis factor. Data were tested for significant differences among selected analysis factors using univariate analysis of variance (unianova) where alpha $=$ 0.05 . P values were reported for significantly different analysis factors.

### 3.0 RESULTS

### 3.1 Habitat Mapping and GIS Database

The SBNEP shoreline has a total length of $2,166,490$ feet ( 410.3 miles). Table 3.1-1 summarizes the total shoreline length and area for each bay segment. The geomorphology, shoreline morphology and intertidal vegetation categories within the study area are summarized in Table 3.1-2. The percentage for each feature within the geomorphology, shore morphology and intertidal vegetation tiers is graphically depicted in Figure 3.1-1.

Altered shorelines are the dominant feature (1,379,127 feet) within the geomorphology tier, totaling $63.7 \%$ of the available shoreline length. The remaining feature classes range from $21.3 \%$ ( 460,823 feet) for linear shoreline down to creek mouth at one percent. Within the shore morphology tier, bulkhead was the dominant shore type at $45.3 \%$ ( 981,120 feet), followed by deep wetland fringing at $25.3 \%$. ( 547,190 feet) The remaining feature types ranged from $3.5(75,201$ feet) to $9 \%(203,089$ feet) of the linear distance of the shoreline. No vegetation is present along $50.5 \%(1,094,429$ feet) of the shoreline of the SBNEP study

Table 3.1-1 Summary of total shoreline length and area by bay segment.

| Bay Segment | Shoreline Length (Feet) | Area (Acres) |
| :---: | :---: | :---: |
| 1 | 147,722 | $1,766.13$ |
| 2 | 160,790 | $2,958.85$ |
| 3 | 187,319 | $2,107.87$ |
| 4 | 11,971 | 136.23 |
| 5 | 128,138 | $2,422.00$ |
| 6 | 152,539 | $5,044.50$ |
| 7 | 157,307 | $5,614.00$ |
| 8 | 114,913 | $3,450.16$ |
| 9 | 26,790 | 213.00 |
| 10 | 123,922 | $3,409.22$ |
| 11 | 112,218 | $2,399.76$ |
| 12 | 29,170 | 625.35 |
| 13 | 445,607 | $1,624.92$ |
| 14 | 175,242 | $2,007.53$ |
| 15 | 22,326 | 74.27 |
| 16 | 170,516 | 840.36 |

Table 3.1-2 Summary of geomorphology, shore morphology and intertidal vegetation features within the NEP study area.

| Feature | Length (feet) | Percentage |
| :--- | ---: | ---: |
| Geomorphology |  |  |
| Altered | $1,379,127$ | 63.7 |
| Bight | 48,346 | 2.2 |
| Tidal Creek | 116,747 | 5.4 |
| Embayment | 150,212 | 6.9 |
| Linear Shoreline | 460,873 | 21.3 |
| Creek Mouth | 11,186 | 0.5 |
|  |  |  |
| Shore Morphology |  |  |
| Bulkhead | 451,120 | 9.3 |
| Riprap | 203,089 | 4.3 |
| Beach | 93,229 | 3.8 |
| Upland Shoreline | 82,445 | 3.5 |
| Patchy Wetland Fringing | 75,201 | 8.5 |
| Solid Wetland Fringing | 184,216 | 25.3 |
| Deep Wetland Fringing | 547,190 |  |
|  |  |  |
| Intertidal Vegetation |  | 0.4 |
| Australian Pine | 7,726 | 1.3 |
| Brazilian Pepper | 29,090 | 0.1 |
| Cattail | 3,067 | 0.1 |
| Juncus | 3,172 | 0.4 |
| Leather Fern | 8,111 | 41.7 |
| Mangrove | 903,210 | 50.5 |
| None | $1,094,429$ | 2.4 |
| Other Vegetation | 52,925 | 0.5 |
| Spartina | 11,898 | 2.4 |
| Terrestrial Vegetation | 52,863 |  |
|  |  |  |



Figure 3.1-1. Pie charts depicting the total percentage of each geomorphology, shore morphology, and intertidal vegetation feature within the SBNEP area.
area, $41.7 \%$ ( 903,210 feet) of the shoreline is dominated by mangrove vegetation and remaining vegetation types each account for less than $2 \%$ of the shoreline.

Table 3.1-3 provides the proportion of each feature class within the geomorphology, shore morphology and intertidal vegetation tiers by bay segment. Figures 3.1-2 through 3.1-4 graphically depict the percentage of each feature class within the geomorphology, shore morphology and intertidal vegetation tiers, respectively. The percent of each feature class within each tier varies widely across the SBNEP study area. This reflects the variable nature of shoreline alteration and natural habitat distribution within the bay system. It is important to note that upland-cut canal systems are included in the bay segmentation scheme. The canal systems tend to overstate the linear shoreline length occupied by altered habitat, seawalls, bulkheads and areas of no intertidal vegetation. The upland-cut canal systems figure most prominently into bay segments $3,6,7,10$ and 13 .

Included in Table 3.1-3 is the linear length of each bay segment as well as the total water area within that bay segment in acres. The shoreline length divided by the total water area of the bay segment results in a value known as the shoreline index (feet/acres). The shoreline index is graphically depicted by bay segment on Figure 3.1-5. The shoreline index essentially provides the ratio of shoreline length relative to the water area in that segment. Within the SBNEP study area, the overall shoreline index is 62 feet/acre.

Appendix A contains graphic depictions of the geomorphology, shore morphology and intertidal vegetation tiers within the SBNEP study area. These map layouts help to provide a general overview of the feature classes within each tier, but are not able to show the detail available on the GIS work product.

Table 3.1-3 Proportion of feature classes within geomorphology, shore morphology and intertidal vegetation tiers.

| SHORE TYPE | Segment 1 | \% | Segment 2 | \% | Segment 3 | \% | Segment 4 | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Segment Length (feet) |  |  |  |  |  |  |  |  |
| Geomorphology |  |  |  |  |  |  |  |  |
| Altered | 133,534 | 90 | 47,074 | 29 | 122,862 | 66 | 5,944 | 50 |
| Bight | 518 | 0 | 8,932 | 6 | 3,115 | 2 | 1,606 | 13 |
| Tidal Creek | 0 | 0 | 500 | 0 | 5,018 | 3 | 0 | 0 |
| Embayment | 359 | 0 | 23,822 | 15 | 14,668 | 8 | 0 | 0 |
| Linear Shoreline | 13,311 | 9 | 80,462 | 50 | 41,656 | 22 | 4,352 | 36 |
| Creek Mouth | 0 | 0 | 0 | 0 | 0 | 0 | 69 | 1 |
| Shore Morphology |  |  |  |  |  |  |  |  |
| Beach | 1,819 | 1 | 4,923 | 3 | 11,348 | 6 | 2,409 | 20 |
| Bulkhead | 121,292 | 82 | 31,635 | 20 | 112,831 | 60 | 5,524 | 46 |
| Deep Wetland Fringing | 6,610 | 4 | 95,149 | 59 | 35,075 | 19 | 1,848 | 15 |
| Patchy Wetland Fringing | 7,140 | 5 | 6,749 | 4 | 6,433 | 3 | 553 | 5 |
| Riprap | 5,907 | 4 | 5,689 | 4 | 5,111 | 3 | 420 | 4 |
| Solid Wetland Fringing | 4,954 | 3 | 15,977 | 10 | 14,500 | 8 | 1,217 | 10 |
| Upland Shoreline | 0 | 0 | 668 | 0 | 2,020 | 1 | 0 | 0 |
| Intertidal Vegetation |  |  |  |  |  |  |  |  |
| Australian Pine | 0 | 0 | 71 | 0 | 600 | 0 | 0 | 0 |
| Brazilian Pepper | 0 | 0 | 0 | 0 | 210 | 0 | 0 | 0 |
| Cattail | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Leather Fern | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Juncus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Mangrove | 24,963 | 17 | 121,990 | 76 | 55,808 | 30 | 3,618 | 30 |
| None | 121,541 | 82 | 38,417 | 24 | 125,604 | 67 | 8,353 | 70 |
| Other Vegetation | 1,065 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Spartina | 0 | 0 | 312 | 0 | 277 | 0 | 0 | 0 |
| Terrestrial Vegetation | 153 | 0 | 0 | 0 | 4,820 | 3 | 0 | 0 |
| Total Segment Length (feet) | 147,722 |  | 160,790 |  | 187,319 |  | 11,971 |  |
| Total Water Area (acres) | 1,766.13 |  | 2,958.85 |  | 2,107.87 |  | 136.23 |  |
| Shoreline Index (ft/acres) | 83.6 |  | 54.3 |  | 88.9 |  | 87.9 |  |

## Table 3.1-3. (Continued)

| SHORE TYPE | Segment 5 | \% | Segment 6 | \% | Segment 7 | \% | Segment 8 | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Segment Length (feet) |  |  |  |  |  |  |  |  |
| Geomorphology |  |  |  |  |  |  |  |  |
| Altered | 86,786 | 68 | 72,112 | 47 | 125,260 | 80 | 98,428 | 86 |
| Bight | 1,008 | 1 | 6,071 | 4 | 1,111 | 1 | 750 | 1 |
| Tidal Creek | 0 | 0 | 1,586 | 1 | 0 | 0 | 0 | 0 |
| Embayment | 1,447 |  | 12,186 | 8 | 1,551 | , | 508 | 0 |
| Linear Shoreline | 38,898 | 30 | 59,999 | 39 | 29,385 | 19 | 14,235 | 12 |
| Creek Mouth | 0 | 0 | 585 | 0 | 0 | 0 | 993 | 1 |
| Shore Morphology |  |  |  |  |  |  |  |  |
| Beach | 7,755 | 6 | 3,484 | 2 | 772 | 1 | 5,134 | 4 |
| Bulkhead | 71,051 | 55 | 56,153 | 37 | 75,617 | 48 | 75,029 | 65 |
| Deep Wetland Fringing | 26,632 | 21 | 58,590 | 38 | 26,407 | 17 | 9,961 | 9 |
| Patchy Wetland Fringing | 2,800 | 2 | 4,428 | 3 | 10,897 | 7 | 4,021 | 4 |
| Riprap | 4,987 | 4 | 7,642 | 5 | 25,459 | 16 | 17,779 | 15 |
| Solid Wetland Fringing | 14,913 | 12 | 22,017 | 14 | 17,871 | 11 | 2,858 | 2 |
| Upland Shoreline | 0 | 0 | 225 | 0 | 284 | 0 | 130 | 0 |
| Intertidal Vegetation |  |  |  |  |  |  |  |  |
| Australian Pine | 0 | 0 | 0 | 0 | 0 | 0 | 340 | 0 |
| Brazilian Pepper | 0 | 0 | 0 | 0 | 329 | 0 | 1,235 | 1 |
| Cattail | 0 | 0 | 400 | 0 | 0 | 0 | 0 | 0 |
| Leather Fern | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Juncus | 0 | 0 | 0 | 0 | 0 | 0 | 389 | 0 |
| Mangrove | 58,838 | 46 | 89,047 | 58 | 71,554 | 45 | 23,117 | 20 |
| None | 69,042 | 54 | 58,771 | 39 | 81,160 | 52 | 88,561 | 77 |
| Other Vegetation | 0 | 0 | 225 | 0 | 619 | 0 | 0 | 0 |
| Spartina | 258 | 0 | 4,096 | 3 | 0 | 0 | 1,177 | 1 |
| Terrestrial Vegetation | 0 | 0 | 0 | 0 | 3,645 | 2 | 94 | 0 |
| Total Segment Length (feet) | 128,138 |  | 152,539 |  | 157,307 |  | 114,913 |  |
| Total Water Area (acres) | 2,422.00 |  | 5,044.50 |  | 5,614.00 |  | 3,450.16 |  |
| Shoreline Index (ft/acres) | 52.9 |  | 30.2 |  | 28.0 |  | 33.3 |  |

## Table 3.1-3. (Continued)

| SHORE TYPE | Segment 9 | \% | Segment 10 | \% | Segment 11 | \% | Segment 12 | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Segment Length (feet) |  |  |  |  |  |  |  |  |
| Geomorphology |  |  |  |  |  |  |  |  |
| Altered | 13,284 | 50 | 106,193 | 86 | 92,248 | 82 | 8,446 | 29 |
| Bight | 2,965 | 11 | 1,911 | 2 | 640 | 1 | 533 | 2 |
| Tidal Creek | 0 | 0 | 3,650 | 3 | 5,413 | 5 | 0 | 0 |
| Embayment | 10,408 | 39 | 2,544 | 2 | 164 | 0 | 10,504 | 36 |
| Linear Shoreline | 133 | 0 | 9,513 | 8 | 13,548 | 12 | 9,687 | 33 |
| Creek Mouth | 0 | 0 | 111 | 0 | 204 | 0 | 0 | 0 |
| Shore Morphology |  |  |  |  |  |  |  |  |
| Beach | 13,809 | 52 | 3,375 | 3 | 15,179 | 14 | 9,491 | 33 |
| Bulkhead | 4,966 | 19 | 85,925 | 69 | 66,761 | 59 | 1,263 | 4 |
| Deep Wetland Fringing | 960 | 4 | 22,783 | 18 | 5,086 | 5 | 11,335 | 39 |
| Patchy Wetland Fringing | 2,067 | 8 | 345 | 0 | 1,728 | 2 | 736 | 3 |
| Riprap | 2,921 | 11 | 6,140 | 5 | 17,412 | 16 | 6,073 | 21 |
| Solid Wetland Fringing | 2,067 | 8 | 4,836 | 4 | 5,524 | 5 | 119 | 0 |
| Upland Shoreline | 0 | 0 | 519 | 0 | 527 | 0 | 153 | 1 |
| Intertidal Vegetation |  |  |  |  |  |  |  |  |
| Australian Pine | 0 | 0 | 767 | 1 | 125 | 0 | 459 | 2 |
| Brazilian Pepper | 0 | 0 | 209 | 0 | 630 | 1 | 52 | 0 |
| Cattail | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Leather Fern | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Juncus | 0 | 0 | 0 | 0 | 180 | 0 | 0 | 0 |
| Mangrove | 5,121 | 19 | 28,930 | 23 | 17,325 | 15 | 12,433 | 43 |
| None | 20,945 | 78 | 89,514 | 72 | 87,569 | 78 | 15,458 | 53 |
| Other Vegetation | 171 | 1 | 3,074 | 2 | 4,306 | 4 | 329 | 1 |
| Spartina | 37 | 0 | 164 | 0 | 725 | 1 | 0 | 0 |
| Terrestrial Vegetation | 516 | 2 | 1,265 | 1 | 1,357 | 1 | 440 | 2 |
| Total Segment Length (feet) | 26,790 |  | 123,922 |  | 112,218 |  | 29,170 |  |
| Total Water Area (acres) | 213.00 |  | 3409.22 |  | 2,399.76 |  | 625.35 |  |
| Shoreline Index (ft/acres) | 125.8 |  | 36.3 |  | 46.8 |  | 46.6 |  |

[^1]
## Table 3.1-3. (Continued)

| SHORE TYPE | Segment 13 | \% | Segment 14 | \% | Segment 15 | \% | Segment 16 | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Segment Length (feet) |  |  |  |  |  |  |  |  |
| Geomorphology |  |  |  |  |  |  |  |  |
| Altered | 305,761 | 69 | 53,146 | 30 | 7,276 | 33 | 100,773 | 59 |
| Bight | 8,553 | 2 | 6,653 | 4 | 593 | 3 | 6,352 | 4 |
| Tidal Creek | 60,170 | 14 | 21,565 | 12 | 0 | 0 | 18,846 | 11 |
| Embayment | 28,394 | 6 | 33,464 | 19 | 6,137 | 27 | 11,499 | 7 |
| Linear Shoreline | 36,972 | 8 | 57,893 | 33 | 8,320 | 37 | 32,234 | 19 |
| Creek Mouth | 5,757 | 1 | 2,522 | 1 | 0 | 0 | 812 | 0 |
| Shore Morphology |  |  |  |  |  |  |  |  |
| Beach | 5,058 | 1 | 4,562 | 3 | 442 | 2 | 3,669 | 2 |
| Bulkhead | 208,142 | 47 | 32,517 | 19 | 7,203 | 32 | 25,210 | 15 |
| Deep Wetland Fringing | 87,378 | 20 | 92,384 | 53 | 10,815 | 48 | 56,176 | 33 |
| Patchy Wetland Fringing | 17,112 | 4 | 4,738 | 3 | 125 | 1 | 5,329 | 3 |
| Riprap | 37,367 | 8 | 11,661 | 7 | 73 | 0 | 48,449 | 28 |
| Solid Wetland Fringing | 35,423 | 8 | 17,490 | 10 | 3,514 | 16 | 20,937 | 12 |
| Upland Shoreline | 55,128 | 12 | 11,891 | 7 | 154 | 1 | 10,746 | 6 |
| Intertidal Vegetation |  |  |  |  |  |  |  |  |
| Australian Pine | 4,229 | 1 | 275 | 0 | 154 | 1 | 707 | 0 |
| Brazilian Pepper | 18,151 | 4 | 4,264 | 2 | 0 | 0 | 4,010 | 2 |
| Cattail | 1,214 | 0 | 1,148 | 1 | 0 | 0 | 305 | 0 |
| Leather Fern | 130 | 0 | 1,571 | 1 | 0 | 0 | 902 | 1 |
| Juncus | 3,974 | 1 | 2,853 | 2 | 0 | 0 | 1,284 | 1 |
| Mangrove | 159,184 | 36 | 112,306 | 64 | 14,884 | 67 | 104,090 | 61 |
| None | 202,696 | 45 | 34,977 | 20 | 6,786 | 30 | 45,036 | 26 |
| Other Vegetation | 32,013 | 7 | 8,085 | 5 | 462 | 2 | 2,576 | 2 |
| Spartina | 1,523 | 0 | 2,165 | 1 | 0 | 0 | 1,164 | 1 |
| Terrestrial Vegetation | 22,493 | 5 | 7,599 | 4 | 40 | 0 | 10,441 | 6 |
| Total Segment Length (feet) | 445,607 |  | 175,242 |  | 22,326 |  | 170,516 |  |
| Total Water Area (acres) | 1,624.92 |  | 2007.53 |  | 74.27 |  | 840.36 |  |
| Shoreline Index (ft/acres) | 274.2 |  | 87.3 |  | 300.6 |  | 202.9 |  |



Geomorphology
$\square$ Altered
$\square$ Bight
-Tidal Creek
QEmbayment
$\square$ Linear Shoreline
-Creek Mouth

Figure 3. 1-2. Percentage of geomorphology feature classes within each bay segment.


Figure 3.1-3 Percentage of shore morphology feature classes within each bay segment.


Figure 3.1-4. Percentage of intertidal vegetation feature classes within each bay segment.


Figure 3.1-5. Shoreline index (feet/acre) for each bay segment.

## Tidal Wetlands

The composition of tidal wetland features within each bay segment is presented in Table 3.1-4. Maps depicting these tidal wetland features are provided in Appendix A. Figure 3.16 depicts the percentages of the tidal wetland features within the SBNEP study area. Bays and Estuaries is the dominant feature class at $91.1 \%$ ( $31,588.6$ acres), followed by Mangrove Swamps at $6.7 \%$ ( $2,340.3$ acres)

The tidal wetland feature class acreage is presented by bay segment in Table 3.1-4 along with its associated percentage of that bay segment. A total intertidal area is also presented in Table 3.1-4 and reflects the acreage of the feature classes less the Bays and Estuaries class. The percentage of the total intertidal area by bay segment is graphically presented in Figure 3.1-7. The percentage of each tidal wetland feature class is presented by bay segment in Appendix A.

### 3.2 Critical Nursery and Restored Habitat Assessment

## Physicochemical Parameters

Fishery sampling activity was limited to two seasonal periods (February and June 2002) and each sampling period of five days duration (Figure 3.2-1). Lacking complete seasonal representation of physicochemical conditions or faunal communities, no attempt was made to analyze the correlation between physicochemical parameters and fish community structure. However, an analysis was made to describe the conditional (dynamic) habitat components of temperature and salinity between fishery sampling stations. Water temperature and salinity values were recorded during SBNEP fishery sampling activity and are plotted for each station as depicted in Figures 3.2-2 and 3.2-3. These results depict typical winter-summer ranges of physicochemical parameters within shallow water, shore fringing habitats of Sarasota Bay.

Table 3.1-4 Composition of tidal wetland features by bay segment.

| Bay Segment Number | 1 |  | 2 |  | 3 |  | 4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) |
| Bays \& Estuaries | 1,664.5 | 94.2 | 1,996.6 | 67.5 | 1,869.8 | 88.7 | 107.1 | 78.6 |
| Beaches Other than Swimming | 0.0 | 0.0 | 0.0 | 0.0 | 12.2 | 0.6 | 1.3 | 0.9 |
| Mangrove Swamps | 72.9 | 4.1 | 846.9 | 28.6 | 202.8 | 9.6 | 9.6 | 7.0 |
| Oyster Bars | 0.0 | 0.0 | 6.5 | 0.2 | 2.0 | 0.1 | 0.0 | 0.0 |
| Salt Barren | 0.0 | 0.0 | 8.3 | 0.3 | 0.6 | 0.0 | 0.0 | 0.0 |
| Saltwater Marshes | 0.0 | 0.0 | 4.4 | 0.1 | 12.0 | 0.6 | 0.0 | 0.0 |
| Unvegetated Bottoms | 28.7 | 1.6 | 96.1 | 3.2 | 8.6 | 0.4 | 18.3 | 13.4 |
| Total intertidal area (no BE) | 101.7 | 5.8 | 962.2 | 32.5 | 238.3 | 11.3 | 29.1 | 21.4 |
| Total area of all water features | 1,766.1 | 100.0 | 2,958.5 | 100.0 | 2,108.1 | 100.0 | 136.2 | 100.0 |


| Bay Segment Number | 5 |  | 6 |  | 7 |  | 8 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature <br> Acreage <br> (Acres) | Percent of Total Water Area in Segment (\%) | Feature <br> Acreage <br> (Acres) | Percent of Total Water Area in Segment (\%) | Feature <br> Acreage <br> (Acres) | Percent of Total Water Area in Segment (\%) |
| Bays \& Estuaries | 2,240.3 | 92.5 | 4,526.3 | 89.7 | 5,453.3 | 97.1 | 3,406.7 | 98.7 |
| Beaches Other than Swimming | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Mangrove Swamps | 147.2 | 6.1 | 417.5 | 8.3 | 130.8 | 2.3 | 11.2 | 0.3 |
| Oyster Bars | 0.0 | 0.0 | 5.7 | 0.1 | 0.0 | 0.0 | 10.3 | 0.3 |
| Salt Barren | $<0.0$ | $<0.1$ | 16.7 | 0.3 | 0.0 | 0.0 | 0.8 | 0.0 |
| Saltwater Marshes | 0.0 | 0.0 | 32.8 | 0.7 | $<0.0$ | 0.0 | 0.0 | 0.0 |
| Unvegetated Bottoms | 33.9 | 1.4 | 45.6 | 0.9 | 29.9 | 0.5 | 21.2 | 0.6 |
| Total intertidal area (no BE) | 181.7 | 7.5 | 518.2 | 10.3 | 160.7 | 2.9 | 43.5 | 1.3 |
| Total area of all water features | 2,422.0 | 100.0 | 5,044.5 | 100.0 | 5614.0 | 100.0 | 3,450.2 | 100.0 |

BE - Bays and Estuaries

Table 3.1-4 (Continued)

| Bay Segment Number | 9 |  | 10 |  | 11 |  | 12 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature Acreage (Acres) | Percent of Total <br> Water Area in <br> Segment <br> (\%) | Feature Acreage (Acres) |  | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) |
| Bays \& Estuaries | 159.7 | 74.9 | 3,312.0 | 97.1 | 2,387.6 | 99.5 | 554.0 | 88.6 |
| Beaches Other than Swimming | 5.0 | 2.4 | 0.0 | 0.0 | 0.0 | 0.0 | 2.0 | 0.3 |
| Mangrove Swamps | 35.4 | 16.6 | 46.4 | 1.4 | 0.6 | 0.0 | 63.9 | 10.2 |
| Oyster Bars | 0.0 | 0.0 | 0.4 | 0.0 | 2.9 | 0.1 | 0.3 | 0.1 |
| Salt Barren | 0.3 | 0.1 | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Saltwater Marshes | 1.2 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Unvegetated Bottoms | 11.5 | 5.4 | 49.7 | 1.5 | 8.7 | 0.4 | 5.1 | 0.8 |
| Total intertidal area (no BE) | 53.4 | 25.1 | 97.2 | 2.9 | 12.2 | 0.5 | 71.3 | 11.4 |
| Total area of all water features | 213.0 | 100.0 | 3,409.2 | 100.0 | 2,399.8 | 100.0 | 625.3 | 100.0 |


| Bay Segment Number | 13 |  | 14 |  | 15 |  | 16 |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) | Feature Acreage (Acres) | Percent of Total Water Area in Segment (\%) |
| Bays \& Estuaries | 1,445.4 | 89.0 | 1,666.6 | 83.0 | 50.2 | 67.6 | 748.7 | 89.1 | 31,588.6 | 91.1 |
| Beaches Other than Swimming | 0.0 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 21.6 | 0.1 |
| Mangrove Swamps | 146.7 | 9.0 | 147.1 | 7.3 | 5.5 | 7.3 | 55.9 | 6.7 | 2,340.3 | 6.7 |
| Oyster Bars | 21.0 | 1.3 | 28.6 | 1.4 | 0.1 | 0.1 | 5.6 | 0.7 | 83.5 | 0.2 |
| Salt Barren | 2.1 | 0.1 | 8.2 | 0.4 | 0.2 | 0.3 | 1.7 | 0.2 | 39.7 | 0.1 |
| Saltwater Marshes | 0.4 | 0.0 | 10.3 | 0.5 | 4.7 | 6.4 | 2.3 | 0.3 | 68.2 | 0.2 |
| Unvegetated Bottoms | 9.2 | 0.6 | 146.3 | 7.3 | 13.6 | 18.3 | 26.1 | 3.1 | 552.3 | 1.6 |
| Total intertidal area (no BE) | 179.4 | 11.0 | 340.9 | 17.0 | 24.1 | 32.4 | 91.6 | 10.9 | 3105.6 | 8.9 |
| Total area of all water features | 1,624.9 | 100.0 | 2,007.5 | 100.0 | 74.3 | 100.0 | 840.4 | 100.0 | 34,694.2 | 100.0 |

BE - Bays and Estuaries


Figure 3.1-6. Tidal wetland feature class percentages for entire Sarasota Bay study area.


Figure 3.1-7. Percentage of total intertidal area (total area less bays and estuaries) by bay segment.

Figure 3.2-1. Plot of water surface elevations and daily sampling activity windows during two 5-day periods of Sarasota Bay Fishery Habitat Assessment in February and June 2002. Water level records derived from continuously recording tide gauge monitoring stations which span north, middle and south portions of SBNEP study area. NOTE: Shaded rectangles (windows) denote time span of daily sampling activity for fishery sample collections.



Figure 3.2-2. Plot of temperature values recorded at sampling stations during Sarasota Bay Fishery Habitat Assessment in February and June 2002. Maximum-minimum range denoted by bar graph. Individual data values denoted by symbols. NOTE: Missing data values with fewer than 4 recorded values indicated at 15 of 40 sample stations.

Temperature Range at SBNEP Fishery Sampling Stations


Figure 3.2-3. Plot of salinity values recorded at sample stations during Sarasota Bay Fishery Habitat Assessment in February and June 2002. Maximum-minimum range denoted by bar graph. Individual data values denoted by symbols. NOTE: Missing data values with fewer than 4 recorded values indicated at 19 of 40 sample stations.

Salinity Range at SBNEP Fishery Sampling Stations


Water temperature values were not significantly different between or within stations (p $>0.05$ ) and ranged from a minimum of $13.4{ }^{\circ} \mathrm{C}$ to a maximum of $32.6^{\circ} \mathrm{C}$ (Figure 3.2-2). The Kruskal-Wallis test statistic $\left(H_{C}=19.56\right)$ as compared to the critical value of the Chisquare statistic $\left(\chi^{2}{ }_{0.05,39}=54.57\right)$ indicated acceptance of the null hypothesis regarding no difference between measured water temperatures. Daytime maximum-minimum water temperatures spanned about $4{ }^{\circ} \mathrm{C}$ in June 2002 but increased to a range of about $7{ }^{\circ} \mathrm{C}$ in February 2002. Qualitatively, temperature range differences appeared related to station depths or differences in duration of tidal inundation and sun exposure, where widest temperature ranges occurred at stations with shallower depths and irregular tidal inundation/exposure cycles, while narrowest temperature ranges occurred at stations with deeper depths and more constant tidal inundation or moderation of sun exposure by canopy cover.

Salinity values were significantly different between stations ( $\mathrm{p} \leq 0.05$ ) and ranged from a minimum of 0.6 ppt to a maximum of 38.7 ppt (Figure 3.2-3). The Kruskal-Wallis test statistic $\left(H_{C}=101.96\right)$ as compared to the critical value of the Chi-square statistic $\left(\chi^{2}{ }_{0.05,39}=\right.$ 54.57) indicated rejection of the null hypothesis regarding no difference between measured salinities. The majority of stations ( 32 locations) exhibited either euhaline salinity ( 23 ppt or greater) conditions or a narrow range of salinity variation ( $<10 \mathrm{ppt}$ ) during sampling events. Highest salinities and narrowest ranges were recorded at C5--Sister Keys ( 38.7 ppt ) and R6--Durante Park ( 37.6 ppt ). Another eight stations ( 8 locations) exhibited either polyhaline salinity ( $<23.0$ ppt or between 15 to 28 ppt ) conditions (e.g., C15 and C16--Phillippi Creek, C21--South Creek), oligohaline salinity ( $<15 \mathrm{ppt}$ or between 2 to 15 ppt ) conditions (e.g., C8--El Conquistador, C15--Phillippi Creek, R4--Durante Park, R12 and R13--Hog Creek), or limnetic salinity (between 0 to 4 ppt ) conditions (e.g., C8--El Conquistador, R14--Selby Gardens Pond). The widest range of salinity variation was recorded at C15--Phillippi Creek which exhibited a single-season range of 14.6 ppt and a two-season range of 24.2 ppt during sampling periods.

An analysis was made of annualized average salinity conditions in both open bay waters and tributary streams throughout the SBNEP study area in order to document the range of dynamic (salinity) habitat available to the fishery community. A seven year period (1986 through 1992) was selected since this period coincides with available water quality monitoring databases for both open bay and tidal streams throughout Manatee and Sarasota County estuarine waters. Bulger et al. (1990) points out that efforts to subdivide estuaries on the basis of salinity have long been based on the observation that estuarine species are not evenly distributed across salinity gradients. For this reason, an estuarine classification scheme (Bulger et al., 1990) was applied to the SBNEP study area which uses salinity zones to organize information on the spatial and temporal distribution of species within estuaries. The classification scheme is composed of five overlapping salinity zones, and was developed through use of multivariate statistical analysis (i.e., principal components analysis) of reported field distribution of fish and invertebrates. A rationale for this approach is that a statistically-based classification scheme may overcome subjective or qualitative assessments of how animals use estuaries.

NOAA Estuarine Classification (Bulger et al., 1990)

| Zone | Salinity Range (ppt) |
| :--- | :--- |
| Euhaline | 23 to Marine $(>30 \mathrm{ppt})$ |
| Polyhaline | 15 to 28 |
| Mesohaline | 11 to 19 |
| Oligohaline | 2 to 15 |
| Limnetic | Freshwater $(0.5 \mathrm{ppt})$ to 4 |

Figures 3.2-4 and 3.2-5 depict plots of mean, maximum and minimum salinity values computed as annualized averages over a seven year period (1986 to 1992) for the SBNEP study area. For the analysis of bay segments (Figure 3.2-4), the database was derived from that previously compiled for the SBNEP water quality trends analysis (Dixon and Heyl, 1999). This database was truncated to include observations with specific conductances greater than or equal to $5 \mathrm{mmhos} / \mathrm{cm}(5000 \mu \mathrm{mhos}$ or 2.67 PSU$)$. As a result, tidal stream environments are not represented in the 1999 water quality trends analysis for the SBNEP. In order to characterize the tidal stream environments within the SBNEP study area, a

Figure 3.2-4. Plot of mean, maximum and minimum salinity values within SBNEP bay segments during 1986 through 1992. Database derived from SBNEP Water Quality Trends Analysis (published September 1999) by Mote Marine Laboratory (MML) for which tidal stream data not reported. Annualized means computed from raw data (daily averages by station) and annual maximum-minimum values reported. Sevenyear composite means of annualized means and annual maximum-minimum values then derived and charted. NOTE: Database truncated by MML to observations with specific conductances greater than or equal to $5 \mathrm{mmhos} / \mathrm{cm}(5000 \mu \mathrm{mhos}$ or 2.67 PSU ) to avoid spurious trends generated by short term tributary sampling programs.

7-YEAR RUNNING MEANS OF SALINITY (1986-1992)


Figure 3.2-5. Plot of mean, maximum and minimum salinity values at Sarasota County tidal stream stations during 1986 through 1992. Database derived from Sarasota County Ambient Water Quality Monitoring Program. Annualized means computed from raw station data and annual maximum-minimum values reported. Seven-year composite means of annualized means and annual maximum-minimum values then derived and charted. Time period corresponds to SBNEP water quality trends analysis for bay segments (published September 1999) by Mote Marine Laboratory (MML) for which tidal stream data not reported.

7-YEAR RUNNING MEANS OF SALINITY (1986-1992)

separate database was compiled as derived from Sarasota County ambient water quality monitoring records for stream runs, such monitoring was discontinued after 1992. Together each salinity analysis supports the characterization of the dynamic (salinity) habitat conditions which are experienced by the juvenile fishery community within the SBNEP study area.

Results of salinity conditions derived from water quality trends analysis by bay segments indicate that, on average, euhaline conditions ( $>23 \mathrm{ppt}$, mean salinity) predominate in the open bay waters (all 16 bay segments) of the SBNEP study area. Minimum salinities lower than euhaline conditions are represented in just seven (7) bay segments, namely Segments 2 and 3 (vicinity Palma Sola Bay)--polyhaline, 20.5 and 21.7 ppt, mean minimum, respectively; Segments 8 (vicinity Bowlee's Creek) and 16 (vicinity South Creek)-mesohaline, 15.1 and 16.2 ppt, mean minimum, respectively; Segment 11 (vicinity Whitaker-Hudson Bayous)-oligohaline, 9.0 ppt , mean minimum; and Segments 13 (vicinity Phillippi Creek) and 14 (vicinity Little Sarasota Bay)--limnetic, 3.5 and 3.9 ppt, mean minimum, respectively. While these seven bay segments experience limited periods of reduced salinity conditions, these episodes are rare enough to maintain average salinity conditions at euhaline levels.

Results of salinity conditions derived for tidal streams (Figure 3.2-5) within the Sarasota County portion of the SBNEP study area indicate that, on average, mesohaline or lower salinities ( $<19 \mathrm{ppt}$, mean salinity) predominate in tidal streams ( 15 of 16 stream stations) of the SBNEP study area (Sarasota County portion). Only Hudson Bayou (Station 583) exhibited a polyhaline condition ( 27.5 ppt , mean salinity) on average. On the other hand, another seven (7) upper tidal stream stations experienced limnetic conditions wherein salinities did not rise above 4 ppt , mean maximum. The remaining eight (8) tidal stream stations are indicative of oligohaline-mesohaline conditions (2-15 ppt and 11-19 ppt zones, respectively) and represent tidally-influenced zones of tributary streams wherein many selected fishery species may aggregate during portions of their early life history.

## Fishery Sampling

Fishery sampling occurred synoptically in north and south portions of the study area during five consecutive days in each of two seasonal periods (February and June 2002) at 40 stations throughout the SBNEP study area which represented both natural and restored habitat conditions. A total of 160 seine samples yielded collections of 315,208 specimens of fishes and macrocrustaceans caught using two gear types and three deployment techniques (Table 3.2-1). The collections consisted of 80 taxa (including individual species or species groups) with some seasonal differences in abundance and numbers of taxa present in each seasonal period. Collections during February 2002 yielded 53 taxa and 128,952 specimens. Collections during June 2002 yielded 70 taxa and 186,256 specimens.

Variable catch efficiencies and selectivity of gears and deployment techniques are quantified in a series of catch statistics Tables (Tables 3.2-2 through 3.2-11). Among the five gear deployment techniques used, the number of all taxa and percentage abundance of the total catch between gears were approximately related in proportion to the sample size. In 93 samples collected using a $21.3-\mathrm{m}$ standard seine technique (sample area $140 \mathrm{~m}^{2}$ ), 71 taxa and 196,333 ( $62.3 \%$ ) individuals were caught over both seasons (Table 3.2-2). In 22 samples collected using a $21.3-\mathrm{m}$ boat seine technique (sample area $68 \mathrm{~m}^{2}$ ), 44 taxa and 104,025 (33.0\%) individuals were caught over both seasons (Table 3.2-4). In 21 samples collected using a $6.1-\mathrm{m}$ standard seine technique (sample area $39 \mathrm{~m}^{2}$ ), 22 taxa and 13,608 $(4.3 \%)$ individuals were caught over both seasons (Table 3.2-6). In 12 samples collected using a $6.1-\mathrm{m}$ boat seine technique (sample area $3 \mathrm{~m}^{2}$ ), 19 taxa and $777(0.2 \%$ ) individuals were caught over both seasons (Table 3.2-8). In 12 samples collected using a $6.1-\mathrm{m}$ nonstandard seine technique (sample areas between $23-36 \mathrm{~m}^{2}$ ), 21 taxa and 465 ( $0.1 \%$ ) individuals were caught over both seasons (Table 3.2-10).

Table 3.2-1. Master species list for all taxa collected by all gear deployment techniques during Sarasota Bay fishery habitat assessment in February and June 2002. Total number of individuals caught and presence or absence during each seasonal period is given.

| Species | Common Name | Guild Name | Selected Taxa ${ }^{1}$ | Number Caught | Feb 2002 | Jun 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Achirus lineatus | Lined sole | Nursery |  | 12 | Yes | Yes |
| Archosargus probatocephalus | Sheepshead | Nursery | X | 10 | Yes | Yes |
| Centropomus undecimalis | Snook | Nursery | X | 12 | Yes | Yes |
| Chaetodipterus faber | Atlantic spadefish | Nursery |  | 3 | No | Yes |
| Cynoscion nebulosus | Spotted seatrout | Nursery | X | 55 | No | Yes |
| Farfantepenaeus duorarum | Pink shrimp | Nursery | X | 336 | Yes | Yes |
| Hyporhamphus spp. | Halfbeak (juv) | Nursery |  | 2 | No | Yes |
| Lagodon rhomboides | Pinfish | Nursery |  | 15,890 | Yes | Yes |
| Leiostomus xanthurus | Spot | Nursery | X | 23,237 | Yes | Yes |
| Lutjanus griseus | Gray snapper | Nursery | X | 5 | Yes | Yes |
| Lutjanus synagris | Lane snapper | Nursery | X | 1 | No | Yes |
| Megalops atlanticus | Tarpon | Nursery | X | 7 | Yes | Yes |
| Menticirrhus saxatilis | Northern kingfish | Nursery | X | 1 | Yes | No |
| Mugil cephalus | Striped mullet | Nursery | X | 3,335 | Yes | Yes |
| Mugil curema | White mullet | Nursery | X | 26 | Yes | Yes |
| Mugil gyrans | Fantail mullet | Nursery | X | 4 | Yes | Yes |
| Mugil spp. | Assorted mullet | Nursery | X | 3,788 | Yes | No |
| Mycteroperca microlepis | Gag | Nursery | X | 10 | No | Yes |
| Opsanus beta | Gulf toadfish | Nursery |  | 35 | No | Yes |
| Orthopristis chrysoptera | Pigfish | Nursery |  | 51 | No | Yes |
| Paralichthys albigutta | Gulf flounder | Nursery | X | 1 | No | Yes |
| Pogonias cromis | Black drum | Nursery | X | 4 | Yes | Yes |
| Sciaenops ocellatus | Red drum | Nursery | X | 34 | Yes | No |
| Scomberomorus maculatus | Spanish mackerel | Nursery | X | 136 | No | Yes |
| Sphyraena barracuda | Great barracuda | Nursery |  | 2 | No | Yes |
| Adinia xenica | Diamond killifish | Resident |  | 3 | Yes | No |
| Bathygobius soporator | Frillfin goby | Resident |  | 12 | Yes | No |
| Cyprinodon variegatus | Sheepshead minnow | Resident |  | 2,973 | Yes | Yes |
| Diapterus plumieri | Striped mojarra | Resident |  | 5 | Yes | Yes |
| Eucinostomus gula | Silver jenny | Resident |  | 724 | Yes | Yes |
| Eucinostomus harengulus | Tidewater mojarra | Resident |  | 1,308 | Yes | Yes |
| Eucinostomus spp. | Assorted mojarra | Resident |  | 16,679 | Yes | Yes |
| Floridichthys carpio | Goldspotted killifish | Resident |  | 2,035 | Yes | Yes |
| Fundulus confluentus | Marsh killifish | Resident |  | 4 | No | Yes |
| Fundulus grandis | Gulf killifish | Resident |  | 1,348 | Yes | Yes |
| Fundulus majalis | Striped killifish | Resident |  | 1,148 | Yes | Yes |
| Fundulus spp. | Assorted killifish | Resident |  | 10 | No | Yes |
| Gambusia holbrooki | Eastern mosquito fish | Resident |  | 14,430 | Yes | Yes |
| Gobiosoma bosc | Naked goby | Resident |  | 13 | Yes | Yes |
| Gobiosoma robustum | Code goby | Resident |  | 21 | Yes | Yes |
| Gobiosoma spp. | Assorted goby | Resident |  | 52 | Yes | Yes |

Table 3.2-1. (Continued)

| Species | Common Name | Guild <br> Name | Selected Taxa | Number Caught | Feb 2002 | Jun 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Heterandria formosa | Least killifish | Resident |  | 31 | No | Yes |
| Lophogobius cyprinoides | Crested goby | Resident |  | 110 | No | Yes |
| Lucania goodei | Bluefin killifish | Resident |  | 22 | No | Yes |
| Lucania parva | Rainwater killifish | Resident |  | 63,313 | Yes | Yes |
| Menidia spp. | Assorted silverside | Resident |  | 18,962 | Yes | Yes |
| Microgobius gulosus | Clown goby | Resident |  | 1,014 | Yes | Yes |
| Poecilia latipinna | Sailfin molly | Resident |  | 7,707 | Yes | Yes |
| Sphoeroides nephelus | Southern puffer | Resident |  | 220 | Yes | Yes |
| Syngnathus floridae | Dusky pipefish | Resident |  | 18 | Yes | Yes |
| Syngnathus louisianae | Chain pipefish | Resident |  | 6 | Yes | Yes |
| Syngnathus scoveli | Gulf pipefish | Resident |  | 110 | Yes | Yes |
| Trinectes maculatus | Hogchoker | Resident |  | 4 | Yes | Yes |
| Aluterus schoepfi | Orange filefish | Transient |  | 2 | No | Yes |
| Anchoa cubana | Cuban anchovy | Transient |  | 1,568 | No | Yes |
| Anchoa hepsetus | Striped anchovy | Transient |  | 433 | Yes | Yes |
| Anchoa mitchilli | Bay anchovy | Transient |  | 116,208 | Yes | Yes |
| Bairdiella chrysoura | Silver perch | Transient |  | 100 | No | Yes |
| Brevoortia spp. | Assorted menhaden | Transient |  | 677 | Yes | Yes |
| Calamus arctifrons | Grass porgy | Transient |  | 2 | Yes | No |
| Callinectes sapidus | Blue crab | Transient | X | 86 | Yes | Yes |
| Chilomycterus schoepfi | Striped burrfish | Transient |  | 37 | No | Yes |
| Clupeidae spp. | Assorted herring | Transient |  | 1,293 | No | Yes |
| Dasyatis say | Bluntnose stingray | Transient |  | 1 | No | Yes |
| Diplodus holbrooki | Spottail pinfish | Transient |  | 16 | No | Yes |
| Haemulon parrai | Sailors choice | Transient |  | 1 | Yes | No |
| Harengula jaguana | Scaled sardine | Transient |  | 15,174 | Yes | Yes |
| Hippocampus zosterae | Dwarf seahorse | Transient |  | 16 | Yes | No |
| Lepomis spp. | Bluegill | Transient |  | 1 | No | Yes |
| Limulus polyphemus | Horseshoe crab | Transient |  | 4 | No | Yes |
| Micropterus salmoides | Largemouth bass | Transient |  | 6 | No | Yes |
| Monacanthus hispidus | Planehead filefish | Transient |  | 2 | Yes | No |
| Oligoplites saurus | Leatherjacket | Transient |  | 94 | No | Yes |
| Selene vomer | Lookdown | Transient |  | 1 | No | Yes |
| Strongylura marina | Atlantic needlefish | Transient |  | 13 | Yes | Yes |
| Strongylura notata | Redfin needlefish | Transient |  | 146 | Yes | Yes |
| Strongylura spp. | Assorted needlefish | Transient |  | 5 | Yes | Yes |
| Strongylura timucu | Timucu | Transient |  | 19 | Yes | Yes |
| Synodus foetens | Inshore lizardfish | Transient |  | 23 | Yes | Yes |
| Urophycis floridana | Southern hake | Transient |  | 1 | Yes | No |
| Total Species: 80 |  |  | Total Catch: | 315,208 | 128,952 | 186,256 |

[^2]Table 3.2-2. Catch statistics for the ten dominant taxa collected in ninety-three 21.3-m standard seine samples (sample area 13,020 $\mathrm{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( 196,333 individuals) represented by that taxon. Percent occurrence (\% occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density.

| Species | Number |  | $\%$ <br> Occur | Density Estimate (animals/100 m${ }^{\text {2 }}$ ) |  |  |  |  | Standard Length (mm) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Anchoa mitchilli | 74,782 | 38.1 | 29.0 | 574.36 | 263.02 | 441.61 | 15,302.86 | 33 | 0.49 | 21 | 54 |
| Lucania parva | 46,068 | 23.5 | 57.0 | 353.82 | 108.80 | 296.53 | 7,474.29 | 23 | 0.25 | 1 | 41 |
| Leiostomus xanthurus | 16,589 | 8.4 | 40.9 | 127.41 | 45.50 | 344.40 | 3,428.57 | 28 | 0.51 | 7 | 165 |
| Eucinostomus spp. | 15,979 | 8.1 | 54.8 | 122.73 | 39.53 | 310.63 | 2,445.71 | 30 | 0.83 | 15 | 330 |
| Lagodon rhomboides | 14,786 | 7.5 | 81.7 | 113.56 | 19.84 | 168.45 | 982.86 | 37 | 0.72 | 9 | 89 |
| Menidia spp. | 9,751 | 5.0 | 64.5 | 74.89 | 18.46 | 237.76 | 1,348.57 | 43 | 0.59 | 16 | 85 |
| Mugil cephalus | 2,977 | 1.5 | 12.9 | 22.86 | 16.56 | 698.58 | 1,462.86 | 47 | 3.43 | 17 | 583 |
| Mugil spp. | 2,645 | 1.3 | 9.7 | 20.31 | 10.34 | 490.95 | 600.00 | 26 | 0.29 | 18 | 63 |
| Floridichthys carpio | 1,645 | 0.8 | 39.8 | 12.63 | 3.27 | 249.75 | 175.00 | 29 | 0.85 | 11 | 62 |
| Cyprinodon variegatus | 1,587 | 0.8 | 19.4 | 12.19 | 5.65 | 447.38 | 372.86 | 33 | 0.90 | 11 | 60 |
| Subtotal | 186,809 | 95.1 | . | . | - | $\cdot$ | . | . | . | 1 | 583 |
| Totals | 196,333 | 100.0 | - | 1,507.93 | 285.26 | 182.43 | 15,997.14 | - | - | 1 | 583 |

Table 3.2-3. Catch statistics for selected taxa collected in ninety-three 21.3-m standard seine samples (sample area $13,020 \mathrm{~m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch (196,333 individuals) represented by that taxon. Percent occurrence (\% occur) is the percentage of samples in which that taxon was collected. $C V$ is the coefficient of variation. Taxa are ranked in order of decreasing mean density.

| Species | Number |  | \% <br> Occur | Density Estimate (animals/100 m${ }^{\text {2 }}$ ) |  |  |  | Standard Length (mm) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Leiostomus xanthurus | 16,589 | 8.4 | 40.9 | 127.41 | 45.50 | 344.40 | 3,428.57 | 28 | 0.51 | 7 | 165 |
| Mugil cephalus | 2,977 | 1.5 | 12.9 | 22.86 | 16.56 | 698.58 | 1,462.86 | 47 | 3.43 | 17 | 583 |
| Mugil spp. | 2,645 | 1.3 | 9.7 | 20.31 | 10.34 | 490.95 | 600.00 | 26 | 0.29 | 18 | 63 |
| Farfantepenaeus duorarum | 331 | 0.2 | 26.9 | 2.54 | 1.34 | 508.84 | 119.29 | 15 | 0.58 | 6 | 33 |
| Scomberomorus maculatus | 135 | 0.1 | 3.2 | 1.04 | 1.01 | 942.84 | 94.29 | 51 | 1.48 | 44 | 67 |
| Cynoscion nebulosus | 53 | 0.0 | 14.0 | 0.41 | 0.15 | 357.93 | 8.57 | 42 | 1.87 | 22 | 86 |
| Callinectes sapidus | 50 | 0.0 | 16.1 | 0.38 | 0.12 | 294.81 | 5.71 | 45 | 4.34 | 14 | 113 |
| Sciaenops ocellatus | 31 | 0.0 | 7.5 | 0.24 | 0.10 | 396.45 | 5.71 | 67 | 3.20 | 49 | 110 |
| Mugil curema | 18 | 0.0 | 3.2 | 0.14 | 0.10 | 727.72 | 9.29 | 55 | 12.83 | 21 | 172 |
| Mycteroperca microlepis | 10 | 0.0 | 2.2 | 0.08 | 0.06 | 793.20 | 5.71 | 114 | 12.31 | 83 | 152 |
| Archosargus probatocephalus | 9 | 0.0 | 7.5 | 0.07 | 0.03 | 404.95 | 2.14 | 60 | 15.34 | 22 | 145 |
| Centropomus undecimalis | 2 | 0.0 | 2.2 | 0.02 | 0.01 | 678.19 | 0.71 | 224 | 17.68 | 199 | 249 |
| Lutjanus griseus | 1 | 0.0 | 1.1 | 0.01 | 0.01 | 964.37 | 0.71 | 80 | 0.00 | 80 | 80 |
| Lutjanus synagris | 1 | 0.0 | 1.1 | 0.01 | 0.01 | 964.37 | 0.71 | 33 | 0.00 | 33 | 33 |
| Menticirrhus saxatilis | 1 | 0.0 | 1.1 | 0.01 | 0.01 | 964.37 | 0.71 | 22 | 0.00 | 22 | 22 |
| Paralichthys albigutta | 1 | 0.0 | 1.1 | 0.01 | 0.01 | 964.37 | 0.71 | 12 | 0.00 | 12 | 12 |
| Pogonias cromis | 1 | 0.0 | 1.1 | 0.01 | 0.01 | 964.37 | 0.71 | 139 | 0.00 | 139 | 139 |
| Totals | 22,855 | 11.6 | 68.8 | 175.54 | 60.74 | 333.70 | 4,891.43 | - | - | 6 | 583 |

Table 3.2-4. Catch statistics for ten dominant taxa collected in twenty-two 21.3-m boat seine samples (sample area $\mathbf{1 , 4 9 6} \mathbf{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( 104,025 individuals) represented by that taxon. Percent occurrence ( $\%$ occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density.

| Species | Number |  | $\begin{gathered} \% \\ \text { Occur } \end{gathered}$ | Density Estimate (animals/100 m${ }^{\text {2 }}$ ) |  |  |  | Standard Length (mm) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Anchoa mitchilli | 41425 | 39.8 | 54.5 | 2,769.05 | 1,722.08 | 291.70 | 31,811.76 | 30 | 0.55 | 21 | 50 |
| Lucania parva | 16863 | 16.2 | 45.5 | 1,127.21 | 550.65 | 229.13 | 9,670.59 | 23 | 0.60 | 14 | 36 |
| Harengula jaguana | 13768 | 13.2 | 13.6 | 920.32 | 915.29 | 466.48 | 20,141.18 | 23 | 0.34 | 20 | 25 |
| Menidia spp. | 8847 | 8.5 | 90.9 | 591.38 | 190.32 | 150.95 | 2,729.41 | 34 | 0.82 | 15 | 99 |
| Gambusia holbrooki | 7784 | 7.5 | 18.2 | 520.32 | 393.16 | 354.42 | 8,364.71 | 20 | 0.46 | 16 | 29 |
| Leiostomus xanthurus | 6380 | 6.1 | 36.4 | 426.47 | 375.33 | 412.79 | 8,282.35 | 28 | 0.77 | 13 | 77 |
| Poecilia latipinna | 1829 | 1.8 | 36.4 | 122.26 | 85.63 | 328.52 | 1,870.59 | 28 | 1.42 | 15 | 67 |
| Anchoa cubana | 1456 | 1.4 | 9.1 | 97.33 | 67.70 | 326.26 | 1,200.00 | 23 | 0.56 | 19 | 29 |
| Mugil spp. | 1081 | 1.0 | 18.2 | 72.26 | 52.36 | 339.89 | 1,129.41 | 30 | 0.45 | 22 | 38 |
| Lagodon rhomboides | 910 | 0.9 | 54.5 | 60.83 | 35.39 | 272.91 | 752.94 | 36 | 2.08 | 15 | 69 |
| Subtotal | 100,343 | 96.5 |  | . | . | . | . | . | . | 1 | 350 |
| Totals | 104,025 | 100.0 | - | 6,953.54 | 2,788.85 | 188.12 | 54,417.65 | - | - | 350 | 350 |

Table 3.2-5. Catch statistics for selected taxa collected in twenty-two 21.3-m boat seine samples (sample area 1,496 $\mathbf{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( 104,025 individuals) represented by that taxon. Percent occurrence ( $\%$ occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density.

| Species | Number |  | $\%$ <br> Occur | Density Estimate (animals/100 m${ }^{\text {2 }}$ ) |  |  |  | Standard Length (mm) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Leiostomus xanthurus | 6,380 | 6.1 | 36.4 | 426.47 | 375.33 | 412.79 | 8,282.35 | 28 | 0.77 | 13 | 77 |
| Mugil spp. | 1,081 | 1.0 | 18.2 | 72.26 | 52.36 | 339.89 | 1,129.41 | 30 | 0.45 | 22 | 38 |
| Mugil cephalus | 282 | 0.3 | 18.2 | 18.85 | 16.53 | 411.33 | 364.71 | 42 | 3.41 | 22 | 83 |
| Mugil gyrans | 4 | 0.0 | 9.1 | 0.27 | 0.21 | 365.47 | 4.41 | 97 | 20.46 | 27 | 129 |
| Lutjanus griseus | 3 | 0.0 | 9.1 | 0.20 | 0.15 | 342.88 | 2.94 | 83 | 20.01 | 48 | 130 |
| Callinectes sapidus | 2 | 0.0 | 9.1 | 0.13 | 0.09 | 323.67 | 1.47 | 34 | 8.13 | 22 | 45 |
| Archosargus probatocephalus | 1 | 0.0 | 4.5 | 0.07 | 0.07 | 469.04 | 1.47 | 350 | 0.00 | 350 | 350 |
| Cynoscion nebulosus | 1 | 0.0 | 4.5 | 0.07 | 0.07 | 469.04 | 1.47 | 59 | 0.00 | 59 | 59 |
| Centropomus undecimalis | 1 | 0.0 | 4.5 | 0.07 | 0.07 | 469.04 | 1.47 | 270 | 0.00 | 270 | 270 |
| Farfantepenaeus duorarum | 1 | 0.0 | 4.5 | 0.07 | 0.07 | 469.04 | 1.47 | 19 | 0.00 | 19 | 19 |
| Mugil curema | 1 | 0.0 | 4.5 | 0.07 | 0.07 | 469.04 | 1.47 | 23 | 0.00 | 23 | 23 |
| Sciaenops ocellatus | 1 | 0.0 | 4.5 | 0.07 | 0.07 | 469.04 | 1.47 | 58 | 0.00 | 58 | 58 |
| Totals | 7,758 | 7.5 | 63.6 | 518.58 | 426.49 | 385.75 | 9,414.71 | - | - | 13 | 350 |

Table 3.2-6. Catch statistics for ten dominant taxa collected in twenty-one 6.1-m standard seine samples (sample area $819 \mathbf{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( $\mathbf{1 3 , 6 0 8}$ individuals) represented by that taxon. Percent occurrence ( $\%$ occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density.

| Species | Number |  | $\%$ <br> Occur | Density Estimate (animals/100 m${ }^{\text {2 }}$ ) |  |  |  | Standard Length (mm) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Gambusia holbrooki | 6,486 | 47.7 | 33.3 | 791.94 | 361.50 | 209.18 | 5,579.49 | 21 | 0.54 | 11 | 33 |
| Poecilia latipinna | 5,173 | 38.0 | 52.4 | 631.62 | 396.98 | 288.02 | 8,287.18 | 34 | 1.27 | 12 | 63 |
| Cyprinodon variegatus | 764 | 5.6 | 57.1 | 93.28 | 39.53 | 194.19 | 761.54 | 28 | 0.99 | 9 | 50 |
| Lucania parva | 333 | 2.4 | 28.6 | 40.66 | 28.14 | 317.21 | 574.36 | 23 | 1.18 | 14 | 34 |
| Menidia spp. | 167 | 1.2 | 52.4 | 20.39 | 7.11 | 159.73 | 92.31 | 44 | 1.26 | 22 | 65 |
| Leiostomus xanthurus | 129 | 0.9 | 9.5 | 15.75 | 15.62 | 454.54 | 328.21 | 26 | 1.15 | 22 | 29 |
| Eucinostomus harengulus | 90 | 0.7 | 19.0 | 10.99 | 9.03 | 376.61 | 189.74 | 56 | 1.98 | 40 | 70 |
| Fundulus majalis | 81 | 0.6 | 28.6 | 9.89 | 6.18 | 286.51 | 128.21 | 31 | 1.13 | 19 | 50 |
| Eucinostomus gula | 74 | 0.5 | 14.3 | 9.04 | 8.65 | 438.96 | 182.05 | 54 | 2.58 | 42 | 66 |
| Mugil cephalus | 72 | 0.5 | 19.0 | 8.79 | 5.53 | 288.38 | 97.44 | 37 | 0.40 | 31 | 48 |
| Subtotal | 13,369 | 98.2 | - | . | . | . | . | . | - | 9 | 330 |
| Totals | 13,608 | 100.0 |  | 1,661.54 | 699.18 | 192.84 | 13,866.67 | - | - | 9 | 330 |

Table 3.2-7. Catch statistics for selected taxa collected in twenty-one $6.1-\mathrm{m}$ standard seine samples (sample area $819 \mathbf{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( 13,608 individuals) represented by that taxon. Percent occurrence ( $\%$ occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density.

| Species | Number | $\%$ <br> Total <br> Catch | $\%$ <br> Occur | Density Estimate (animals/100 m${ }^{\text {2 }}$ ) |  |  |  | Standard Length (mm) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Leiostomus xanthurus | 129 | 0.9 | 9.5 | 15.75 | 15.62 | 454.54 | 328.21 | 26 | 1.15 | 22 | 29 |
| Mugil cephalus | 72 | 0.5 | 19.0 | 8.79 | 5.53 | 288.38 | 97.44 | 37 | 0.40 | 31 | 48 |
| Callinectes sapidus | 33 | 0.2 | 9.5 | 4.03 | 3.90 | 443.89 | 82.05 | 44 | 18.03 | 18 | 69 |
| Centropomus undecimalis | 9 | 0.1 | 9.5 | 1.10 | 0.76 | 317.98 | 12.82 | 64 | 7.75 | 40 | 111 |
| Megalops atlanticus | 7 | 0.1 | 9.5 | 0.85 | 0.60 | 319.37 | 10.26 | 233 | 31.75 | 69 | 330 |
| Mugil curema | 7 | 0.1 | 9.5 | 0.85 | 0.60 | 319.37 | 10.26 | 37 | 1.53 | 29 | 41 |
| Pogonias cromis | 3 | 0.0 | 9.5 | 0.37 | 0.27 | 334.66 | 5.13 | 195 | 33.11 | 151 | 276 |
| Sciaenops ocellatus | 1 | 0.0 | 4.8 | 0.12 | 0.12 | 458.26 | 2.56 | 76 | 0.00 | 76 | 76 |
| Totals | 261 | 1.9 | 42.9 | 31.87 | 19.89 | 286.07 | 410.26 | - | - | 18 | 330 |

Table 3.2-8. Catch statistics for ten dominant taxa collected in twelve 6.1-m boat seine samples (sample area $36 \mathbf{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( 777 individuals) represented by that taxon. Percent occurrence (\% occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density.

| Species | Number |  | $\begin{gathered} \% \\ \text { Occur } \end{gathered}$ | Density Estimate (animals/100 m²) |  |  |  |  | Standard Length (mm) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Lagodon rhomboides | 140 | 18.0 | 16.7 | 388.89 | 270.28 | 240.76 | 2,866.67 | 17 | 1.11 | 11 | 34 |
| Leiostomus xanthurus | 139 | 17.9 | 50.0 | 386.11 | 323.92 | 290.61 | 3,933.33 | 24 | 0.52 | 17 | 35 |
| Menidia spp. | 135 | 17.4 | 25.0 | 375.00 | 198.95 | 183.78 | 1,766.67 | 41 | 0.96 | 35 | 59 |
| Eucinostomus spp. | 88 | 11.3 | 33.3 | 244.44 | 151.78 | 215.09 | 1,366.67 | 30 | 1.06 | 21 | 39 |
| Fundulus grandis | 76 | 9.8 | 33.3 | 211.11 | 127.61 | 209.40 | 1,466.67 | 34 | 2.54 | 15 | 71 |
| Mugil spp. | 62 | 8.0 | 8.3 | 172.22 | 172.22 | 346.41 | 2,066.67 | 30 | 0.45 | 25 | 40 |
| Cyprinodon variegatus | 60 | 7.7 | 25.0 | 166.67 | 119.83 | 249.07 | 1,366.67 | 29 | 1.33 | 19 | 40 |
| Lucania parva | 46 | 5.9 | 33.3 | 127.78 | 80.28 | 217.64 | 900.00 | 23 | 0.54 | 18 | 30 |
| Poecilia latipinna | 12 | 1.5 | 25.0 | 33.33 | 20.52 | 213.20 | 233.33 | 24 | 1.20 | 18 | 32 |
| Farfantepenaeus duorarum | 4 | 0.5 | 16.7 | 11.11 | 7.49 | 233.55 | 66.67 | 16 | 3.02 | 7 | 24 |
| Subtotal | 762 | 98.1 | . | . | . | . | . | . | . | 7 | 155 |
| Totals | 777 | 100.0 | - | 2,158.33 | 399.64 | 64.14 | 4,066.67 | - | - | 7 | 155 |

Table 3.2-9. Catch statistics for selected taxa collected in twelve 6.1-m boat seine samples (sample area $36 \mathbf{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( 777 individuals) represented by that taxon. Percent occurrence (\% occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density.

| Species | Number |  | $\%$ <br> Occur | Density Estimate (animals/ $100 \mathrm{~m}^{2}$ ) |  |  |  | Standard Length (mm) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Leiostomus xanthurus | 139 | 17.9 | 50.0 | 386.11 | 323.92 | 290.61 | 3,933.33 | 24 | 0.52 | 17 | 35 |
| Mugil spp. | 62 | 8.0 | 8.3 | 172.22 | 172.22 | 346.41 | 2,066.67 | 30 | 0.45 | 25 | 40 |
| Farfantepenaeus duorarum | 4 | 0.5 | 16.7 | 11.11 | 7.49 | 233.55 | 66.67 | 16 | 3.02 | 7 | 24 |
| Callinectes sapidus | 1 | 0.1 | 8.3 | 2.78 | 2.78 | 346.41 | 33.33 | 132 | 0.00 | 132 | 132 |
| Sciaenops ocellatus | 1 | 0.1 | 8.3 | 2.78 | 2.78 | 346.41 | 33.33 | 52 | 0.00 | 52 | 52 |
| Totals | 207 | 26.6 | 50.0 | 575.00 | 359.76 | 216.74 | 3,933.33 | - | - | 7 | 132 |

Table 3.2-10. Catch statistics for ten dominant taxa collected in twelve 6.1-m non-standard seine samples (sample aera $332 \mathbf{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( 465 individuals) represented by that taxon. Percent occurrence (\% occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density. Note: Variable sample areas resulted from use of non-standard seine deployment techniques.

| Species | Number |  | $\%$ <br> Occur | Density Estimate (animals/100 m${ }^{\text {2 }}$ ) |  |  |  | Standard Length (mm) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Gambusia holbrooki | 139 | 29.9 | 50.0 | 41.87 | 23.57 | 188.59 | 278.26 | 26 | 1.60 | 7 | 65 |
| Poecilia latipinna | 125 | 26.9 | 41.7 | 37.65 | 23.69 | 185.48 | 221.74 | 37 | 3.04 | 11 | 83 |
| Menidia spp. | 62 | 13.3 | 33.3 | 18.67 | 18.29 | 295.62 | 220.83 | 44 | 2.11 | 14 | 54 |
| Heterandria formosa | 31 | 6.7 | 16.7 | 9.34 | 9.43 | 290.83 | 113.04 | 18 | 0.86 | 12 | 24 |
| Eucinostomus harengulus | 25 | 5.4 | 33.3 | 7.53 | 3.58 | 181.81 | 38.89 | 56 | 2.04 | 40 | 79 |
| Lucania goodei | 22 | 4.7 | 16.7 | 6.63 | 5.48 | 238.14 | 56.52 | 26 | 1.44 | 12 | 35 |
| Lagodon rhomboides | 22 | 4.7 | 33.3 | 6.63 | 4.90 | 219.93 | 58.33 | 48 | 3.95 | 18 | 70 |
| Oligoplites saurus | 12 | 2.6 | 16.7 | 3.61 | 3.80 | 316.23 | 45.83 | 62 | 2.71 | 42 | 78 |
| Micropterus salmoides | 6 | 1.3 | 16.7 | 1.81 | 1.47 | 233.55 | 13.04 | 48 | 4.66 | 25 | 62 |
| Mugil cephalus | 4 | 0.9 | 16.7 | 1.20 | 1.07 | 266.29 | 12.50 | 62 | 3.77 | 55 | 73 |
| Subtotal | 448 | 96.3 | - | - | . | - | - | . | - | 7 | 100 |
| Totals | 465 | 100.0 | - | 140.06 | 59.70 | 132.68 | 682.61 | - | - | 7 | 100 |

Table 3.2-11. Catch statistics for selected taxa collected in twelve $6.1-\mathrm{m}$ non-standard seine samples (sample area $332 \mathbf{m}^{2}$ ) during Sarasota Bay fishery habitat assessment in February and June 2002. Percent total catch (\% total catch) is the percent of the total catch ( 465 individuals) represented by that taxon. Percent occurrence ( $\%$ occur) is the percentage of samples in which that taxon was collected. CV is the coefficient of variation. Taxa are ranked in order of decreasing mean density. Note: Variable sample areas resulted from use of non-standard seine deployment techniques.

| $\overline{\text { Species }}$ | Number | $\%$ <br> Total <br> Catch | $\%$ <br> Occur | Density Estimate (animals/100 m${ }^{\text {2 }}$ ) |  |  |  | Standard Length (mm) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Mean | Stderr | CV | Max | Mean | Stderr | Min | Max |
| Mugil cephalus | 4 | 0.9 | 16.7 | 1.20 | 1.07 | 266.29 | 12.50 | 62 | 3.77 | 55 | 73 |
| Cynoscion nebulosus | 1 | 0.2 | 8.3 | 0.30 | 0.36 | 346.41 | 4.35 | 48 | 0.00 | 48 | 48 |
| Lutjanus griseus | 1 | 0.2 | 8.3 | 0.30 | 0.35 | 346.41 | 4.17 | 100 | 0.00 | 100 | 100 |
| Scomberomorus maculatus | 1 | 0.2 | 8.3 | 0.30 | 0.35 | 346.41 | 4.17 | 50 | 0.00 | 50 | 50 |
| Totals | 7 | 1.5 | 25.0 | 2.11 | 1.74 | 246.11 | 20.83 | - | - | 48 | 100 |

On a catch-per-unit-area (CPUA) basis (i.e., abundance values were converted to density and expressed as the number of animals per $100 \mathrm{~m}^{2}$ ), for 93 samples collected using a 21.3m standard seine, mean density over all taxa was approximately 1,508 animals per $100 \mathrm{~m}^{2}$ (Table 3.2-2). In 22 samples collected using a $21.3-\mathrm{m}$ boat seine technique, mean density over all taxa was approximately 6,954 animals per $100 \mathrm{~m}^{2}$ (Table 3.2-4). In 21 samples collected using a $6.1-\mathrm{m}$ standard seine technique, mean density over all taxa was approximately 1,662 animals per $100 \mathrm{~m}^{2}$ (Table 3.2-6). In 12 samples collected using a 6.1m boat seine technique, mean density over all taxa was approximately 2,158 animals per 100 $\mathrm{m}^{2}$ (Table 3.2-8). In 12 samples collected using a $6.1-\mathrm{m}$ non-standard seine technique, mean density over all taxa was approximately 156 animals per $100 \mathrm{~m}^{2}$ (Table 3.2-10).

Among the five gear deployment techniques used, the number of selected taxa and the percentage abundance of the total catch between gears were not related in direct proportion to the sample size. In ninety-three $21.3-\mathrm{m}$ standard seine samples, 17 selected taxa accounted for $11.6 \%$ of the total catch, and selected taxa occurred in approximately $69 \%$ of samples taken with this gear (Table 3.2-3). In twenty-two $21.3-\mathrm{m}$ boat seine samples, 12 selected taxa accounted for $7.5 \%$ of the total catch, and selected taxa occurred in approximately $64 \%$ of samples taken with this gear (Table 3.2-5). In twenty-one $6.1-\mathrm{m}$ standard seine samples, 8 selected taxa accounted for $1.9 \%$ of the total catch, and selected taxa occurred in approximately $43 \%$ of samples taken with this gear (Table 3.2-7). In twelve $6.1-\mathrm{m}$ boat seine samples, 5 selected taxa accounted for $26.6 \%$ of the total catch, and selected taxa occurred in approximately $50 \%$ of samples taken with this gear (Table 3.2-9). In twelve $6.1-\mathrm{m}$ non-standard seine samples, 4 selected taxa accounted for $1.5 \%$ of the total catch, and selected taxa occurred in approximately $25 \%$ of samples taken with this gear (Table 3.2-11).

An analysis of faunal density on a CPUA basis for selected taxa was made to compare the abundance of selected taxa across different gear deployment techniques. In 93 samples collected using a $21.3-\mathrm{m}$ standard seine, mean density of selected taxa was approximately 176 animals per $100 \mathrm{~m}^{2}$ (Table 3.2-3). In 22 samples collected using a $21.3-\mathrm{m}$ boat seine
technique, mean density of selected taxa was approximately 519 animals per $100 \mathrm{~m}^{2}$ (Table 3.2-5). In 21 samples collected using a $6.1-\mathrm{m}$ standard seine technique, mean density of selected taxa was approximately 32 animals per $100 \mathrm{~m}^{2}$ (Table 3.2-7). In 12 samples collected using a $6.1-\mathrm{m}$ boat seine technique, mean density of selected taxa was approximately 575 animals per $100 \mathrm{~m}^{2}$ (Table 3.2-9). In 12 samples collected using a $6.1-\mathrm{m}$ non-standard seine technique, mean density of selected taxa was approximately two animals per $100 \mathrm{~m}^{2}$ (Table 3.2-11).

Four of the five gear deployment techniques yielded taxa that were only collected by that single gear and deployment technique. Use of the $21.3-\mathrm{m}$ boat seine technique did not result in any species unique to that technique. The $21.3-\mathrm{m}$ standard seine captured 24 taxa not collected by any other technique. These taxa included Aluterus schoepfi, Bairdiella chrysoura, Bathygobius soporator, Chaetodipterus faber, Chilomycterus schoepfi, Diplodus holbrooki, Dasyatis say, Fundulus confluentus, Gobiosoma bosc, Gobiosoma spp., Haemulon parrai, Hippocampus zosterae, Hyporhamphus spp., Limulus polyphemus, Lutjanus synagris, Monacanthus hispidus, Menticirrhus saxatilis, Orthopristis chrysoptera, Paralichthys albigutta, Sphyraena barracuda, Syngnathus floridae, Syngnathus louisianae, Selene vomer and Trinectes maculatus. The $6.1-\mathrm{m}$ standard seine captured 1 taxon (i.e., Megalops atlanticus) not collected by any other technique. The $6.1-\mathrm{m}$ boat seine captured 2 taxa (i.e., Adinia xenica and Urophycis floridana) not collected by any other technique. The 6.1-m non-standard seine captured 4 taxa (i.e., Heterandria formosa, Lucania goodei, Lepomis spp., and Micropterus salmoides) not collected by any other technique.

Dominant taxa (top ten) collected over all gear deployment techniques (Table 3.2-12), in order of greatest mean density (animals per $100 \mathrm{~m}^{2}$ ) and percent total catch (\%), were Anchoa mitchilli (32.1\%), Lucania parva (16.9\%), Gambusia holbrooki (8.0\%), Leiostomus xanthurus (7.4\%), Menidia spp. (7.1\%), Harengula jaguana (6.0\%), Poecilia latipinna (4.9\%), Lagodon rhomboides (4.7\%), Eucinostomus spp. (4.3\%), and Cyprinodon variegatus (1.7\%). Two taxa, Menidia spp. and Mugil spp., ranked in the top ten dominant taxa in all collections regardless of gear type used.

Table 3.2-12. Summary catch statistics for the ten dominant taxa and selected taxa collected by gear type and sampling procedure.

|  | Method |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 21.3-m Standard Seine |  |  | $\begin{gathered} 21.3-\mathrm{m} \text { Boat Seine } \\ \% \end{gathered}$ |  |  | 6.1-m Standard Seine |  |  | 6.1-m Boat Seine \% |  |  | 6.1-m Non-Standard Seine |  |  | Overall |  |  |
|  | Mean <br> Density | \% Total Catch | $\begin{gathered} \% \\ \text { Occur } \end{gathered}$ | Mean Density | Total Catch | $\%$ | Mean <br> Density | \% Total Catch | \% <br> Occur | Mean <br> Density | Total Catch | $\%$ <br> Occur | Mean <br> Density | \% Total Catch | $\begin{gathered} \% \\ \text { Occur } \end{gathered}$ | Mean Density | \% Total <br> Catch | \% <br> Occur |
| All Taxa | 1,507.93 | 100.0 |  | 6,953.54 | 100.0 |  | 1,661.54 | 100.0 |  | 2158.33 | 100.0 |  | 155.86 | 100.0 |  | 2223.3 | 100.0 |  |
| Dominant Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Anchoa mitchilli | 574.36 | 38.1 | 29.0 | 2,769.05 | 39.8 | 54.5 |  |  |  |  |  |  |  |  |  | 714.6 | 36.9 | 25 |
| Lucania parva | 353.82 | 23.5 | 57.0 | 1,127.21 | 16.2 | 45.5 | 40.66 | 2.4 | 28.6 | 127.78 | 5.9 | 33.3 |  |  |  | 375.7 | 20.1 | 46.9 |
| Gambusia holbrooki |  |  |  | 520.32 | 7.5 | 18.2 | 791.94 | 47.7 | 33.3 |  |  |  | 41.87 | 29.9 | 50.0 | 178.8 | 4.6 | 12.5 |
| Leiostomus xanthurus | 127.41 | 8.4 | 40.9 | 426.47 | 6.1 | 36.4 | 15.75 | 0.9 | 9.5 | 386.11 | 17.9 | 50.0 |  |  |  | 163.7 | 7.4 | 33.8 |
| Menidia spp. | 74.89 | 5.0 | 64.5 | 591.38 | 8.5 | 90.9 | 20.39 | 1.2 | 52.4 | 375.00 | 17.4 | 25.0 | 18.67 | 13.3 | 33.3 | 157.3 | 6.0 | 61.3 |
| Harengula jaguana |  |  |  | 920.32 | 13.2 | 13.6 |  |  |  |  |  |  |  |  |  | 132.8 | 4.8 | 6.3 |
| Poecilia latipinna |  |  |  | 122.26 | 1.8 | 36.4 | 631.62 | 38.0 | 52.4 | 33.33 | 1.5 | 25.0 | 37.65 | 26.9 | 41.7 | 108.1 | 2.4 | 21.9 |
| Lagodon rhomboides | 113.56 | 7.5 | 81.7 | 60.83 | 0.9 | 54.5 |  |  |  | 388.89 | 18.0 | 16.7 | 6.63 | 4.7 | 33.3 | 104.6 | 5.0 | 59.4 |
| Eucinostomus spp. | 122.73 | 8.1 | 54.8 |  |  |  |  |  |  | 244.44 | 11.3 | 33.3 |  |  |  | 95.6 | 5.3 | 45.0 |
| Cyprinodon variegatus | 12.19 | 0.8 | 19.4 |  |  |  | 93.28 | 5.6 | 57.1 | 166.67 | 7.7 | 25.0 |  |  |  | 37.0 | 0.9 | 25.6 |
| Mugil spp. | 20.31 | 1.3 | 9.7 | 72.26 | 1.0 | 18.2 |  |  |  | 172.22 | 8.0 | 8.3 |  |  |  | 34.7 | 1.2 | 8.8 |
| Fundulus grandis |  |  |  |  |  |  |  |  |  | 211.11 | 9.8 | 33.3 |  |  |  | 23.4 | 0.4 | 25.0 |
| Mugil cephalus | 22.86 | 1.5 | 12.9 |  |  |  | 8.79 | 0.5 | 19.0 |  |  |  | 1.20 | 0.9 | 16.7 | 17.1 | 1.1 | 13.1 |
| Anchoa cubana |  |  |  | 97.33 | 1.4 | 9.1 |  |  |  |  |  |  |  |  |  | 13.9 | 0.5 | 1.9 |
| Floridichthys carpio | 12.63 | 0.8 | 39.8 |  |  |  |  |  |  |  |  |  |  |  |  | 11.7 | 0.6 | 31.3 |
| Eucinostomus harengulus |  |  |  |  |  |  | 10.99 | 0.7 | 19.0 |  |  |  | 7.53 | 5.4 | 33.3 | 10.4 | 0.4 | 24.4 |
| Fundulus majalis |  |  |  |  |  |  | 9.89 | 0.6 | 28.6 |  |  |  |  |  |  | 6.4 | 0.4 | 15.0 |
| Eucinostomus gula |  |  |  |  |  |  | 9.04 | 0.5 | 14.3 |  |  |  |  |  |  | 4.1 | 0.2 | 22.5 |
| Farfantepenaeus duorarum |  |  |  |  |  |  |  |  |  | 11.11 | 0.5 | 16.7 |  |  |  | 2.3 | 0.1 | 17.5 |
| Heterandria formosa |  |  |  |  |  |  |  |  |  |  |  |  | 9.34 | 6.7 | 16.7 | 0.8 | 0.0 | 1.3 |
| Oligoplites saurus |  |  |  |  |  |  |  |  |  |  |  |  | 3.61 | 2.6 | 16.7 | 0.7 | 0.0 | 16.9 |
| Lucania goodei |  |  |  |  |  |  |  |  |  |  |  |  | 6.63 | 4.7 | 16.7 | 0.6 | 0.0 | 1.3 |
| Micropterus salmoides |  |  |  |  |  |  |  |  |  |  |  |  | 1.81 | 1.3 | 16.7 | 0.2 | 0.0 | 1.3 |
| Subtotal | 1,434.76 | 95.0 |  | 6,707.43 | 96.4 |  | 1,632.35 | 98.2 |  | 2,116.66 | 98.1 |  | 134.94 | 96.3 |  | 2194.5 | 98.3 |  |

Table 3.1-12 (Continued)

|  | Method |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 21.3-m Standard Seine |  |  | 21.3-m Boat Seine |  |  | 6.1-m Standard Seine |  |  | 6.1-m Boat Seine |  |  | 6.1-m Non-Standard Seine |  |  | Overall |  |  |
|  | Mean <br> Density | \% Total Catch | \% <br> Occur | Mean <br> Density | Total Catch | \% <br> Occur | Mean <br> Density | \% Total <br> Catch | $\begin{gathered} \% \\ \text { Occur } \\ \hline \end{gathered}$ | Mean <br> Density | Total Catch | $\begin{gathered} \% \\ \text { Occur } \\ \hline \end{gathered}$ | Mean <br> Density | \% Total <br> Catch | \% <br> Occur | Mean <br> Density | \% Total <br> Catch | $\begin{gathered} \% \\ \text { Occur } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Leiostomus xanthurus | 127.41 | 8.4 | 40.9 | 426.47 | 6.1 | 36.4 | 15.75 | 0.9 | 9.5 | 386.11 | 17.9 | 50.0 |  |  |  | 163.7 | 7.4 | 33.8 |
| Mugil spp. | 20.31 | 1.3 | 9.7 | 72.26 | 1.0 | 18.2 |  |  |  | 172.22 | 8.0 | 8.3 |  |  |  | 34.7 | 1.2 | 8.8 |
| Mugil cephalus | 22.86 | 1.5 | 12.9 | 18.85 | 0.3 | 18.2 | 8.79 |  | 19.0 |  |  |  | 1.20 | 0.9 | 16.7 | 17.1 | 1.1 | 13.1 |
| Farfantepenaeus duorarum | 2.54 | 0.2 | 26.9 | 0.07 | 0.0 | 4.5 |  |  |  | 11.11 | 0.5 | 16.7 |  |  |  | 2.3 | 0.1 | 17.5 |
| Callinectes sapidus | 0.38 | 0.0 | 16.1 | 0.13 | 0.0 | 9.1 | 4.03 |  | 9.5 | 2.78 | 0.1 | 8.3 |  |  |  | 1.0 | 0.0 | 12.5 |
| Scomberomorus maculatus | 1.04 | 0.1 | 3.2 |  |  |  |  |  |  |  |  |  | 0.30 | 0.2 | 8.3 | 0.6 | 0.0 | 2.5 |
| Sciaenops ocellatus | 0.24 | 0.0 | 7.5 | 0.07 | 0.0 | 4.5 | 0.12 |  | 4.8 | 2.78 | 0.1 | 8.3 |  |  |  | 0.4 | 0.0 | 6.3 |
| Cynoscion nebulosus | 0.41 | 0.0 | 14.0 | 0.07 | 0.0 | 4.5 |  |  |  |  |  |  | 0.30 | 0.2 | 8.3 | 0.3 | 0.0 | 9.4 |
| Mugil curema | 0.14 | 0.0 | 3.2 | 0.07 | 0.0 | 4.5 | 0.85 |  | 9.5 |  |  |  |  |  |  | 0.2 | 0.0 | 4.4 |
| Centropomus undecimalis | 0.02 | 0.0 | 2.2 | 0.07 | 0.0 | 4.5 | 1.10 |  | 9.5 |  |  |  |  |  |  | 0.2 | 0.0 | 3.1 |
| Lutjanus griseus | 0.01 | 0.0 | 1.1 | 0.20 | 0.0 | 9.1 |  |  |  |  |  |  | 0.30 | 0.2 | 8.3 | 0.1 | 0.0 | 2.5 |
| Pogonias cromis | 0.01 | 0.0 | 1.1 |  |  |  | 0.37 |  | 9.5 |  |  |  |  |  |  | 0.1 | 0.0 | 1.9 |
| Megalops atlanticus |  |  |  |  |  |  | 0.85 |  | 9.5 |  |  |  |  |  |  | 0.1 | 0.0 | 1.3 |
| Mycteroperca microlepis | 0.08 | 0.0 | 2.2 |  |  |  |  |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 1.3 |
| Archosargus |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| probatocephalus | 0.07 | 0.0 | 7.5 | 0.07 | 0.0 | 4.5 |  |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 5.0 |
| Lutjanus synagris | 0.01 | 0.0 | 1.1 |  |  |  |  |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 0.6 |
| Menticirrhus saxatilis | 0.01 | 0.0 | 1.1 |  |  |  |  |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 0.6 |
| Paralichthys albigutta | 0.01 | 0.0 | 1.1 |  |  |  |  |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 0.6 |
| Mugil gyrans |  |  |  | 0.27 | 0.0 | 9.1 |  |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 1.3 |
| Subtotal | 175.55 | 11.5 | 68.8 | 518.60 | 7.5 | 63.6 | 31.86 | 1.9 | 42.9 | 575.00 | 26.6 | 50.0 | 2.10 | 1.5 | 25.0 | 220.8 | 9.8 |  |

The two most abundant taxa in the $21.3-\mathrm{m}$ standard and boat seine collections were Anchoa mitchilli ( $38.1 \%$ and $39.8 \%$, respectively) and Lucania parva $(23.5 \%$ and $16.2 \%$, respectively). Gambusia holbrooki (47.7\%) and Poecilia latipinna (38.8\%) were the two most abundant taxa in 6.1-m standard seine collections. Lagodon rhomboides (18.0\%) and Leiostomus xanthurus (17.9\%) were the two most abundant taxa in $6.1-\mathrm{m}$ boat seine collections. Gambusia holbrooki (29.9\%) and Poecilia latipinna (26.9\%) were the two most abundant taxa in $6.1-\mathrm{m}$ non-standard seine collections.

The rank order of all taxa in terms of percent occurrence in samples exhibited a different pattern than for percentage abundance. Lagodon rhomboides (81.7\%) and Menidia spp. ( $64.5 \%$ ) were the two most commonly encountered taxa on a percentage occurrence basis in 21.3-m standard seine collections. Menidia spp. (90.0\%) and Anchoa mitchilli (54.3\%) were the two most commonly encountered taxa on a percentage occurrence basis in 21.3-m boat seine collections. Cyprinodon variegatus $(57.1 \%)$ and Menidia spp. $(52.4 \%)$ were the two most commonly encountered taxa on a percentage occurrence basis in $6.1-\mathrm{m}$ standard seine collections. Leiostomus xanthurus (50.0\%) and Eucinostomus spp. (33.3\%) were the two most commonly encountered taxa on a percentage occurrence basis in $6.1-\mathrm{m}$ boat seine collections. Gambusia holbrooki (50.0\%) and Poecilia latipinna (41.7\%) were the two most commonly encountered taxa on a percentage occurrence basis in $6.1-\mathrm{m}$ non-standard seine collections.

Dominant selected taxa (top ten) over all gear deployment techniques (Table 3.2-12), in order of greatest mean density (animals per $100 \mathrm{~m}^{2}$ ) and percent total catch (\%), were Leiostomus xanthurus (163.7/100 m², 7.4\%), Mugil spp. (34.77/100 m², 1.2\%), Mugil cephalus (17.1/100 m², 1.1\%), Farfantepenaeus duorarum (2.3/100 m², 0.1\%), Callinectes sapidus ( $1.0 / 100 \mathrm{~m}^{2},<0.1 \%$ ), Scomberomorus maculates $\left(0.6 / 100 \mathrm{~m}^{2},<0.1 \%\right.$ ), Sciaenops oscellatus ( $0.4 / 100 \mathrm{~m}^{2},<0.1 \%$ ), Cynoscion nebulosus ( $0.3 / 100 \mathrm{~m}^{2},<0.1 \%$ ), Mugil curema $\left(0.2 / 100 \mathrm{~m}^{2},<0.1 \%\right)$, and Centropomus undecimalis $\left(0.2 / 100 \mathrm{~m}^{2},<0.1 \%\right)$. Other selected taxa, in order of decreasing mean density over all gear deployment techniques, were less common in collections and included Megalops atlanticus, Lutjanus griseus, Pogonias romis,

Archosargus probatocephalus, Mycteroperca microlepis, Mugil gyrans, Lutjanus synagris, Menticirrhus saxatilis, Paralichthys albigutta.

Four of the five gear deployment techniques, yielded Leiostomus xanthurus and Mugil spp. (likely Mugil cephalus) as the two most abundant selected taxa. In 21.3-m standard seine collections, Leiostomus xanthurus (8.4\%) and Mugil spp. (1.5\%) were the two most abundant selected taxa. In 21.3-m boat seine collections, Leiostomus xanthurus (6.1\%) and Mugil spp. (likely Mugil cephalus, 1.0\%) were the two most abundant selected taxa. In 6.1m standard seine collections, Leiostomus xanthurus, ( $0.9 \%$ ) and Mugil cephalus, ( $0.5 \%$ ) were the two most abundant selected taxa. In $6.1-\mathrm{m}$ boat seine collections, Leiostomus xanthurus (17.9\%) and Mugil spp. (likely Mugil cephalus, $8.0 \%$ ) were the two most abundant selected taxa. In 6.1-m non-standard seine collections, Mugil cephalus $(0.9 \%)$ and Cynoscion nebulosus ( $0.2 \%$ ) were the two most abundant selected taxa.

Among selected taxa, the rank order of percentage occurrence in samples exhibited a different pattern than that for percentage abundance. Leiostomus xanthurus (40.9\%) and Farfantepenaeus duorarum (26.9\%) were the two most commonly encountered selected taxa on a percentage occurrence basis in 21.3-m standard collections. Leiostomus xanthurus (36.4\%) and Mugil spp. (likely Mugil cephalus, 18.2\%) were the two most commonly encountered taxa on a percentage occurrence basis in 21.3-m boat seine collections. Mugil cephalus ( $19.0 \%$ ) and Leiostomus xanthurus (9.5\%) were the two most commonly encountered taxa on a percentage occurrence basis in $6.1-\mathrm{m}$ standard seine collections. However, several other selected taxa (e.g., Callinectes sapidus, Centropomus undecimalis, Megalops atlanticus, Mugil curema and Pogonias cromis) were equally common in $6.1-\mathrm{m}$ standard seine collections. Leiostomus xanthurus (50.0\%) and Farfantepenaeus duorarum (16.7\%) were the two most commonly encountered taxa on a percentage occurrence basis in 6.1-m boat seine collections. Mugil cephalus (16.7\%) and Cynoscion nebulosus (8.3\%) were the two most commonly encountered taxa on a percentage occurrence basis in 6.1-m non-standard seine collections. However, Lutjanus griseus and Scomberomorus maculatus were equally common in $6.1-\mathrm{m}$ non-standard seine collections.

Data for selected taxa (i.e., commercially or recreationally important species) were pooled among all sites and seasons to calculate mean sizes (i.e., standard length for fishes, postorbital head length for shrimp, and carapace width for crabs), relative abundance and density. Several selected taxa were relatively abundant in Sarasota Bay collections during 2002, occurring in 12 to $50 \%$ of samples and with 50 or more individuals collected. These taxa included Leiostomus xanthurus, Mugil spp. (likely Mugil cephalus), Farfantepenaeus duorarum, Scomberomorus maculatus, Callinectes sapidus, and Cynoscion nebulosus. Other selected taxa were rare in Sarasota Bay collections during 2002, occurring in less than $10 \%$ of the samples and with 34 or fewer (no less than 4 ) individuals collected. These rare selected taxa included Sciaenops ocellatus, Mugil curema, Centropomus undecimalis, Archosargus probatocephalus, Mycteroperca microlepis, Megalops atlanticus, Lutjanus griseus, Mugil gyrans and Pogonias cromis. A few selected taxa were extremely rare in Sarasota Bay collections during 2002 wherein only 1 individual was collected. These extremely rare selected taxa included Lutjanus synagris, Menticirrhus saxatilis and Paralichthys albigutta.

Among frequently occurring selected taxa, a total of 23,237 Leiostomus xanthurus were collected in 54 seine samples. Standard length of Leiostomus xanthurus ranged from 7 to 165 mm with a mean over all gear deployment techniques between 23 to 28 mm . A total of 7,153 Mullet (Mugil spp.) were collected in 44 seine samples. Standard length of Mugil spp ranged from 17 to 583 mm with a mean over all gear deployment techniques between 23 to 62 mm . A total of 336 Farfantepenaeus duorarum were collected in 28 seine samples. Post-orbital head length of Farfantepenaeus duorarum ranged from 6 to 33 mm with a mean over all gear deployment techniques between 15 to 19 mm . A total of 136 Scomberomorus maculatus were collected in 4 seine samples. Standard length of Scomberomorus maculatus ranged from 44 to 67 mm with a mean over all gear deployment techniques between 50 to 51 mm . A total of 86 Callinectes sapidus were collected in 20 seine samples. Carapace width of Callinectes sapidus ranged from 14 to 132 mm with a mean over all gear deployment techniques between 34 to 45 mm . A total of 55 Cynoscion nebulosus were
collected in 15 seine samples. Standard length of Cynoscion nebulosus ranged from 22 to 86 mm with a mean over all gear deployment techniques between 42 to 59 mm .

Among rarely occurring selected taxa, a total of 34 Sciaenops ocellatus were collected in 10 seine samples. Standard length of Sciaenops ocellatus ranged from 49 to 110 mm with a mean of approximately 67 mm collected by four of five gear deployment techniques at nine stations. A total of 26 Mugil curema were collected in 6 seine samples. Standard length of Mugil curema ranged from 21 to 172 mm with a mean between 23 to 55 mm collected by three of five gear deployment techniques at five stations. A total of 12 Centropomus undecimalis were collected in 5 seine samples. Standard length of Centropomus undecimalis ranged from 40 to 270 mm with a mean of approximately 64 mm collected by three of five gear deployment techniques at four stations (C15, C21, C22 and R18). A total of 10 Archosargus probatocephalus were collected in 8 seine samples. Standard length of Archosargus probatocephalus ranged from 22 to 350 mm with a mean of approximately 60 mm collected by the $21.3-\mathrm{m}$ standard or boat seines at six stations. A total of 10 Mycteroperca microlepis were collected in 2 seine samples. Standard length of Mycteroperca microlepis ranged from 83 to 152 mm with a mean of 114 mm collected by the $21.3-\mathrm{m}$ standard seine at only one station (C4--Leffis Key). A total of 7 Megalops atlanticus were collected in 2 seine samples. Standard length of Megalops atlanticus ranged from 69 to 330 mm with a mean of 233 mm collected by the $6.1-\mathrm{m}$ standard seine at only one station (R4--Durante Park). A total of 5 Lutjanus griseus were collected in 4 seine samples. Standard length of Lutjanus griseus ranged from 48 to 130 mm with a mean of approximately 83 mm collected by three of five gear deployment techniques at four stations (C11, R13, R15 and R16). A total of four Mugil gyrans were collected in two samples. Standard length of Mugil gyrans ranged from 27 to 129 mm with a mean of 97 mm collected by the $21.3-\mathrm{m}$ boat seine at only one station (R16--Palmer Point Park). A total of four Pogonias cromis were collected in three samples. Standard length of Pogonias cromis ranged from 139 to 276 mm with a mean of 195 mm collected by the $6.1-\mathrm{m}$ or $21.3-\mathrm{m}$ standard seines at only one station (R7--Sarasota Baywalk).

Extremely rare selected taxa (only 1 individual collected) were represented by Lutjanus synagris collected in one $21.3-\mathrm{m}$ standard seine sample at a station (C4--Leffis Key); Menticirrhus saxatilis collected in one 21.3-m standard seine sample at a station (R9-Sarasota Baywalk); and Paralichthys albigutta collected in one 21.3-m standard seine sample at a station (C2--Tidy Island).

Species ranked in the top 10 by numerical abundance by gear deployment technique were not consistent across gear techniques (Table 3.2-13). Similarly, species ranks by numerical abundance were not similar between 21.3 m seine deployment techniques (i.e., 71 and 44 species in 21.3 m standard or boat seines, respectively). However, species ranks were similar between all 6.1 m seine deployment techniques (i.e., 22,19 and 21 species in 6.1 m standard or boat or non-standard seines, respectively). These results suggest that generally different habitat conditions were sampled by the five gear deployment techniques. By design, each of the five gear deployment techniques was used across two habitat sites (i.e., natural v. restored) and several microhabitat types (i.e., channel, marsh edge, subtidal pool, or seagrass).

Relative abundance plots (Figure 3.2-6) for the total collection (all gears combined) and each gear deployment technique are not similar indicating that the Sarasota Bay fish community was not equally accessible to all gear deployment techniques used in this study. Differences in species rank order and numerical abundance between gear techniques are paralleled by sample size differences between gear type (Table 3.2-13). For example, the 21.3 m standard seine captured nearly $89 \%$ ( 71 of 80 species) of the total fauna in 93 samples. The 21.3 m boat seine captured $55 \%$ ( 44 of 80 species) of the total fauna in 22 samples. The 6.1 m standard seine captured $27 \%$, the 6.1 m boat seine captured $24 \%$, and the 6.1 m non-standard seine captured $26 \%$ of the total fauna in 21,12 and 12 samples, respectively. This result suggests that limitations due to differences in sample size or gear sample area are important in interpreting community structure. It is apparent that the number of rare species represented by less than 100 individuals are proportionately fewer and similar for collections made with any of the 6.1 m seine deployment techniques (i.e.,

Table 3.2-13. List of rank order in relative abundance of all taxa collected by each gear deployment technique during Sarasota Bay fishery habitat assessment in February and June 2002. Separate lists provided for each gear deployment technique and total over all gear types.

| ALL GEAR TOTAL |  |  | 21.3 m STANDARD SEINE |  |  |  | 21.3 m BOAT SEINE |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Rank | Number | Species | Rank | Number | Species | Rank | Number | Species |
| 1 | 116,208 | Anchoa mitchilli | 1 | 74,782 | Anchoa mitchilli | 1 | 41,425 | Anchoa mitchilli |
| 2 | 63,313 | Lucania parva | 2 | 46,068 | Lucania parva | 2 | 16,863 | Lucania parva |
| 3 | 23,237 | Leiostomus xanthurus | 3 | 16,589 | Leiostomus xanthurus | 3 | 13,768 | Harengula jaguana |
| 4 | 18,962 | Menidia spp. | 4 | 15,979 | Eucinostomus spp. | 4 | 8,847 | Menidia spp. |
| 5 | 16,679 | Eucinostomus spp. | 5 | 14,786 | Lagodon rhomboides | 5 | 7,784 | Gambusia holbrooki |
| 6 | 15,890 | Lagodon rhomboides | 6 | 9,751 | Menidia spp. | 6 | 6,380 | Leiostomus xanthurus |
| 7 | 15,174 | Harengula jaguana | 7 | 2,977 | Mugil cephalus | 7 | 1,829 | Poecilia latipinna |
| 8 | 14,430 | Gambusia holbrooki | 8 | 2,645 | Mugil spp. | 8 | 1,456 | Anchoa cubana |
| 9 | 7,707 | Poecilia latipinna | 9 | 1,645 | Floridichthys carpio | 9 | 1,081 | Mugil spp. |
| 10 | 3,788 | Mugil spp. | 10 | 1,587 | Cyprinodon variegatus | 10 | 910 | Lagodon rhomboides |
| 11 | 3,335 | Mugil cephalus | 11 | 1,406 | Harengula jaguana | 11 | 561 | Cyprinodon variegatus |
| 12 | 2,973 | Cyprinodon variegatus | 12 | 1,285 | Clupeidae spp. | 12 | 561 | Eucinostomus spp. |
| 13 | 2,035 | Floridichthys carpio | 13 | 1,046 | Fundulus majalis | 13 | 529 | Eucinostomus harengulus |
| 14 | 1,568 | Anchoa cubana | 14 | 947 | Fundulus grandis | 14 | 452 | Microgobius gulosus |
| 15 | 1,348 | Fundulus grandis | 15 | 676 | Brevoortia spp. | 15 | 368 | Anchoa hepsetus |
| 16 | 1,308 | Eucinostomus harengulus | 16 | 661 | Eucinostomus harengulus | 16 | 334 | Floridichthys carpio |
| 17 | 1,293 | Clupeidae spp. | 17 | 647 | Eucinostomus gula | 17 | 282 | Fundulus grandis |
| 18 | 1,148 | Fundulus majalis | 18 | 568 | Poecilia latipinna | 18 | 282 | Mugil cephalus |
| 19 | 1,014 | Microgobius gulosus | 19 | 561 | Microgobius gulosus | 19 | 87 | Strongylura notata |
| 20 | 724 | Eucinostomus gula | 20 | 331 | Farfantepenaeus duorarum | 20 | 78 | Lophogobius cyprinoides |
| 21 | 677 | Brevoortia spp. | 21 | 219 | Sphoeroides nephelus | 21 | 33 | Opsanus beta |
| 22 | 433 | Anchoa hepsetus | 22 | 135 | Scomberomorus maculatus | 22 | 24 | Syngnathus scovelli |
| 23 | 336 | Farfantepenaeus duorarum | 23 | 112 | Anchoa cubana | 23 | 20 | Fundulus majalis |
| 24 | 220 | Sphoeroides nephelus | 24 | 100 | Bairdiella chrysoura | 24 | 17 | Strongylura timucu |
| 25 | 146 | Strongylura notata | 25 | 84 | Syngnathus scovelli | 25 | 10 | Fundulus spp. |
| 26 | 136 | Scomberomorus maculatus | 26 | 76 | Oligoplites saurus | 26 | 8 | Clupeidae spp. |
| 27 | 110 | Lophogobius cyprinoides | 27 | 65 | Anchoa hepsetus | 27 | 6 | Oligoplites saurus |
| 28 | 110 | Syngnathus scovelli | 28 | 57 | Strongylura notata | 28 | 4 | Mugil gyrans |
| 29 | 100 | Bairdiella chrysoura | 29 | 53 | Cynoscion nebulosus | 29 | 4 | Strongylura marina |
| 30 | 94 | Oligoplites saurus | 30 | 52 | Gobiosoma spp. | 30 | 3 | Lutjanus griseus |
| 31 | 86 | Callinectes sapidus | 31 | 51 | Orthopristis chrysoptera | 31 | 3 | Strongylura spp. |
| 32 | 55 | Cynoscion nebulosus | 32 | 50 | Callinectes sapidus | 32 | 2 | Callinectes sapidus |
| 33 | 52 | Gobiosoma spp. | 33 | 37 | Chilomycterus schoepfi | 33 | 2 | Diapterus plumieri |
| 34 | 51 | Orthopristis chrysoptera | 34 | 32 | Lophogobius cyprinoides | 34 | 2 | Gobiosoma robustum |
| 35 | 37 | Chilomycterus schoepfi | 35 | 31 | Sciaenops ocellatus | 35 | 1 | Achirus lineatus |
| 36 | 35 | Opsanus beta | 36 | 21 | Gambusia holbrooki | 36 | 1 | Archosargus probatocephalus |
| 37 | 34 | Sciaenops ocellatus | 37 | 21 | Synodus foetens | 37 | 1 | Brevoortia spp. |
| 38 | 31 | Heterandria formosa | 38 | 19 | Gobiosoma robustum | 38 | 1 | Cynoscion nebulosus |
| 39 | 26 | Mugil curema | 39 | 18 | Mugil curema | 39 | 1 | Centropomus undecimalis |
| 40 | 23 | Synodus foetens | 40 | 18 | Syngnathus floridae | 40 | 1 | Farfantepenaeus duorarum |
| 41 | 22 | Lucania goodei | 41 | 16 | Diplodus holbrooki | 41 | 1 | Mugil curema |
| 42 | 21 | Gobiosoma robustum | 42 | 16 | Hippocampus zosterae | 42 | 1 | Synodus foetens |
| 43 | 19 | Strongylura timucu | 43 | 13 | Gobiosoma bosc | 43 | 1 | Sphoeroides nephelus |
| 44 | 18 | Syngnathus floridae | 44 | 12 | Bathygobius soporator | 44 | 1 | Sciaenops ocellatus |
| 45 | 16 | Diplodus holbrooki | 45 | 11 | Achirus lineatus |  |  |  |
| 46 | 16 | Hippocampus zosterae | 46 | 10 | Mycteroperca microlepis |  |  |  |
| 47 | 13 | Gobiosoma bosc | 47 | 9 | Archosargus probatocephalus |  |  |  |
| 48 | 13 | Strongylura marina | 48 | 9 | Strongylura marina |  |  |  |
| 49 | 12 | Achirus lineatus | 49 | 6 | Syngnathus louisianae |  |  |  |
| 50 | 12 | Bathygobius soporator | 50 | 4 | Fundulus confluentus |  |  |  |
| 51 | 12 | Centropomus undecimalis | 51 |  | Limulus polyphemus |  |  |  |
| 52 | 10 | Archosargus probatocephalus | 52 | 4 | Trinectes maculatus |  |  |  |
| 53 | 10 | Fundulus spp. | 53 | 3 | Chaetodipterus faber |  |  |  |
| 54 | 10 | Mycteroperca microlepis | 54 | 3 | Diapterus plumieri |  |  |  |
| 55 | 7 | Megalops atlanticus | 55 | 2 | Aluterus schoepfi |  |  |  |
| 56 | 6 | Micropterus salmoides | 56 | 2 | Centropomus undecimalis |  |  |  |
| 57 | 6 | Syngnathus louisianae | 57 | 2 | Hyporhamphus spp. |  |  |  |
| 58 | 5 | Diapterus plumieri | 58 | 2 | Monacanthus hispidus |  |  |  |
| 59 | 5 | Lutjanus griseus | 59 | 2 | Opsanus beta |  |  |  |
| 60 | 5 | Strongylura spp. | 60 | 2 | Sphyraena barracuda |  |  |  |
| 61 | 4 | Fundulus confluentus | 61 | 2 | Strongylura timucu |  |  |  |
| 62 | 4 | Limulus polyphemus | 62 |  | Strongylura spp. |  |  |  |
| 63 | 4 | Mugil gyrans | 63 |  | Calamus arctifrons |  |  |  |
| 64 | 4 | Pogonias cromis | 64 | 1 | Dasyatis say |  |  |  |
| 65 | 4 | Trinectes maculatus | 65 | , | Haemulon parrai |  |  |  |
| 66 | 3 | Adinia xenica | 66 | 1 | Lutjanus griseus |  |  |  |
| 67 | 3 | Chaetodipterus faber | 67 | , | Lutjanus synagris |  |  |  |
| 68 | 2 | Aluterus schoepfi | 68 | 1 | Menticirrhus saxatilis |  |  |  |
| 69 | 2 | Calamus arctifrons | 69 | , | Paralichthys albigutta |  |  |  |
| 70 | 2 | Hyporhamphus spp. | 70 | 1 | Pogonias cromis |  |  |  |
| 71 | 2 | Monacanthus hispidus | 71 | 1 | Selene vomer |  |  |  |
| 72 | 2 | Sphyraena barracuda |  |  |  |  |  |  |
| 73 | 1 | Dasyatis say |  |  |  |  |  |  |
| 74 | 1 | Haemulon parrai |  |  |  |  |  |  |
| 75 | 1 | Lutjanus synagris |  |  |  |  |  |  |
| 76 | , | Lepomis spp. |  |  |  |  |  |  |
| 77 | 1 | Menticirrhus saxatilis |  |  |  |  |  |  |
| 78 | 1 | Paralichthys albigutta |  |  |  |  |  |  |
| 79 | 1 | Selene vomer |  |  |  |  |  |  |
| 80 | 1 | Urophycis floridana |  |  |  |  |  |  |

Table 3.2-13 (Continued)

| 6.1 | m STANDARD SEINE |  |
| ---: | ---: | :--- |
|  |  |  |
| Rank | Number | Species |
| 1 | 6,486 | Gambusia holbrooki |
| 2 | 5,173 | Poecilia latipinna |
| 3 | 764 | Cyprinodon variegatus |
| 4 | 333 | Lucania parva |
| 5 | 167 | Menidia spp. |
| 6 | 129 | Leiostomus xanthurus |
| 7 | 90 | Eucinostomus harengulus |
| 8 | 81 | Fundulus majalis |
| 9 | 74 | Eucinostomus gula |
| 10 | 72 | Mugil cephalus |
| 11 | 53 | Floridichthys carpio |
| 12 | 49 | Eucinostomus spp. |
| 13 | 43 | Fundulus grandis |
| 14 | 33 | Callinectes sapidus |
| 15 | 32 | Lagodon rhomboides |
| 16 | 9 | Centropomus undecimalis |
| 17 | 7 | Megalops atlanticus |
| 18 | 7 | Mugil curema |
| 19 | 3 | Pogonias cromis |
| 20 | 1 | Anchoa mitchilli |
| 21 | 1 | Microgobius gulosus |
| 22 | 1 | Sciaenops ocellatus |

6.1 m BOAT SEINE
$\begin{array}{rrl}\text { Rank } & \text { Number } & \text { Species } \\ 1 & 140 & \text { Lagodon rhomboides } \\ 2 & 139 & \text { Leiostomus xanthurus } \\ 3 & 135 & \text { Menidia spp. } \\ 4 & 88 & \text { Eucinostomus spp. } \\ 5 & 76 & \text { Fundulus grandis } \\ 6 & 62 & \text { Mugil spp. } \\ 7 & 60 & \text { Cyprinodon variegatus } \\ 8 & 46 & \text { Lucania parva } \\ 9 & 12 & \text { Poecilia latipinna } \\ 10 & 4 & \text { Farfantepenaeus duorarum } \\ 11 & 3 & \text { Adinia xenica } \\ 12 & 3 & \text { Eucinostomus harengulus } \\ 13 & 2 & \text { Floridichthys carpio } \\ 14 & 2 & \text { Syngnathus scovelli } \\ 15 & 1 & \text { Callinectes sapidus } \\ 16 & 1 & \text { Fundulus majalis } \\ 17 & 1 & \text { Synodus foetens } \\ 18 & 1 & \text { Sciaenops ocellatus } \\ 19 & 1 & \text { Urophycis floridana } \\ & & \\ & & \end{array}$
6.1 m NON-STANDARD SEINE

| Rank | Number | Species |
| ---: | ---: | :--- |
| 1 | 139 | Gambusia holbrooki |
| 2 | 125 | Poecilia latipinna |
| 3 | 62 | Menidia spp. |
| 4 | 31 | Heterandria formosa |
| 5 | 25 | Eucinostomus harengulus |
| 6 | 22 | Lucania goodei |
| 7 | 22 | Lagodon rhomboides |
| 8 | 12 | Oligoplites saurus |
| 9 | 6 | Micropterus salmoides |
| 10 | 4 | Mugil cephalus |
| 11 | 3 | Eucinostomus gula |
| 12 | 3 | Lucania parva |
| 13 | 2 | Eucinostomus spp. |
| 14 | 2 | Strongylura notata |
| 15 | 1 | Cynoscion nebulosus |
| 16 | 1 | Cyprinodon variegatus |
| 17 | 1 | Calamus arctifrons |
| 18 | 1 | Floridichthys carpio |
| 19 | 1 | Lutjanus griseus |
| 20 | 1 | Lepomis spp. |
| 21 | 1 | Scomberomorus maculatus |

Figure 3.2-6. Relative abundance curves for all taxa in total collection and each gear deployment technique during Sarasota Bay Fishery Habitat Assessment in February and June 2002. The number of individuals collected for each species are plotted in sequence based on rank abundance values shown in Table 1.

steeply-sloped relative abundance plot) than for collections made with the 21.3 m seine deployment techniques. However, all gear deployment techniques yielded collections with high dominance wherein the most numerically dominant species (i.e., typically the top 10 taxa) comprised over $95 \%$ of the total catch.

In order to measure heterogeneity and assess community organization related to how relative abundances vary among different species in the community, an effort was made to fit the logarithmic series by applying a log-normal curve to the data. The logarithmic series implies that the largest number of species are rare in abundance, and the number of species represented by a single specimen are maximal. The dominance-diversity relationship (i.e., relative abundance plots) as previously derived indicates that this model of species abundance may be applicable to the dataset.

A comparison of faunal representation by each gear deployment technique was made using log-normal plots of the number of individuals of a species in geometric classes as a function of the number of species (Figures 3.2-7 and 3.2-8). An approximate log-normal curve was then fit to the plots. The log-normal plots reveal a trimodal curve for the total collection, 21.3 m boat seine and 6.1 m standard seine, and a quadmodal curve for the 21.3 m standard seine collection. Too few individuals were collected in 6.1 m boat or non-standard seines to reveal similar faunal groupings. For three gear techniques with adequate data, the most species rich group is the least abundant in which nursery and transient species predominate over resident species. A middle group consists of separate clusters of predominantly nursery/transient species (lesser occurrence) and resident species (greater occurrence). The least species rich group with the greatest abundance consists of a mix of nursery/transient and resident species in equal proportion. These results consistently show the numerical and spatial dominance of fauna consisting of approximately 11 resident species in the Sarasota Bay system.

Figure 3.2.7. Log normal distribution of taxa collected by all gear deployment techniques (top) and 21.3 m standard seine (bottom) during Sarasota Bay Fishery Habitat Assessment in February and June 2002. Arithmetic classes of number of individuals of a species are derived from abundance.


Log Normal Distribution-21.3m Standard Seine



Figure 3.2.8. Log normal distribution of taxa collected by 21.3 m boat seine (top) and 6.1 m standard seine (bottom) during Sarasota Bay Fishery Habitat Assessment in February and June 2002. Arithmetic classes of number of individuals of a species are derived from abundance values.

Frequently encountered and abundant species comprise the shore fringe resident complex or guild (Table 3.2-1). These species include: Lucania parva, Menidia spp., a Eucinostomus complex consisting of Eucinostomus harengulus and Eucinostomus gula, Gambusia holbrooki, Poecilia latipinna, Floridichthys carpio, Cyprinodon variegatus, a Fundulus complex consisting of Fundulus grandis and Fundulus majalis, and Microgobius gulosus. Numerous species encountered less frequently and less abundantly are classified as either nursery or transient species (Table 3.2-1). Nursery species include most of the selected taxa (i.e., commercially or recreationally important) which utilize the Bay system during portions of their early life history.

The mean density and standard error of selected taxa at each sampling station is provided (Table 3.2-14). For the selected taxa, the highest density was observed at the Bowlees Creek Station C11 (2,433.1/100 $\mathrm{m}^{2}$ ). The highest mean density for a restored station $\left(1,225.1 / 100 \mathrm{~m}^{2}\right)$ occurred at Station R9, Sarasota Baywalk. The lowest average density (no selected species observed) for selected taxa was observed at the Hog Creek Station R12 and the Selby Shoreline Station R14. The lowest density observed at a natural station was Philippi Creek Station C16 with $0.2 / 100 \mathrm{~m}^{2}$. Overall species richness was highest for selected taxa at Palmer Point Station R16 with eight species, followed by Tidy Island Station C2, Leffis Key Station C4, Leffis Key Station R1 and the Sarasota Baywalk Station R7 all with seven species. The lowest observed number of species was again at restored Stations R12, Hog Creek, and R14, Selby Shoreline, where no species were observed. An additional six stations were observed to have only one species.

Table B-2 in Appendix B contains a summary of the mean density and standard error of taxa pooled for each sampling location for all sampling events (e.g., all the stations in the Bowlees Creek area). The highest observed density was found at the Bowlees Creek Site $\left(8,324.9 / 100 \mathrm{~m}^{2}\right)$, with the lowest observed mean density occurred at the Hidden Bay Site (76.9/100 m ${ }^{2}$ ). The highest average species richness occurred at the Palmer Point Site (11.5 species) and the lowest average species richness occurred at the Hog Creek Site (3.0

Table 3.2-14. Mean density ( $\# / 100 \mathrm{~m}^{2}$ ) and standard error ( $\pm$ SE) by selected taxa at each sampling station for all sampling events.
$\left.\begin{array}{llcccccc}\hline \hline & & & & \text { Tidy Island } & \text { Tidy Island } & \text { Tidy Island } & \text { Leffis Key } \\ \text { Site } 2\end{array}\right)$

Table 3.2-14. (Continued)

|  |  | Sister Keys | El Conquistador | El Conquistador | Bowlees Creek | Bowlees Creek | Bowlees |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Creek |  |  |  |  |  |  |  |
|  |  | Site C 7 | Site C 8 | Site C 9 | Site C 10 | Site C 11 | Site C 12 |
| Species name | Common name | $\underline{\text { Mean } \pm \text { SE }}$ | Mean $\pm$ SE | $\underline{\text { Mean } \pm \text { SE }}$ | Mean $\pm$ SE | Mean $\pm$ SE | $\underline{\text { Mean } \pm \text { SE }}$ |
| Leiostomus xanthurus | Spot | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $82.1 \pm 82.1$ | $2.0 \pm 1.1$ | $2082.4 \pm 2066.7$ | $74.8 \pm 74.2$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $350.0 \pm 267.5$ | $0.0 \pm 0.0$ |
| Mugil cephalus | Striped mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $2.0 \pm 1.3$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $1.1 \pm 0.7$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Callinectes sapidus | Blue crab | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $20.5 \pm 20.5$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $0.2 \pm 0.2$ | $2.0 \pm 1.3$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

## Table 3.2-14. (Continued)

|  |  | S Lido Beach | S Lido Lagoon | Phillippi Creek | Phillippi Creek | Heron Lagoon | Heron |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Lagoon |  |  |  |  |  |  |  |
|  |  | Site C 13 | Site C 14 | Site C 15 | Site C 16 | Site C 17 | Site C 18 |
| Species name | Common name | Mean $\pm$ SE | $\underline{\text { Mean } \pm \text { SE }}$ | Mean $\pm$ SE | Mean $\pm$ SE | $\underline{\text { Mean } \pm \text { SE }}$ | $\underline{\text { Mean } \pm \text { SE }}$ |
| Leiostomus xanthurus | Spot | $70.9 \pm 52.6$ | $4.3 \pm 4.3$ | $1.1 \pm 1.1$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil cephalus | Striped mullet | $1.4 \pm 1.4$ | $6.8 \pm 6.8$ | $1.1 \pm 1.1$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $5.9 \pm 5.9$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table 3.2-14. Continued)

| Species name | Common name | N Catfish <br> Site C 19 <br> Mean $\pm$ SE | Bird Key <br> Site C 20 <br> Mean $\pm$ SE | South Creek <br> Site C 21 <br> Mean $\pm$ SE | South Creek <br> Site C 22 <br> Mean $\pm$ SE | Leffis Key <br> Site R 1 <br> Mean $\pm$ SE | Leffis Key Site R 2 Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Leiostomus xanthurus | Spot | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $289.8 \pm 238$ | $25 \pm 16$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $267.3 \pm 154.7$ | $0.0 \pm 0.0$ |
| Mugil cephalus | Striped mullet | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.0 \pm 0.0$ | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $13.2 \pm 7.3$ | $33.3 \pm 19.2$ |
| Callinectes sapidus | Blue crab | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $8.3 \pm 8.3$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $23.6 \pm 23.6$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.4 \pm 0.8$ | $8.3 \pm 8.3$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.7 \pm 2.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

## Table 3.2-14. (Continued)

|  |  | Leffis Key | Durante Park | Durante Park | Durante Park | Sarasota Baywalk | Sarasota |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Baywalk |  |  |  |  |  |  |  |
|  |  | Site R 3 | Site R 4 | Site R 5 | Site R 6 | Site R 7 | Site R 8 |
| Species name | Common name | $\underline{\text { Mean } \pm \text { SE }}$ | $\underline{\text { Mean } \pm \text { SE }}$ | $\underline{\text { Mean } \pm \text { SE }}$ | $\underline{\text { Mean } \pm \text { SE }}$ | Mean $\pm$ SE | $\underline{\text { Mean } \pm \text { SE }}$ |
| Leiostomus xanthurus | Spot | $38.7 \pm 21.8$ | $0.0 \pm 0.0$ | $991.7 \pm 98.6$ | $256.4 \pm 253.6$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Mugil spp. | Assorted mullet | $1.3 \pm 1.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil cephalus | Striped mullet | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $23.7 \pm 15.5$ | $25.6 \pm 24$ |
| Farfantepenaeus duorarum | Pink shrimp | $4.5 \pm 2.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $30.2 \pm 29.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $1.3 \pm 0.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.6 \pm 0.6$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.7 \pm 2.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.1 \pm 1.9$ | $2.6 \pm 2.6$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $4.5 \pm 2.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.1 \pm 1.1$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table 3.2-14. (Continued)


## Table 3.2-14. (Continued)

| Species name | Common name | Selby Shoreline Site R 15 <br> Mean $\pm$ SE | Palmer Point Site R 16 Mean $\pm$ SE | Bird Key <br> Site R 17 <br> Mean $\pm$ SE | Hidden Bay <br> Site R 18 <br> Mean $\pm$ SE | Total by species Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Leiostomus xanthurus | Spot | $0.5 \pm 0.5$ | $158.1 \pm 151.3$ | $27.7 \pm 16.2$ | $0.6 \pm 0.6$ | $163.7 \pm 60.6$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $34.7 \pm 16.8$ |
| Mugil cephalus | Striped mullet | $0.0 \pm 0.0$ | $91.2 \pm 91.2$ | $145.0 \pm 119.0$ | $0.0 \pm 0.0$ | $17.1 \pm 9.9$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.2 \pm 0.2$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.3 \pm 1.1$ |
| Callinectes sapidus | Blue crab | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.7 \pm 0.7$ | $0.6 \pm 0.6$ | $1.0 \pm 0.5$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.6 \pm 0.6$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $1.3 \pm 1.3$ | $0.0 \pm 0.0$ | $0.4 \pm 0.2$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.3 \pm 0.1$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $2.3 \pm 2.3$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.8 \pm 3.4$ | $0.2 \pm 0.1$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Lutjanus griseus | Gray snapper | $0.2 \pm 0.2$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $1.5 \pm 1.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

species). Mean species diversity was also highest at the Palmer Point Site (1.3), whereas the Selby Shoreline Site (0.5) had the lowest average species diversity.

The mean density and standard error for selected taxa were pooled and analyzed for natural and restored sampling locations for all sampling event (Table 3.2-15). Spot (Leiostomus xanthurus) was the most abundant selected taxa at both natural ( $158.5 / 100 \mathrm{~m}^{2}$ ) and restored (170.1/100 $\mathrm{m}^{2}$ ) stations, followed by Mugil spp. (50.8/100 $\mathrm{m}^{2}$ at natural and $14.9 / 100 \mathrm{~m}^{2}$ at restored sites) both stations being the second most common species found. The overall density for selected taxa was slightly higher for the restored stations (230.7/100 $\mathrm{m}^{2}$ compared to $212.2 / 100 \mathrm{~m}^{2}$ ). The overall selected taxa species richness was also greater at the restored stations, 16 species being observed, whereas 10 selected taxa were observed at natural stations.

The mean density and standard error were also evaluated by guild pooled for both natural and restored location for all sampling events. Table 3.2-16 presents the results of the analysis by guild for natural and restored stations. Overall, natural stations had the greatest mean densities for the resident $\left(1,190.5 / 100 \mathrm{~m}^{2}\right)$ and transient $\left(1,196.3 / 100 \mathrm{~m}^{2}\right)$ guilds, with the exception of the nursery $\left(368.5 / 100 \mathrm{~m}^{2}\right.$ at restored sites) guild. The most apparent difference between natural and restored stations was the observed density of the transient species $\left(1,196.3 / 100 \mathrm{~m}^{2}\right.$ at natural sites and $493.0 / 100 \mathrm{~m}^{2}$ at restored sites). The overall mean density was $2,678 / 100 \mathrm{~m}^{2}$ for the natural stations and $1669 / 100 \mathrm{~m}^{2}$ for the restored habitats.

Five extreme density values above $9,670 / 100 \mathrm{~m}^{2}$ were observed for natural stations and five extreme values above $5,579 / 100 \mathrm{~m}^{2}$ were observed for the restored habitat stations. These extreme high values contributed to a non-normal distribution of the mean density data for both habitat types. The data was subsequently $\log$ transformed, however both sets of stations were still non-normally distributed.

Table 3.2-15. Mean density ( $\# / 100 \mathrm{~m}^{2}$ ) and standard error ( $\pm$ SE) for selected taxa pooled by natural ( $n=88$ ) and restored ( $n=72$ ) location for all sampling events.

| Site Type <br> Species name | Natural <br> Mean $\pm$ SE | Restored <br> Mean $\pm \mathrm{SE}$ |  |
| :--- | :--- | :---: | :---: |
|  | Common name |  |  |
| Leiostomus xanthurus | Spot | $158.5 \pm 96.1$ | $170.1 \pm 75.4$ |
| Mugil spp. | Assorted mullet | $50.8 \pm 27.5$ | $14.9 \pm 10.4$ |
| Mugil cephalus | Striped mullet | $1.1 \pm 0.5$ | $36.5 \pm 21.9$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.3 \pm 0.1$ | $4.8 \pm 2.1$ |
| Callinectes sapidus | Blue crab | $1.1 \pm 0.9$ | $0.8 \pm 0.5$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $1.4 \pm 1.3$ |
| Sciaenops ocellatus | Red drum | $0.1 \pm 0.1$ | $0.7 \pm 0.5$ |
| Cynoscion nebulosus | Spotted seatrout | $0.2 \pm 0.2$ | $0.3 \pm 0.3$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.3 \pm 0.3$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.3 \pm 0.2$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Archosargus probatocephalus | Sheepshead | $0.1 \pm 0.1$ | $0.1 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ |  |
| Total Mean Density |  | $212.2 \pm 111.6$ | $230.7 \pm 90.5$ |

Table 3.2-16. Mean density (\#/100 $\mathrm{m}^{2}$ ) and standard error ( $\pm$ SE) by guild pooled for natural and restored sampling locations for all sampling events.

| Site type | Natural | Restored <br> 72 |
| :--- | :---: | :---: |
| Number of Samples | 88 |  |
| Guild Name | Mean $\pm \mathrm{SE}$ | Mean $\pm \mathrm{SE}$ |
|  |  |  |
| Nursery | $291.2 \pm 121.7$ | $368.5 \pm 106.0$ |
| Resident | $1190.5 \pm 275.5$ | $807.6 \pm 222.4$ |
| Transient | $1196.3 \pm 629.9$ | $493.0 \pm 268.6$ |

For the selected taxa, the mean overall density was $212.1 / 100 \mathrm{~m}^{2}$ for the natural habitats and $231.2 / 100 \mathrm{~m}^{2}$ for the restored habitats. Within the established guilds, the overall mean densities were $291.2 / 100 \mathrm{~m}^{2}, 1,190.7 / 100 \mathrm{~m}^{2}$ and $1,196.3 / 100 \mathrm{~m}^{2}$ for the nursery, resident and transient guilds, respectively, within the natural habitats and $368.5 / 100 \mathrm{~m}^{2}$, $807.6 / 100 \mathrm{~m}^{2}$, and $493.0 / 100 \mathrm{~m}^{2}$ for the nursery, resident and transient guilds, respectively, within the restored habitats.

The fishery sampling stations were further categorized into four microhabitat types: channel, marsh edge, subtidal pool and seagrass. These microhabitat types more closely resembled the habitat types actually sampled using the FMRI FIM methodology and represent, in many cases, the microhabitat type adjacent to the shoreline targeted. Tables 3.2-17 and 3.2-18 present the results for all taxa and selected taxa, respectively, for the four-microhabitat types divided amongst natural and restored sites. The breakdown of stations included within each microhabitat type is provided in Table 2.3-1.

For the natural monitoring sites, the greatest mean density $\left(6,363.0 / 100 \mathrm{~m}^{2}\right)$ was found for the channel microhabitat, followed by seagrass ( $2996.9 / 100 \mathrm{~m}^{2}$ ), marsh edge (925.9/100 $\mathrm{m}^{2}$ ) and subtidal pool $466.6 / 100 \mathrm{~m}^{2}$ ), respectively. The higher mean densities observed for the natural channel and seagrass microhabitats reflect the presence of a large number of schooling species, such as the bay anchovy. For the restored monitoring sites, the seagrass exhibited the greatest overall mean density $\left(2,617.7 / 100 \mathrm{~m}^{2}\right)$ followed by the subtidal pool $\left(1,780.0 / 100 \mathrm{~m}^{2}\right)$, marsh edge $\left(1,427.5 / 100 \mathrm{~m}^{2}\right)$ and channel $\left(1,320.6 / 100 \mathrm{~m}^{2}\right)$ microhabitats, respectively. The mean density of taxa for the natural seagrass $\left(2,996.9 / 100 \mathrm{~m}^{2}\right)$ site is only slightly higher than the mean density for the restored site $\left(2617.7 / 100 \mathrm{~m}^{2}\right)$, whereas the natural channel sites $\left(6,363.0 / 100 \mathrm{~m}^{2}\right)$ have a significantly higher density than the restored channel sites (1,320.6/100 $\mathrm{m}^{2}$ ). Conversely, the restored site subtidal pool $\left(1,780.0 / 100 \mathrm{~m}^{2}\right)$ and marsh edge $\left(1,427.5 / 100 \mathrm{~m}^{2}\right)$ microhabitats have a greater mean density than the natural sites $\left(466.6 / 100 \mathrm{~m}^{2}\right.$ and $925.9 / 100 \mathrm{~m}^{2}$ for subtidal pool and marsh edge, respectively).

Table 3.2-17. Summary of mean density $\left(\# / 100 \mathrm{~m}^{2}\right)$ by taxa within four microhabitat types at natural and restored sites.

| Species Name | Microhabitat Type <br> Common Name | Channel |  |  |  | Marsh Edge |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Natural |  | Restored |  | Natural |  | Restored |  |
|  |  | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error |
| Anchoa mitchilli | Bay anchovy | 3747.9 | 3675.8 | 0.0 | 0.0 | 109.5 | 60.3 | 800.3 | 770.1 |
| Lucania parva | Rainwater killifish | 72.0 | 67.4 | 94.0 | 61.2 | 113.4 | 94.1 | 9.5 | 9.2 |
| Gambusia holbrooki | Eastern mosquito fish | 23.1 | 23.1 | 0.0 | 0.0 | 0.6 | 0.6 | 140.6 | 131.4 |
| Leiostomus xanthurus | Spot | 546.7 | 512.5 | 254.3 | 245.8 | 51.0 | 40.9 | 114.8 | 81.2 |
| Menidia spp. | Assorted silverside | 128.1 | 83.4 | 303.4 | 166.7 | 302.6 | 263.4 | 85.7 | 76.5 |
| Harengula jaguana | Scaled sardine | 1264.7 | 1264.7 | 0.0 | 0.0 | 1.0 | 1.0 | 0.9 | 0.6 |
| Poecilia latipinna | Sailfin molly | 35.7 | 30.7 | 20.8 | 13.8 | 1.1 | 1.0 | 90.6 | 58.0 |
| Lagodon rhomboides | Pinfish | 79.5 | 48.3 | 296.1 | 290.2 | 23.2 | 13.5 | 32.7 | 30.3 |
| Eucinostomus spp. | Assorted mojarra | 39.0 | 28.1 | 191.0 | 164.4 | 209.2 | 139.0 | 14.0 | 9.7 |
| Cyprinodon variegatus | Sheepshead minnow | 50.2 | 29.6 | 87.5 | 84.7 | 4.3 | 4.3 | 69.8 | 50.1 |
| Mugil spp. | Assorted mullet | 99.4 | 84.3 | 0.0 | 0.0 | 36.3 | 29.1 | 0.0 | 0.0 |
| Fundulus grandis | Gulf killifish | 25.4 | 23.0 | 42.5 | 41.4 | 0.7 | 0.3 | 4.5 | 3.3 |
| Mugil cephalus | Striped mullet | 1.4 | 1.4 | 1.0 | 1.0 | 1.8 | 1.5 | 23.4 | 17.7 |
| Anchoa cubana | Cuban anchovy | 133.8 | 133.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Floridichthys carpio | Goldspotted killifish | 22.1 | 12.7 | 0.0 | 0.0 | 10.9 | 6.7 | 1.2 | 1.2 |
| Eucinostomus harengulus | Tidewater mojarra | 24.7 | 24.4 | 10.5 | 5.2 | 23.6 | 15.2 | 15.2 | 11.4 |
| Microgobius gulosus | Clown goby | 6.3 | 6.3 | 0.0 | 0.0 | 10.7 | 5.0 | 2.4 | 2.4 |
| Fundulus majalis | Striped killifish | 2.7 | 1.6 | 0.0 | 0.0 | 7.4 | 6.7 | 8.7 | 8.6 |
| Clupeidae spp. | Assorted herring | 0.7 | 0.7 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| Eucinostomus gula | Silver jenny | 0.0 | 0.0 | 0.5 | 0.5 | 4.6 | 3.3 | 4.7 | 4.2 |
| Anchoa hepsetus | Striped anchovy | 33.8 | 33.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Brevoortia spp. | Assorted menhaden | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| Farfantepenaeus duorarum | Pink shrimp | 0.0 | 0.0 | 8.3 | 8.3 | 0.1 | 0.0 | 0.7 | 0.6 |
| Strongylura notata | Redfin needlefish | 2.6 | 1.7 | 0.5 | 0.5 | 1.4 | 0.9 | 3.9 | 3.5 |
| Syngnathus scovelli | Gulf pipefish | 0.0 | 0.0 | 4.2 | 4.2 | 0.9 | 0.8 | 0.2 | 0.2 |
| Sphoeroides nephelus | Southern puffer | 0.0 | 0.0 | 0.0 | 0.0 | 1.0 | 0.9 | 0.4 | 0.3 |
| Callinectes sapidus | Blue crab | 0.4 | 0.3 | 2.2 | 2.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| Fundulus spp. | Assorted killifish | 0.9 | 0.9 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lophogobius cyprinoides | Crested goby | 1.4 | 1.4 | 0.0 | 0.0 | 5.0 | 5.0 | 0.0 | 0.0 |
| Heterandria formosa | Least killifish | 8.4 | 8.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Oligoplites saurus | Leatherjacket | 0.3 | 0.3 | 3.1 | 3.1 | 0.2 | 0.1 | 0.2 | 0.1 |
| Scomberomorus maculatus | Spanish mackerel | 0.0 | 0.0 | 0.3 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 |
| Adinia xenica | Diamond killifish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lucania goodei | Bluefin killifish | 6.0 | 6.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Sciaenops ocellatus | Red drum | 0.0 | 0.0 | 2.1 | 2.1 | 0.0 | 0.0 | 0.1 | 0.1 |
| Opsanus beta | Gulf toadfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| Synodus foetens | Inshore lizardfish | 0.0 | 0.0 | 2.1 | 2.1 | 0.0 | 0.0 | 0.1 | 0.1 |
| Cynoscion nebulosus | Spotted seatrout | 0.3 | 0.3 | 0.0 | 0.0 | 0.3 | 0.2 | 0.0 | 0.0 |
| Gobiosoma spp. | Assorted goby | 0.7 | 0.7 | 0.0 | 0.0 | 0.4 | 0.4 | 0.0 | 0.0 |
| Orthopristis chrysoptera | Pigfish | 0.0 | 0.0 | 0.0 | 0.0 | 1.1 | 0.7 | 0.4 | 0.4 |
| Urophycis floridana | Southern hake | 0.0 | 0.0 | 2.1 | 2.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| Mugil curema | White mullet | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.5 |
| Bairdiella chrysoura | Silver perch | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |

[^3]Table 3.2-17. (Continued)

| Species Name | Microhabitat Type <br> Common name | Channel |  |  |  | Marsh Edge |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Natural |  | Restored |  | Natural |  | Restored |  |
|  |  | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error |
| Strongylura timиси | Timucu | 1.6 | 1.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Chilomycterus schoepfi | Striped burrfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.8 | 0.8 | 0.0 | 0.0 |
| Micropterus salmoides | Largemouth bass | 1.6 | 1.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Centropomus undecimalis | Snook | 0.0 | 0.0 | 1.4 | 1.4 | 0.1 | 0.1 | 0.0 | 0.0 |
| Megalops atlanticus | Tarpon | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.9 | 0.9 |
| Gobiosoma robustum | Code goby | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.6 | 0.1 | 0.1 |
| Syngnathus floridae | Dusky pipefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.6 | 0.0 | 0.0 |
| Strongylura marina | Atlantic needlefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| Diplodus holbrooki | Spottail pinfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Hippocampus zosterae | Dwarf seahorse | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.6 | 0.0 | 0.0 |
| Achirus lineatus | Lined sole | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.1 | 0.1 | 0.1 |
| Lutjanus griseus | Gray snapper | 0.2 | 0.2 | 0.3 | 0.3 | 0.0 | 0.0 | 0.1 | 0.1 |
| Gobiosoma bosc | Naked goby | 0.1 | 0.1 | 0.0 | 0.0 | 0.2 | 0.2 | 0.0 | 0.0 |
| Bathygobius soporator | Frillfin goby | 0.5 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Pogonias cromis | Black drum | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Archosargus probatocephalus | Sheepshead | 0.1 | 0.1 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| Mycteroperca microlepis | Gag | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Mugil gyrans | Fantail mullet | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.3 |
| Strongylura spp. | Assorted needlefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.2 |
| Diapterus plumieri | Striped mojarra | 0.1 | 0.1 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| Calamus arctifrons | Grass porgy | 0.0 | 0.0 | 0.3 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lepomis spp. | Bluegill | 0.3 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Syngnathus louisianae | Chain pipefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Trinectes maculatus | Hogchoker | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Fundulus confluentus | Marsh killifish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Limulus polyphemus | Horseshoe crab | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Chaetodipterus faber | Atlantic spadefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Aluterus schoepfi | Orange filefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Hyporhamphus spp. | Halfbeak (juv) | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Monacanthus hispidus | Planehead filefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Sphyraena barracuda | Great barracuda | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Dasyatis say | Bluntnose stingray | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Haemulon parrai | Sailors choice | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lutjanus synagris | Lane snapper | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Menticirrhus saxatilis | Northern kingfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Paralichthys albigutta | Gulf flounder | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Selene vomer | Lookdown | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  | Total Mean Density | 6363.0 |  | 1328.6 |  | 925.9 |  | 1427.5 |  |
|  | Standard Error of Means | 49.5 |  | 6.6 |  | 5.0 |  | 10.3 |  |
|  | Total Species Richness | 40 |  | 23 |  | 42 |  | 38 |  |

STD Error: Standard Error of Means

Table 3.2-17. (Continued)

| Species Name | Microhabitat Type <br> Common name | Subtidal Pool |  |  |  | Seagrass |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Natural |  | Restored |  | Natural |  | Restored |  |
|  |  | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error |
| Anchoa mitchilli | Bay anchovy | 1.8 | 1.8 | 15.4 | 10.2 | 488.8 | 424.6 | 1143.8 | 576.5 |
| Lucania parva | Rainwater killifish | 11.4 | 6.0 | 11.7 | 9.8 | 1410.6 | 668.8 | 240.0 | 100.7 |
| Gambusia holbrooki | Eastern mosquito fish | 0.1 | 0.1 | 873.1 | 873.1 | 318.0 | 318.0 | 0.0 | 0.0 |
| Leiostomus xanthurus | Spot | 35.9 | 35.3 | 9.7 | 9.7 | 100.2 | 56.0 | 357.9 | 154.2 |
| Menidia spp. | Assorted silverside | 11.4 | 10.8 | 51.8 | 35.8 | 200.3 | 106.4 | 141.6 | 40.1 |
| Harengula jaguana | Scaled sardine | 55.7 | 55.7 | 0.0 | 0.0 | 1.6 | 1.6 | 1.8 | 1.4 |
| Poecilia latipinna | Sailfin molly | 4.3 | 4.1 | 719.4 | 715.8 | 73.2 | 72.8 | 21.1 | 16.8 |
| Lagodon rhomboides | Pinfish | 10.9 | 9.5 | 5.7 | 4.8 | 136.7 | 42.1 | 276.7 | 68.9 |
| Eucinostomus spp. | Assorted mojarra | 28.6 | 24.2 | 16.2 | 12.4 | 142.1 | 90.9 | 82.8 | 28.4 |
| Cyprinodon variegatus | Sheepshead minnow | 37.8 | 37.4 | 17.0 | 9.8 | 17.3 | 11.5 | 46.1 | 36.8 |
| Mugil spp. | Assorted mullet | 129.2 | 129.2 | 0.3 | 0.3 | 2.5 | 2.4 | 66.8 | 53.5 |
| Fundulus grandis | Gulf killifish | 116.8 | 116.6 | 2.0 | 1.2 | 9.7 | 5.8 | 18.7 | 6.6 |
| Mugil cephalus | Striped mullet | 0.2 | 0.1 | 6.1 | 5.9 | 1.6 | 0.9 | 127.8 | 71.7 |
| Anchoa cubana | Cuban anchovy | 0.0 | 0.0 | 0.0 | 0.0 | 2.2 | 2.2 | 0.0 | 0.0 |
| Floridichthys carpio | Goldspotted killifish | 4.7 | 4.0 | 2.0 | 1.2 | 27.9 | 7.7 | 10.6 | 5.3 |
| Eucinostomus harengulus | Tidewater mojarra | 0.4 | 0.3 | 19.1 | 12.3 | 0.1 | 0.1 | 0.5 | 0.3 |
| Microgobius gulosus | Clown goby | 2.3 | 1.5 | 3.5 | 3.5 | 14.3 | 14.0 | 6.0 | 4.8 |
| Fundulus majalis | Striped killifish | 2.1 | 2.1 | 5.6 | 3.3 | 11.3 | 9.1 | 8.5 | 5.5 |
| Clupeidae spp. | Assorted herring | 0.0 | 0.0 | 0.5 | 0.5 | 23.1 | 19.5 | 4.6 | 3.5 |
| Eucinostomus gula | Silver jenny | 0.6 | 0.6 | 16.8 | 10.8 | 3.8 | 2.2 | 3.1 | 1.2 |
| Anchoa hepsetus | Striped anchovy | 2.1 | 2.1 | 0.2 | 0.2 | 0.3 | 0.3 | 0.0 | 0.0 |
| Brevoortia spp. | Assorted menhaden | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 30.0 | 24.0 |
| Farfantepenaeus duorarum | Pink shrimp | 0.3 | 0.3 | 1.2 | 1.1 | 0.6 | 0.2 | 11.1 | 5.9 |
| Strongylura notata | Redfin needlefish | 0.1 | 0.1 | 0.1 | 0.1 | 0.3 | 0.2 | 0.5 | 0.1 |
| Syngnathus scovelli | Gulf pipefish | 0.2 | 0.2 | 0.0 | 0.0 | 1.4 | 0.9 | 1.2 | 0.6 |
| Sphoeroides nephelus | Southern puffer | 1.2 | 1.1 | 0.1 | 0.1 | 1.6 | 1.2 | 3.1 | 1.4 |
| Callinectes sapidus | Blue crab | 0.1 | 0.1 | 0.4 | 0.3 | 2.6 | 2.2 | 0.7 | 0.2 |
| Fundulus spp. | Assorted killifish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lophogobius cyprinoides | Crested goby | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 0.4 | 0.0 | 0.0 |
| Heterandria formosa | Least killifish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Oligoplites saurus | Leatherjacket | 0.3 | 0.2 | 0.5 | 0.5 | 0.6 | 0.3 | 1.0 | 0.3 |
| Scomberomorus maculatus | Spanish mackerel | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 5.9 | 4.7 |
| Adinia xenica | Diamond killifish | 6.3 | 6.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lucania goodei | Bluefin killifish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Sciaenops ocellatus | Red drum | 0.0 | 0.0 | 0.3 | 0.2 | 0.2 | 0.2 | 0.8 | 0.3 |
| Opsanus beta | Gulf toadfish | 0.0 | 0.0 | 0.0 | 0.0 | 1.3 | 1.3 | 0.0 | 0.0 |
| Synodus foetens | Inshore lizardfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.2 | 0.3 | 0.1 |
| Cynoscion nebulosus | Spotted seatrout | 0.5 | 0.5 | 0.0 | 0.0 | 0.1 | 0.0 | 1.4 | 0.6 |
| Gobiosoma spp. | Assorted goby | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 1.0 | 0.8 |
| Orthopristis chrysoptera | Pigfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.1 | 0.0 | 0.0 |
| Urophycis floridana | Southern hake | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Mugil curema | White mullet | 0.2 | 0.2 | 0.5 | 0.5 | 0.0 | 0.0 | 0.6 | 0.5 |
| Bairdiella chrysoura | Silver perch | 0.0 | 0.0 | 0.0 | 0.0 | 0.8 | 0.8 | 0.0 | 0.0 |

Table 3.2-17. (Continued)

| Species Name | Microhabitat Type <br> Common name | Subtidal Pool |  |  |  | Seagrass |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Natural |  | Restored |  | Natural |  | Restored |  |
|  |  | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error |
| Strongylura timucu | Timucu | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Chilomycterus schoepfi | Striped burrfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.4 | 0.3 |
| Micropterus salmoides | Largemouth bass | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Centropomus undecimalis | Snook | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Megalops atlanticus | Tarpon | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Gobiosoma robustum | Code goby | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Syngnathus floridae | Dusky pipefish | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Strongylura marina | Atlantic needlefish | 0.2 | 0.2 | 0.0 | 0.0 | 0.2 | 0.1 | 0.0 | 0.0 |
| Diplodus holbrooki | Spottail pinfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.3 | 0.0 | 0.0 |
| Hippocampus zosterae | Dwarf seahorse | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Achirus lineatus | Lined sole | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| Lutjanus griseus | Gray snapper | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Gobiosoma bosc | Naked goby | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.1 |
| Bathygobius soporator | Frillfin goby | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Pogonias cromis | Black drum | 0.0 | 0.0 | 0.5 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 |
| Archosargus probatocephalus | Sheepshead | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.1 | 0.1 |
| Mycteroperca microlepis | Gag | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.2 | 0.0 | 0.0 |
| Mugil gyrans | Fantail mullet | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Strongylura spp. | Assorted needlefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Diapterus plumieri | Striped mojarra | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Calamus arctifrons | Grass porgy | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lepomis spp. | Bluegill | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Syngnathus louisianae | Chain pipefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.2 |
| Trinectes maculatus | Hogchoker | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Fundulus confluentus | Marsh killifish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.1 |
| Limulus polyphemus | Horseshoe crab | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| Chaetodipterus faber | Atlantic spadefish | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Aluterus schoepfi | Orange filefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Hyporhamphus spp. | Halfbeak (juv) | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Monacanthus hispidus | Planehead filefish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Sphyraena barracuda | Great barracuda | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Dasyatis say | Bluntnose stingray | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Haemulon parrai | Sailors choice | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lutjanus synagris | Lane snapper | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Menticirrhus saxatilis | Northern kingfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Paralichthys albigutta | Gulf flounder | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Selene vomer | Lookdown | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  | Total Mean Density | 466.6 |  | 1780.0 |  | 2996.9 |  | 2617.7 |  |
|  | Standard Error of Means | 2.3 |  | 14 |  | 19.1 |  | 15.6 |  |
|  | Total Species Richness | 40 |  | 33 |  | 48 |  | 46 |  |

STD Error: Standard Error of Means

Table 3.2-18 Mean dnsity $\left(\# 100 \mathrm{~m}^{2}\right)$ of slected txa within fur microhabitat types at natural and restored sites.

| Species Name | Microhabitat Type <br> Common name | Channel |  |  |  | Marsh Edge |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Natural |  | Restored |  | Natural |  | Restored |  |
|  |  | Mean <br> Density | STD Error | Mean <br> Density | STD Error | Mean Density | STD Error | Mean <br> Density | STD Error |
| Leiostomus xanthurus | Spot | 546.7 | 512.5 | 254.3 | 245.8 | 51.0 | 40.9 | 114.8 | 81.2 |
| Mugil spp. | Assorted mullet | 99.4 | 84.3 | 0.0 | 0.0 | 36.3 | 29.1 | 0.0 | 0.0 |
| Mugil cephalus | Striped mullet | 1.4 | 1.4 | 1.0 | 1.0 | 1.8 | 1.5 | 23.4 | 17.7 |
| Farfantepenaeus duorarum | Pink shrimp | 0.0 | 0.0 | 8.3 | 8.3 | 0.1 | 0.0 | 0.7 | 0.6 |
| Callinectes sapidus | Blue crab | 0.4 | 0.3 | 2.2 | 2.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| Scomberomorus maculatus | Spanish mackerel | 0.0 | 0.0 | 0.3 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 |
| Sciaenops ocellatus | Red drum | 0.0 | 0.0 | 2.1 | 2.1 | 0.0 | 0.0 | 0.1 | 0.1 |
| Cynoscion nebulosus | Spotted seatrout | 0.3 | 0.3 | 0.0 | 0.0 | 0.3 | 0.2 | 0.0 | 0.0 |
| Mugil curema | White mullet | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.5 |
| Centropomus undecimalis | Snook | 0.0 | 0.0 | 1.4 | 1.4 | 0.1 | 0.1 | 0.0 | 0.0 |
| Megalops atlanticus | Tarpon | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.9 | 0.9 |
| Lutjanus griseus | Gray snapper | 0.2 | 0.2 | 0.3 | 0.3 | 0.0 | 0.0 | 0.1 | 0.1 |
| Pogonias cromis | Black drum | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Archosargus probatocephalus | Sheepshead | 0.1 | 0.1 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| Mycteroperca microlepis | Gag | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Mugil gyrans | Fantail mullet | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.3 |
| Lutjanus synagris | Lane snapper | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Menticirrhus saxatilis | Northern kingfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Paralichthys albigutta | Gulf flounder | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Total Mean Density |  | 648.5 |  | 270.0 |  | 89.7 |  | 140.9 |  |
| Standard Error of Means |  | 27.1 |  | 12.9 |  | 2.6 |  | 4.3 |  |
| Total Species Richness |  | 8 |  | 8 |  | 10 |  | 10 |  |

STD Error: Standard Error of Means

Table 3.2-18. (Continued)

| Species Name | Microhabitat Type <br> Common name | Subtidal Pool |  |  |  | Seagrass |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Natural |  | Restored |  | Natural |  | Restored |  |
|  |  | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error | Mean Density | STD Error |
| Leiostomus xanthurus | Spot | 35.9 | 35.3 | 9.7 | 9.7 | 100.2 | 56.0 | 357.9 | 154.2 |
| Mugil spp. | Assorted mullet | 129.2 | 129.2 | 0.3 | 0.3 | 2.5 | 2.4 | 66.8 | 53.5 |
| Mugil cephalus | Striped mullet | 0.2 | 0.1 | 6.1 | 5.9 | 1.6 | 0.9 | 127.8 | 71.7 |
| Farfantepenaeus duorarum | Pink shrimp | 0.3 | 0.3 | 1.2 | 1.1 | 0.6 | 0.2 | 11.1 | 5.9 |
| Callinectes sapidus | Blue crab | 0.1 | 0.1 | 0.4 | 0.3 | 2.6 | 2.2 | 0.7 | 0.2 |
| Scomberomorus maculatus | Spanish mackerel | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 5.9 | 4.7 |
| Sciaenops ocellatus | Red drum | 0.0 | 0.0 | 0.3 | 0.2 | 0.2 | 0.2 | 0.8 | 0.3 |
| Cynoscion nebulosus | Spotted seatrout | 0.5 | 0.5 | 0.0 | 0.0 | 0.1 | 0.0 | 1.4 | 0.6 |
| Mugil curema | White mullet | 0.2 | 0.2 | 0.5 | 0.5 | 0.0 | 0.0 | 0.6 | 0.5 |
| Centropomus undecimalis | Snook | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Megalops atlanticus | Tarpon | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lutjanus griseus | Gray snapper | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Pogonias cromis | Black drum | 0.0 | 0.0 | 0.5 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 |
| Archosargus probatocephalus | Sheepshead | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.1 | 0.1 |
| Mycteroperca microlepis | Gag | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.2 | 0.0 | 0.0 |
| Mugil gyrans | Fantail mullet | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Lutjanus synagris | Lane snapper | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Menticirrhus saxatilis | Northern kingfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Paralichthys albigutta | Gulf flounder | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Total Mean Density |  | 166.3 |  | 19.2 |  | 107.9 |  | 573.3 |  |
| Standard Error of Means |  | 6.9 |  | 0.6 |  | 2.9 |  | 8.9 |  |
| Total Species Richness |  | 7 |  | 9 |  | 10 |  | 12 |  |

STD Error: Standard Error of Means

When the microhabitat analysis is refined to just the selected taxa, a different set of results is found. The natural channel sites mean density $\left(648.5 / 100 \mathrm{~m}^{2}\right)$ is still the greatest for all sites and is more than two times that for the restored sites $\left(270.0 / 100 \mathrm{~m}^{2}\right)$. The natural subtidal pool sites $\left(166.3 / 100 \mathrm{~m}^{2}\right)$ also have substantially greater abundance than the restored sites (19.2/100 m${ }^{2}$ ). Conversely, the restored marsh edge (140.9/100 m${ }^{2}$ ) and seagrass (573.3/100 $\mathrm{m}^{2}$ ) sites had a higher total mean density than that of the natural sites $\left(89.7 / 100 \mathrm{~m}^{2}\right.$ and $107.9 / 100 \mathrm{~m}^{2}$ for marsh edge and seagrass sites, respectively). When evaluating total species richness, both the channel ( 8 species) and the marsh edge ( 10 species) microhabitat types have the same total species richness for both natural and restored sites. For the seagrass and subtidal pool microhabitat types, the restored sites actually recorded two more total species than their natural counterpart.

Evaluating the individual selected taxa, different results are found between natural and restored sites (Table 3.2-18). For Leiostomus xanthurus, the channel (546.7/100 $\mathrm{m}^{2}$ ), and subtidal pool ( $35.9 / 100 \mathrm{~m}^{2}$ ) sites had a greater density than their restored site counterparts (254.3/100 $\mathrm{m}^{2}$ for channel and $9.7 / 100 \mathrm{~m}^{2}$ for subtidal pools), whereas the opposite is true for both the marsh edge and the seagrass sites, where the restored sites $\left(114.8 / 100 \mathrm{~m}^{2}\right.$ for marsh edge and $357.9 / 100 \mathrm{~m}^{2}$ for seagrass) have a greater mean density than that of the natural sites ( $51.0 / 100 \mathrm{~m}^{2}$ for marsh edge and $100.2 / 100 \mathrm{~m}^{2}$ for seagrass). Leiostomus xanthurus reached its greatest mean density within the natural channel (546.7/100 $\mathrm{m}^{2}$ ) and was lowest within the restored subtidal pool sites $\left(9.7 / 100 \mathrm{~m}^{2}\right)$. The differences were even more dramatic for the Mugil spp. (combined total of all Mugil spp.) where higher numbers were found in natural channel ( $100.8 / 100 \mathrm{~m}^{2}$ ) and subtidal pool sites ( $129.6 / 100 \mathrm{~m}^{2}$ ) and observed densities for the marsh edge were similar between natural $\left(38.1 / 100 \mathrm{~m}^{2}\right)$ and restored sites $\left(24.3 / 100 \mathrm{~m}^{2}\right)$. Mugil spp. densities within the restored seagrass habitats (296.3/100 $\mathrm{m}^{2}$ ) were substantially higher than that of the natural seagrass $\left(4.1 / 100 \mathrm{~m}^{2}\right)$ microhabitats. Looking at the selected taxa that are game fish, spotted seatrout (Cynoscion nebulosus) was observed at every natural microhabitat type but was only found in one restored microhabitat type (seagrass). Conversely, the red drum (Sciaenops ocellatus) was
observed at every restored habitat type but was only observed within the natural seagrass microhabitat type.

Table 3.2-19 provides the mean density of taxa within three guilds for natural and restored sites within the four-microhabitat types. Within the channel microhabitat type, the natural nursery ( $727.5 / 100 \mathrm{~m}^{2}$ ) and transient (5,210.9/100 $\mathrm{m}^{2}$ ) guilds have higher mean density of that of their restored counterpart ( $563.8 / 100 \mathrm{~m}^{2}$ for nursery and $10.3 / 100 \mathrm{~m}^{2}$ for transient), whereas the restored resident species were greater in mean density $\left(754.4 / 100 \mathrm{~m}^{2}\right)$ than the natural channel $\left(424.5 / 100 \mathrm{~m}^{2}\right)$ sites. The opposite of this held true for the marsh edge sites where the observed densities for nursery $\left(174.2 / 100 \mathrm{~m}^{2}\right)$ and transient $\left(946.5 / 100 \mathrm{~m}^{2}\right)$ guilds for restored marsh edge were greater than the mean densities for natural sites $\left(115.0 / 100 \mathrm{~m}^{2}\right.$ for nursery and $114.3 / 100 \mathrm{~m}^{2}$ for transient guilds) and the resident natural marsh edge species ( $696.6 / 100 \mathrm{~m}^{2}$ ) had a greater mean density of that of the restored sites (306.8/100 $\mathrm{m}^{2}$ ). The seagrass microhabitat type also followed this pattern with a greater mean density for the nursery $\left(852.6 / 100 \mathrm{~m}^{2}\right)$ and transient $\left(1,183.3 / 100 \mathrm{~m}^{2}\right)$ guilds within the restored sites (compared to $245.2 / 100 \mathrm{~m}^{2}$ for nursery and $839.2 / 100 \mathrm{~m}^{2}$ for transients at natural sites) and a lower mean density for the resident ( $581.8 / 100 \mathrm{~m}^{2}$ compared to $1,912.4 / 100 \mathrm{~m}^{2}$ for natural) guild. For subtidal pool, the nursery guild mean density was greater for the natural sites $\left(170.5 / 100 \mathrm{~m}^{2}\right)$ than the restored sites $\left(24.5 / 100 \mathrm{~m}^{2}\right)$, whereas the resident $\left(865.0 / 100 \mathrm{~m}^{2}\right)$ and transient $\left(890.4 / 100 \mathrm{~m}^{2}\right)$ species were in greater mean density in the restored sites than at the natural sites $\left(227.3 / 100 \mathrm{~m}^{2}\right.$ for the resident species and $60.8 / 100 \mathrm{~m}^{2}$ for transient species) subtidal pools.

The total species richness observed within the various microhabitat types and guilds presents some contrast. For the nursery guild, the number of species (between natural and restored sites) is closely tied between all four microhabitat types, not varying by more than three species. The resident species guild is dramatically different between the natural and restored sites, especially for the channel sites where 19 species were observed for natural channel sites compared to 9 for restored channel sites. The total number of resident species observed

Table 3.2-19. Mean density ( $\# 100 \mathrm{~m}^{2}$ ) of taxa within three guilds within four microhabitat types at natural and restored sites.

|  |  | Microhabitat |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Guild |  | Channel | Marsh Edge | Subtidal Pool | Seagrass |
| Nursery | Total Mean Density | 727.5 | 115.0 | 178.5 | 245.2 |
|  | Standard Error of Means | 22.1 | 2.6 | 5.3 | 6.6 |
|  | Total Species Richness | 8 | 14 | 9 | 15 |
| Resident | Total Mean Density | 424.5 | 696.6 | 227.3 | 1912.4 |
|  | Standard Error of Means | 5.6 | 13.7 | 4.5 | 52.3 |
|  | Total Species Richness | 19 | 19 | 19 | 15 |
| Transient | Total Mean Density | 5210.9 | 114.3 | 60.8 | 839.2 |
|  | Standard Error of Means | 139.4 | 3.9 | 2.0 | 20.4 |
|  | Total Species Richness | 13 | 9 | 12 | 18 |

## Restored

|  | Microhabitat |  |  |  |  |
| :--- | :--- | ---: | ---: | ---: | ---: |
| Guild |  | Channel | Marsh Edge | Subtidal Pool | Seagrass |
| Nursery | Total Mean Density | 563.8 | 174.2 | 24.5 | 852.6 |
|  | Standard Error of Means | 15.3 | 4.8 | 0.5 | 18.1 |
|  | Total Species Richness | 8 | 12 | 10 | 17 |
|  |  |  |  |  |  |
| Resident | Total Mean Density | 754.4 | 306.8 | 865.0 | 581.8 |
|  | Standard Error of Means | 13.4 | 5.0 | 26.5 | 10.3 |
|  | Total Species Richness | 9 | 14 | 12 | 17 |
|  |  |  |  | 890.4 | 1183.3 |
|  | Transient | Total Mean Density | 10.3 | 946.5 | 31.2 |
|  | Standard Error of Means | 0.2 | 28.8 | 10 | 40.8 |
|  | Total Species Richness | 6 | 12 | 12 |  |

for the natural marsh edge ( 19 species) and subtidal pool ( 19 species) sites were also above that of the resident species within the restored site counterpart ( 14 species for marsh edge and 12 species for subtidal pool), whereas the restored seagrass sites actually had two more species ( 17 species) observed than the natural seagrass sites ( 15 species). The total species richness for transient species was also higher at three of the four natural microhabitats, the exception being the marsh edge where three more species were observed within restored habitat types ( 12 species compared to 9 species).

### 3.3 Artificial Reef Assessment Results

The artificial reef assessment surveys occurred during late winter (February 21, 22 and 26, 2002) and early summer (June 17, 20, 27 and 28, 2002) at the 14 artificial reef stations and 3 reef sites. Table 3.3-1 presents a master species list with scientific name, common name, family name and selected taxa status for all observed species. The number of each species, by size class, is also provided for both sampling seasons.

During the February surveys, there was an extensive amount of drift algae present at all sites. The drift algae were 0.3 to 0.5 m in depth around the reef material. Since all individuals within one-meter of the reef structure were counted, some individuals counted may have been related to the drift algae more so than the reef material. However, the species that were found over the drift algae were also found over the reef material. However, the ubiquitous Lagodon rhomboides, primarily the early juvenile stage, was most prevalent over the drift algae. The drift algae disappeared before the June surveys. Reduced visibility during the June surveys may have reduced the number of observations, especially at the Hart and Gerkin sites.

A total of approximately 59,027 individuals within 27 species or species groups were observed. Eighteen species were observed during the February surveys and 23 species were observed during the June surveys. Clupeidae spp., Haemulon aurolineatum, larval fish and Lagodon rhomboides were the dominant species observed (see Table 3.3-1). These four

Table 3.3-1. Master species list for the artificial reef assessment surveys and total number of individuals observed at each station by sampling season.

| Scientific name | Common Name | Family Name | $\begin{aligned} & \text { Selected } \\ & \text { Taxa } \end{aligned}$ | February |  |  |  | June |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | EJ | JUV | SA | A | EJ | JUV | SA | A |
| Acanthostracion quadricornis | Cowfish | Ostraciidae |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Archosargus probatocephalus | Sheepshead | Sparidae | X | 1 | 8 | 11 | 101 | 0 | 34 | 48 | 51 |
| Balistidae spp. | Filefish | Balistidae |  | 0 | 0 | 0 | 0 | 0 | 15 | 2 | 2 |
| Blenniidae spp. | Blenny | Blenniidae |  | 0 | 0 | 0 | 6 | 0 | 0 | 0 | 1 |
| Calamus sp. | Porgy | Sparidae |  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Callinectes sapidus | Blue Crab | Portunidae | X | 0 | 0 | 0 | 7 | 0 | 2 | 0 | 5 |
| Chaetodipterus faber | Atlantic Spadefish | Ephippidae |  | 0 | 0 | 155 | 27 | 0 | 25 | 17 | 19 |
| Chilomycterus schoepfi | Striped Burrfish | Diodontidae |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Clupeidae spp. | Sardine | Clupeidae spp. |  | 0 | 0 | 0 | 0 | 15,000 | 25,100 | 0 | 0 |
| Diplectrum formosum | Sand Perch | Serranidae |  | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| Diplodus holbrooki | Spottail Pinfish | Sparidae |  | 118 | 151 | 0 | 11 | 0 | 11 | 16 | 6 |
| Eucinostomus spp. | Mojarra | Gerridae |  | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 0 |
| Gobiidae sp. | Goby | Gobiidae sp. |  | 0 | 0 | 0 | 6 | 0 | 1 | 0 | 0 |
| Haemulon aurolineatum | Tomtate | Pomadasyidae |  | 6,318 | 5 | 1 | 0 | 0 | 4 | 5 | 0 |
| Haemulon plumieri | White Grunt | Pomadasyidae |  | 0 | 6 | 4 | 0 | 0 | 1 | 7 | 5 |
| Lagodon rhomboides | Pinfish | Sparidae |  | 3,054 | 331 | 16 | 1 | 1 | 85 | 524 | 15 |
| Larval fish | Larval Fish | --- |  | 5,641 | 0 | 0 | 0 | 600 | 0 | 0 | 0 |
| Leiostomus xanthurus | Spot | Sciaenidae | X | 0 | 1 | 0 | 0 | 0 | 143 | 3 | 0 |
| Lutjanus griseus | Gray Snapper | Lutjanidae | X | 11 | 60 | 57 | 49 | 0 | 257 | 378 | 77 |
| Lutjanus synagris | Lane Snapper | Lutjanidae | X | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Menippe mercenaria | Stone Crab | Xanthidae | X | 1 | 21 | 23 | 66 | 0 | 15 | 47 | 62 |
| Mycteroperca microlepsis | Gag | Serranidae | X | 0 | 1 | 9 | 7 | 0 | 39 | 15 | 4 |
| Mycteroperca phenax | Scamp | Serranidae | X | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 |
| Opsanus beta | Oyster toadfish | Batrachoididae |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| Orthopristis chrysoptera | Pigfish | Pomadasyidae |  | 0 | 0 | 0 | 0 | 0 | 46 | 27 | 2 |
| Prionotus sp. | Searobin | Trigilidae |  | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| Scorpaenidae sp. | Scorpion fish | Scorpaenidae sp. |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  |  |  |  | 15,144 | 584 | 280 | 286 | 15,609 | 25,778 | 1,091 | 255 |

${ }^{1}$ Selected taxa- Commercially and/or Recreationally important species.
EJ - Early-Juvenile, JUV - Juvenile, SA - Subadult, A - Adult
species groups comprised $96 \%$ of all individuals. The early-juvenile $(30,753)$ and juvenile $(26,362)$ size classes were most often observed. These two size classes combined for roughly $97 \%$ of the total fish count, with a little over $2 \%$ for subadults and less than $1 \%$ for adults. Over $72 \%$ of the individuals observed were recorded for the June season, however, the percentage drops to $14 \%$ when Clupeidae spp. are excluded.

Table 3.3-2 provides summary statistic for the non-selected taxa pool for both sampling season and all reef sites. A total of 57,408 individuals were estimated comprising over 97\% of the total number of individuals counted.

As seen in Table 3.3-2, Clupeidae spp was the dominant non-selected taxa for both the early juvenile ( 15,000 individuals) and juvenile ( 25,100 individuals) life classes. Although comprising a large percentage ( $69.9 \%$ ) of the number of individuals, this species occurred with relative infrequency at the stations. Lagodon rhomboides was the most frequently encountered species within all four size classes (16 to 64\%). Haemulon aurolineatum was the second most abundant ( 6,333 individuals) and the second frequently encountered species for both the early juvenile (23\%) and juvenile classes (11\%), while the Chaetodipterus faber was the second most frequently encountered species within the juvenile (11\%), subadult ( $14 \%$ ) and adult ( $14 \%$ ) categories. The number of non-selected taxa individuals decreased as the size class increased ( 30,740 early-juvenile to 112 adults). Conversely, the number of species encountered actually increased as the size class increased (from six at the early juvenile stage to 14 at the adult stage).

The summary statistics for the selected taxa observed in the artificial reef assessment are presented in Table 3.3-3. Select taxa accounted for less than 3\% of the total number of individuals observed during the survey. Lutjanus griseus was the dominant selected taxa observed (889 individuals), followed by Archosargus probatocephalus at 254 individuals and Menippe mercenaria at 235 individuals. Contrary to that which was observed for the non-selected taxa, the early juvenile category had the lowest number of observations with

Table 3.3-2 Summary statistics for non-selected ${ }^{1}$ taxa observed during the artificial reef assessment.

| Species name | Early Juvenile |  |  | Juvenile |  |  | Subadult |  |  | Adult |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number | Percent | Frequency | Number | Percent | Frequency | Number | Percent | Frequency | Number | Percent | Frequency |
| Acanthostracion quadricornis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 100 | 2 |
| Balistidae spp. | 0 | 0 | 0 | 15 | 79 | 7 | 2 | 11 | 4 | 2 | 11 | 4 |
| Blenniidae spp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 100 | 5 |
| Calamus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 100 | 2 | 0 | 0 | 0 |
| Chaetodipterus faber | 0 | 0 | 0 | 25 | 10 | 11 | 172 | 71 | 14 | 46 | 19 | 14 |
| Chilomycterus schoepfi | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 100 | 2 |
| Clupeidae spp. | 15,000 | 37 | 3 | 25,100 | 63 | 6 | 0 | 0 | 0 | 0 | 0 | 0 |
| Diplectrum formosum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 100 | 4 |
| Diplodus holbrooki | 118 | 38 | 13 | 162 | 52 | 7 | 16 | 5 | 5 | 17 | 5 | 13 |
| Eucinostomus spp. | 8 | 100 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gobiidae sp. | 0 | 0 | 0 | 1 | 14 | 2 | 0 | 0 | 0 | 6 | 86 | 5 |
| Haemulon aurolineatum | 6,318 | 100 | 23 | 9 | 0 | 11 | 6 | 0 | 7 | 0 | 0 | 0 |
| Haemulon plumieri | 0 | 0 | 0 | 7 | 30 | 5 | 11 | 48 | 14 | 5 | 22 | 9 |
| Lagodon rhomboides | 3,055 | 76 | 45 | 416 | 10 | 64 | 540 | 13 | 29 | 16 | 0 | 16 |
| Larval fish | 6,241 | 100 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Opsanus beta | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 100 | 5 |
| Orthopristis chrysoptera | 0 | 0 | 0 | 46 | 61 | 5 | 27 | 36 | 7 | 2 | 3 | 4 |
| Prionotus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 100 | 4 |
| Scorpaenidae sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 100 | 2 |
| Total | 30,740 |  |  | 25,781 |  |  | 775 |  |  | 112 |  |  |

${ }^{1}$ Species not listed as Commercially or Recreationally important.

Table 3.3-3 Summary statistics for selected taxa ${ }^{1}$ observed during the artificial reef assessment.

|  | Early Juvenile |  |  | Juvenile |  |  | Subadult |  |  | Adult |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species name | Number | Percent | Frequency | Number | Percent | Frequency | Number | Percent | Frequency | Number | Percent | Frequency |
| Archosargus probatocephalus | 1 | 0 | 2 | 42 | 17 | 29 | 59 | 23 | 38 | 152 | 60 | 54 |
| Callinectes sapidus | 0 | 0 | 0 | 2 | 14 | 0 | 0 | 0 | 0 | 12 | 86 | 4 |
| Leiostomus xanthurus | 0 | 0 | 0 | 144 | 98 | 7 | 3 | 2 | 2 | 0 | 0 | 0 |
| Lutjanus griseus | 11 | 1 | 9 | 317 | 36 | 73 | 435 | 49 | 64 | 126 | 14 | 38 |
| Lutjanus synagris | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 100 | 2 | 0 | 0 | 0 |
| Menippe mercenaria | 1 | 0 | 4 | 36 | 15 | 30 | 70 | 30 | 38 | 128 | 54 | 59 |
| Mycteroperca microlepsis | 0 | 0 | 0 | 40 | 53 | 25 | 24 | 32 | 16 | 11 | 15 | 14 |
| Mycteroperca phenax | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 100 | 4 | 0 | 0 | 0 |
| Total | 13 |  |  | 581 |  |  | 596 |  |  | 429 |  |  |

${ }^{1} \mathrm{Sp}$ ecies listed as Commercially or Recreationally important.
only 13 individuals being noted within three species. The number of selected taxa observed was actually higher in the subadult category ( 596 individuals), followed closely by the juvenile category (581 individuals) and then the adult category (429 individuals). Of the eight selected taxa observed, three species were observed at the early juvenile stage, six at the juvenile stage, seven at the subadult stage and five at the adult stage.

In addition to being the most abundant species of selected taxa observed, Lutjanus griseus was the most frequently encountered species during the sampling for both the early juvenile ( $9 \%$ ), juvenile ( $73 \%$ ) and sub adult ( $64 \%$ ) categories. Menippe mercenaria was the most frequently observed species in the adult category (59\%) and was the second most commonly frequency encountered species for the early-juvenile (4\%), juvenile (30\%) and subadult (38\%) categories. Archosargus probatocephalus was also the second most frequently observed subadult ( $38 \%$ ) and adult (54\%). Most selected taxa were observed at multiple life stages, with the only exceptions being Lutjanus synagris and Mycteroperca phenax, which were only observed as subadults. Of the eight selected taxa, six were finfish and two were shellfish (Callinectes sapidus and Menippe mercenaria).

Table 3.3-4 provides a comparison of the mean density of selected and non-selected taxa between the three different reef sites. For the early-juvenile size class, the highest overall density $\left(6,343.9 / 100 \mathrm{~m}^{2}\right)$ and the greatest occurrence $(6.3 \%)$ were at the Hart site, while species richness ( 7 species) was highest at the Walker site. The selected taxa density (21.4/100 $\mathrm{m}^{2}$ ) and percent occurrence ( $2.5 \%$ ) were highest at the Gerkin site, while the nonselected species were at their highest density $\left(6,342.3 / 100 \mathrm{~m}^{2}\right)$ and percent occurrence (8.7\%) at the Hart site.

The density of early-juvenile Lutjanus griseus was substantially higher at the Gerkin site $\left(20.5 / 100 \mathrm{~m}^{2}\right)$. The only other early-juvenile observed at this site was Archosargus probatocephalus. One early-juvenile Menippe mercenaria was observed at the Hart site. Haemulon aurolineatum density was highest at the Gerkin site $\left(3,758.5 / 100 \mathrm{~m}^{2}\right)$, but this

Table 3.3-4. Comparison of mean density ( $\# / 100 \mathrm{~m}^{2}$ ) of selected and non-selected taxa between the Gerkin, Hart and Walker reef sites.

| Species Name | Gerkin |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Early-Juvenile |  |  | Juvenile |  |  | Subadult |  |  | Adult |  |  |
|  | Mean <br> Density | $\begin{aligned} & \hline \text { STD } \\ & \text { Error } \\ & \hline \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{aligned} & \hline \text { STD } \\ & \text { Error } \\ & \hline \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \hline \text { STD } \\ & \text { Error } \\ & \hline \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \mathrm{OCC} \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 20.5 | 14.5 | 15 | 379.4 | 108.1 | 85 | 296.0 | 82.2 | 85 | 58.1 | 18.2 | 55 |
| Archosargus probatocephalus | 0.9 | 0.9 | 5 | 116.9 | 41.4 | 65 | 83.1 | 35.0 | 60 | 46.3 | 16.7 | 60 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 4.7 | 2.6 | 30 | 15.2 | 10.4 | 10 | 48.8 | 23.2 | 50 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 24.5 | 13.6 | 20 | 0.2 | 0.2 | 5 | 0.9 | 0.9 | 5 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 21.4 | 2.6 | 2.5 | 525.5 | 47.0 | 25.0 | 394.4 | 36.7 | 20.0 | 154.1 | 9.4 | 21.3 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 3758.5 | 2366.0 | 15 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 106.7 | 106.7 | 5 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 1847.8 | 998.3 | 45 | 138.1 | 58.0 | 50 | 28.4 | 16.1 | 30 | 9.5 | 7.6 | 20 |
| Diplodus holbrooki | 0.0 | 0.0 | 0 | 0.2 | 0.2 | 5 | 0.4 | 0.4 | 5 | 22.4 | 10.4 | 35 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 18.9 | 10.1 | 25 | 122.8 | 61.6 | 25 | 2.6 | 2.3 | 10 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 3.8 | 3.8 | 5 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 7.6 | 7.6 | 5 | 3.8 | 3.8 | 5 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.2 | 0.2 | 5 | 1.8 | 1.8 | 5 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 1.5 | 0.9 | 15 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 3.8 | 3.8 | 5 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 11.4 | 8.3 | 10 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 15.2 | 10.4 | 10 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 5713.0 | 226.8 | 3.6 | 157.4 | 7.7 | 4.7 | 161.0 | 6.9 | 3.9 | 73.9 | 1.5 | 6.4 |
| Total All Taxa | 5734.4 | 158.3 | 3.3 | 683.0 | 15.7 | 11.0 | 555.5 | 12.3 | 8.8 | 228.0 | 3.3 | 11.0 |
| Percent Dominant's Density | 97.8 |  |  | 92.9 |  |  | 95.5 |  |  | 83.7 |  |  |
| Percent Non-dominant's Density | 2.2 |  |  | 7.1 |  |  | 4.5 |  |  | 16.3 |  |  |
| Species Richness | 5 |  |  | 8 |  |  | 9 |  |  | 13 |  |  |

Table 3.3-4. (Continued)

| Species Name | Hart |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Early-Juvenile |  |  | Juvenile |  |  | Subadult |  |  | Adult |  |  |
|  | Mean <br> Density | $\begin{gathered} \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \mathrm{OCC} \\ \hline \end{gathered}$ | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \mathrm{OCC} \\ \hline \end{gathered}$ | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 1.6 | 1.6 | 6 | 42.2 | 20.5 | 50 | 91.4 | 36.2 | 56 | 46.9 | 21.4 | 44 |
| Archosargus probatocephalus | 0.0 | 0.0 | 0 | 3.2 | 3.2 | 6 | 39.0 | 20.3 | 44 | 273.2 | 81.4 | 88 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 28.4 | 20.6 | 13 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 10.2 | 9.2 | 25 | 9.8 | 9.4 | 19 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 1.6 | 0.2 | 0.8 | 45.4 | 5.2 | 7.0 | 140.6 | 11.6 | 15.6 | 358.3 | 33.2 | 20.3 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 1993.1 | 1132.9 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 2385.8 | 1696.1 | 19 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 1881.4 | 706.8 | 50 | 235.2 | 104.6 | 50 | 9.5 | 42.9 | 6 | 0.0 | 0.0 | 0 |
| Diplodus holbrooki | 53.5 | 32.8 | 25 | 0.0 | 0.0 | 0 | 0.3 | 0.3 | 6 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 6.9 | 5.0 | 13 | 52.4 | 26.3 | 38 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 1.0 | 0.9 | 13 | 1.0 | 0.6 | 25 | 28.6 | 15.2 | 25 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.3 | 0.2 | 13 |
| Eucinostomus spp. | 28.4 | 20.6 | 13 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.2 | 0.2 | 6 | 0.0 | 0.0 | 0 | 1.6 | 1.6 | 6 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 9.5 | 9.5 | 6 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.2 | 0.2 | 6 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 6342.3 | 189.4 | 8.7 | 236.4 | 13.1 | 3.8 | 17.8 | 0.6 | 2.8 | 92.6 | 3.2 | 5.2 |
| Total All Taxa | 6343.9 | 133.9 | 6.3 | 281.8 | 9.1 | 4.8 | 158.4 | 3.8 | 6.7 | 450.9 | 10.6 | 9.9 |
| Percent Dominant's Density | 0.0 |  |  | 98.4 |  |  | 94.8 |  |  | 95.3 |  |  |
| Percent Non-dominant's Density | 100.0 |  |  | 1.6 |  |  | 5.2 |  |  | 4.7 |  |  |
| Species Richness | 0 |  |  | 5 |  |  | 7 |  |  | 10 |  |  |

Table 3.3-4. (Continued)

| Species Name | Walker |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Early-Juvenile |  |  | Juvenile |  |  | Subadult |  |  | Adult |  |  |
|  | Mean <br> Density | STD <br> Error | $\begin{gathered} \text { \% } \\ \text { OCC } \end{gathered}$ | Mean <br> Density | STD <br> Error | $\begin{gathered} \text { \% } \\ \text { OCC } \end{gathered}$ | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 0.3 | 0.3 | 5 | 325.6 | 111.1 | 80 | 573.6 | 256.7 | 50 | 11.4 | 6.9 | 15 |
| Archosargus probatocephalus | 0.0 | 0.0 | 0 | 2.9 | 2.8 | 10 | 1.8 | 1.5 | 10 | 7.5 | 3.6 | 20 |
| Menippe mercenaria | 0.5 | 0.5 | 5 | 53.4 | 14.2 | 50 | 113.6 | 21.2 | 90 | 167.6 | 23.9 | 100 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 46.9 | 24.8 | 20 | 1.4 | 1.4 | 5 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 65.4 | 23.4 | 50 | 4.5 | 2.4 | 20 | 1.8 | 0.9 | 20 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 1.0 | 0.8 | 5 | 0.0 | 0.0 | 0 | 26.6 | 15.5 | 30 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 1.0 | 0.8 | 10 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 3.8 | 3.8 | 5 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 0.7 | 0.1 | 1.3 | 495.2 | 38.9 | 26.9 | 699.7 | 70.8 | 23.8 | 214.9 | 20.4 | 23.1 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 38.4 | 37.9 | 10 | 17.6 | 9.2 | 30 | 2.4 | 1.5 | 20 | 0.0 | 0.0 | 0 |
| Larval fish | 801.9 | 755.7 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 981.7 | 582.8 | 40 | 746.2 | 161.7 | 90 | 598.9 | 216.4 | 45 | 14.1 | 8.3 | 25 |
| Diplodus holbrooki | 26.0 | 25.0 | 15 | 40.3 | 32.9 | 15 | 3.0 | 3.0 | 5 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.5 | 0.4 | 5 | 0.5 | 0.5 | 5 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 12.0 | 8.8 | 15 | 25.7 | 19.2 | 20 | 0.5 | 0.5 | 5 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 3.8 | 3.3 | 5 | 1.2 | 0.7 | 15 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 10.0 | 5.6 | 15 | 0.5 | 0.5 | 5 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 1.3 | 1.3 | 5 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 1.3 | 0.9 | 10 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.5 | 0.5 | 5 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.5 | 0.5 | 5 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.5 | 0.5 | 5 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.3 | 0.3 | 5 |
| Subtotal Non-selected Taxa | 1849.2 | 68.1 | 5.3 | 830.3 | 41.3 | 9.7 | 632.7 | 33.2 | 6.7 | 17.1 | 0.8 | 3.1 |
| Total All Taxa | 1850.0 | 47.7 | 4.0 | 1325.5 | 30.6 | 15.0 | 1332.4 | 31.2 | 11.9 | 231.9 | 6.5 | 9.2 |
| Percent Dominant's Density | 96.4 |  |  | 80.9 |  |  | 96.5 |  |  | 89.8 |  |  |
| Percent Non-dominant's Density | 3.6 |  |  | 19.1 |  |  | 3.5 |  |  | 10.2 |  |  |
| Species Richness | 7 |  |  | 13 |  |  | 15 |  |  | 11 |  |  |

\% OCC - Percent Occurrence (percent of samples in which the species was observed)
\% OCC - Percent Occurrence (percent of
STD Error - Standard Error of the Means
species was more frequently observed at the Hart site ( $50 \%$ occurrence). All other early-juvenile non-selected taxa reached their highest density at the Hart site and the species were generally observed here more frequently.

The Walker site displayed the highest overall density $\left(1,325.5 / 100 \mathrm{~m}^{2}\right)$, species richness (13) and percent occurrence ( $15.0 \%$ ) for juveniles. The selected taxa density of juveniles was highest at the Gerkin site $\left(525.5 / 100 \mathrm{~m}^{2}\right)$, while selected taxa juveniles were more frequently observed at the Walker site ( $26.9 \%$ occurrence). For non-selected taxa, mean density ( $830.3 / 100 \mathrm{~m}^{2}$ ) and percent occurrence (9.7\%) were highest at the Walker reef site. Lutjanus griseus mean density (379.4/100 $\mathrm{m}^{2}$ ) and percent occurrence ( $85 \%$ ) were highest at the Gerkin site followed closely by the Walker site ( $325.6 / 100 \mathrm{~m}^{2}$ and $80 \%$ ). Archosargus probatocephalus mean density (116.9/100 $\mathrm{m}^{2}$ ) and percent occurrence ( $65 \%$ ) were substantially higher at the Gerkin site than on the other reef sites. The mean density and percent occurrence of the other four select taxa observed had greater mean density and percent occurrence at the Walker reef site. With the exception of the Chaetodipterus faber, the majority of the non-selected taxa reached their highest densities and highest percent occurrence in the juvenile size class at the Walker reef site.

As observed with the juvenile size class, the Walker reef site also had the highest overall mean density ( $1,332.4 / 100 \mathrm{~m}^{2}$ ), percent occurrence (11.9\%) and species richness (15), for the subadult size class. The Walker reef site also had the highest mean density and percentage occurrence for both the selected ( $699.7 / 100 \mathrm{~m}^{2}$ and $23.8 \%$, respectively) and non-selected taxa ( $632.7 / 100 \mathrm{~m}^{2}$ and $6.7 \%$, respectively). Lutjanus griseus was the most frequently encountered (50\%) and dominant selected ( $573.6 / 100 \mathrm{~m}^{2}$ ) taxa at the Walker site while its percent occurrence was actually higher for the Gerkin site (85\%). Archosargus probatocephalus mean density (83.1/100 $\mathrm{m}^{2}$ ) and percent occurrence ( $60 \%$ ) were both highest at the Gerkin site, while Mycteroperca microlepsis mean density ( $10.2 / 100 \mathrm{~m}^{2}$ ) and percent occurrence ( $25 \%$ ) were highest at the Hart site. All other subadult's mean density and percent occurrence were highest at the Walker site.

For the adult size class, the mean density $\left(450.9 / 100 \mathrm{~m}^{2}\right)$ was highest at the Hart reef site, while the percent occurrence ( $11.0 \%$ ) and the species richness (13) were highest at the Gerkin site.

Mean density of the adult size class was highest for the selected taxa at the Hart reef (358.3/100 $\mathrm{m}^{2}$ ), while the percent occurrence ( $23.1 \%$ ) was highest at the Walker site. Adult non-selected taxa mean density was highest at the Hart site ( $92.6 / 100 \mathrm{~m}^{2}$ ), while percent occurrence (11.8\%) was highest at the Gerkin site. Adult Lutjanus griseus mean density $\left(58.1 / 100 \mathrm{~m}^{2}\right)$ and percent occurrence ( $55 \%$ ) were highest at the Gerkin site, Archosargus probatocephalus mean density (273.2/100 $\mathrm{m}^{2}$ ) and percent occurrence ( $88 \%$ ) were highest at the Hart site, while Menippe mercenaria ( $167.6 / 100 \mathrm{~m}^{2}$ ) and Callinectes sapidus $\left(26.6 / 100 \mathrm{~m}^{2}\right)$ density and percent occurrence were highest at the Walker site ( $100 \%$ and $30 \%$, respectively). Adult Mycteroperca microlepsis had their highest mean density at the Hart site $\left(9.8 / 100 \mathrm{~m}^{2}\right)$, while the highest percent occurrence ( $20 \%$ ) was at the Walker site. For the non-select taxa, six adults species had their highest mean density and percent occurrence at the Hart site, while five species had the highest mean density and percent occurrence at the Gerkin site, followed by three species at the Walker site.

Tables 3.3-5 through 3.3-7 provide a comparison of the mean density of selected and nonselected taxa between stations at the Gerkin, Hart and Walker reef sites, respectively.

## GERKIN

For the Gerkin reef site (Table 3.3-5), early-juvenile species were generally at their greatest density $\left(20,503.4 / 100 \mathrm{~m}^{2}\right)$ and percent occurrence $(5.6 \%)$ at station GP3, yet the greatest richness (4 species) was at station GL1. Only two selected species of early-juveniles, Lutjanus griseus and Archosargus probatocephalus, were observed at any of the five stations. An additional three species of early-juveniles from non-selected taxa were observed among these same five sites with Lagodon rhomboides being present in fair numbers at all sites. Haemulon aurolineatum reached very high density $\left(16,659.6 / 100 \mathrm{~m}^{2}\right)$ at station GP3, but otherwise was only present at station GL1 (2,133.1/100 m ${ }^{2}$ ).

Table 3.3-5. Comparison of mean density $\left(\# / 100 \mathrm{~m}^{2}\right)$ of selected and non-selected taxa between stations at the Gerkin reef site.

| Species Name | Early-Juvenile |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean <br> Density | $\begin{aligned} & \frac{\text { GL1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | $\begin{aligned} & \frac{\text { GP1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\frac{\mathrm{GP2}}{\mathrm{STD}}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | $\begin{gathered} \underset{\text { GP3 }}{\text { STD }} \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | $\begin{gathered} \underset{\text { GP4 }}{\text { STD }} \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 1.1 | 1.1 | 25 | 0.0 | 0.0 | 0 | 37.9 | 37.9 | 25 | 0.0 | 0.0 | 0 | 63.6 | 63.6 | 25 |
| Archosargus probatocephalus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 4.3 | 4.3 | 25 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 1.1 | 0.1 | 3.1 | 0.0 | 0.0 | 0.0 | 37.9 | 4.7 | 3.1 | 4.3 | 0.5 | 3.1 | 63.6 | 8.0 | 3.1 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 2133.1 | 2133.1 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 16659.6 | 10111.0 | 50 | 0.0 | 0.0 | 0 |
| Larval fish | 533.3 | 533.3 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 1205.2 | 1162.9 | 50 | 38.2 | 38.2 | 25 | 4128.8 | 3735.7 | 50 | 3839.6 | 3295.4 | 50 | 27.3 | 17.4 | 50 |
| Diplodus holbrooki | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 3871.6 | 133.5 | 5.6 | 38.2 | 2.1 | 1.4 | 4128.8 | 229.4 | 2.8 | 20499.1 | 937.5 | 5.6 | 27.3 | 1.5 | 2.8 |
| Total All Taxa | 3872.7 | 93.7 | 4.8 | 38.2 | 1.5 | 1.0 | 4166.7 | 158.7 | 2.9 | 20503.4 | 651.8 | 4.8 | 90.9 | 2.6 | 2.9 |
| Percent Dominant's Density | 100.0 |  |  | 100.0 |  |  | 99.1 |  |  | 100.0 |  |  | 100.0 |  |  |
| Percent Non-dominant's Density | 0.0 |  |  | 0.0 |  |  | 0.9 |  |  | 0.0 |  |  | 0.0 |  |  |
| Species Richness | 4 |  |  | 1 |  |  | 2 |  |  | 3 |  |  | 2 |  |  |

Table 3.3-5. (Continued)

| Species Name | Juvenile |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean <br> Density | $\begin{gathered} \frac{\text { GL1 }}{\text { STD }} \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \frac{\text { GP1 }}{\text { STD }} \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \frac{\text { GP2 }}{\text { STD }} \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \hline \text { GP3 } \\ \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \hline \underline{\text { GP4 }} \\ \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 65.1 | 21.1 | 100 | 706.1 | 252.0 | 100 | 719.7 | 393.0 | 75 | 251.7 | 140.6 | 75 | 154.5 | 52.2 | 75 |
| Archosargus probatocephalus | 4.3 | 3.0 | 50 | 248.1 | 122.2 | 75 | 265.2 | 129.4 | 75 | 21.3 | 16.2 | 50 | 45.5 | 17.4 | 75 |
| Menippe mercenaria | 6.4 | 1.2 | 100 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 17.1 | 12.1 | 50 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 95.4 | 57.3 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 27.3 | 17.4 | 50 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 75.7 | 8.0 | 31.3 | 1049.6 | 87.8 | 28.1 | 984.8 | 91.3 | 18.8 | 290.1 | 30.9 | 21.9 | 227.3 | 19.0 | 25.0 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 56.5 | 41.3 | 50 | 114.5 | 73.1 | 50 | 189.4 | 143.4 | 50 | 302.9 | 249.3 | 50 | 27.3 | 17.4 | 50 |
| Diplodus holbrooki | 1.1 | 1.1 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 7.5 | 4.4 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 59.7 | 40.3 | 50 | 27.3 | 27.3 | 25 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 1.1 | 1.1 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 66.1 | 3.1 | 8.3 | 114.5 | 6.4 | 2.8 | 189.4 | 10.5 | 2.8 | 362.6 | 17.0 | 5.6 | 54.5 | 2.1 | 4.2 |
| Total All Taxa | 141.9 | 3.2 | 15.4 | 1164.1 | 28.6 | 10.6 | 1174.2 | 29.6 | 7.7 | 652.7 | 14.8 | 10.6 | 281.8 | 6.2 | 10.6 |
| Percent Dominant's Density | 91.0 |  |  | 100.0 |  |  | 100.0 |  |  | 94.1 |  |  | 100.0 |  |  |
| Percent Non-dominant's Density | 9.0 |  |  | 0.0 |  |  | 0.0 |  |  | 5.9 |  |  | 0.0 |  |  |
| Species Richness | 7 |  |  | 4 |  |  | 3 |  |  | 5 |  |  | 5 |  |  |

Table 3.3-5. (Continued)

| Species Name | Subadult |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean Density | $\begin{aligned} & \frac{\text { GL1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \underline{\text { GP1 }} \\ \text { STror } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \mathrm{GP2} \\ & \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean <br> Density | $\begin{gathered} \hline \text { GP3 } \\ \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \underset{\text { Error }}{\text { GP4 }} \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 65.1 | 32.7 | 100 | 572.5 | 257.9 | 75 | 454.5 | 269.6 | 50 | 115.2 | 51.4 | 100 | 272.7 | 90.9 | 100 |
| Archosargus probatocephalus | 8.5 | 5.2 | 50 | 209.9 | 129.9 | 50 | 113.6 | 113.6 | 25 | 46.9 | 12.8 | 100 | 36.4 | 14.8 | 75 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 75.8 | 43.7 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 1.1 | 1.1 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 74.7 | 8.0 | 21.9 | 782.4 | 72.6 | 15.6 | 643.9 | 55.7 | 15.6 | 162.1 | 14.8 | 25.0 | 309.1 | 33.7 | 21.9 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 3.2 | 3.2 | 25 | 95.4 | 72.3 | 50 | 0.0 | 0.0 | 0 | 34.1 | 28.7 | 50 | 9.1 | 9.1 | 25 |
| Diplodus holbrooki | 2.1 | 2.1 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 48.0 | 28.0 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 384.0 | 224.4 | 50 | 181.8 | 181.8 | 25 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 38.2 | 38.2 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 9.1 | 9.1 | 25 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 53.3 | 2.7 | 5.6 | 133.6 | 5.6 | 4.2 | 0.0 | 0.0 | 0.0 | 418.1 | 21.3 | 5.6 | 200.0 | 10.1 | 4.2 |
| Total All Taxa | 128.0 | 3.0 | 10.6 | 916.0 | 23.2 | 7.7 | 643.9 | 17.9 | 4.8 | 580.2 | 15.2 | 11.5 | 509.1 | 12.3 | 9.6 |
| Percent Dominant's Density | 95.0 |  |  | 95.8 |  |  | 100.0 |  |  | 100.0 |  |  | 96.4 |  |  |
| Percent Non-dominant's Density | 5.0 |  |  | 4.2 |  |  | 0.0 |  |  | 0.0 |  |  | 3.6 |  |  |
| Species Richness | 6 |  |  | 4 |  |  | 3 |  |  | 4 |  |  | 5 |  |  |

Table 3.3-5. (Continued)

| Species Name | Adult |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean Density | $\begin{aligned} & \frac{\text { GL1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{aligned} & \frac{\text { GP1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{gathered} \mathrm{GP2} \\ \text { STD } \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \frac{\text { GP3 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{gathered} \text { GP4 } \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 19.2 | 2.8 | 100 | 57.3 | 57.3 | 25 | 37.9 | 37.9 | 25 | 85.3 | 42.4 | 75 | 90.9 | 52.5 | 50 |
| Archosargus probatocephalus | 13.9 | 4.0 | 100 | 38.2 | 22.0 | 50 | 75.8 | 75.8 | 25 | 85.3 | 28.7 | 100 | 18.2 | 18.2 | 25 |
| Menippe mercenaria | 1.1 | 1.1 | 25 | 38.2 | 22.0 | 50 | 151.5 | 107.1 | 50 | 17.1 | 12.1 | 50 | 36.4 | 14.8 | 75 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 4.3 | 4.3 | 25 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 34.1 | 2.7 | 28.1 | 133.6 | 8.4 | 15.6 | 265.2 | 19.5 | 12.5 | 192.0 | 13.5 | 31.3 | 145.5 | 11.4 | 18.8 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 1.1 | 1.1 | 25 | 0.0 | 0.0 | 0 | 37.9 | 37.9 | 25 | 8.5 | 4.9 | 50 | 0.0 | 0.0 | 0 |
| Diplodus holbrooki | 9.6 | 6.1 | 50 | 38.2 | 38.2 | 25 | 37.9 | 37.9 | 25 | 17.1 | 12.1 | 50 | 9.1 | 9.1 | 25 |
| Chaetodipterus faber | 12.8 | 11.4 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 19.1 | 19.1 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 19.1 | 19.1 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 7.5 | 3.2 | 75 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 19.1 | 19.1 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 19.1 | 19.1 | 25 | 37.9 | 37.9 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 75.8 | 43.7 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 30.9 | 0.9 | 11.1 | 114.5 | 2.7 | 6.9 | 189.4 | 5.1 | 6.9 | 25.6 | 1.0 | 5.6 | 9.1 | 0.5 | 1.4 |
| Total All Taxa | 65.1 | 1.0 | 16.3 | 248.1 | 3.2 | 9.6 | 454.5 | 7.0 | 8.7 | 217.6 | 4.5 | 13.5 | 154.5 | 3.7 | 6.7 |
| Percent Dominant's Density | 96.7 |  |  | 100.0 |  |  | 100.0 |  |  | 94.1 |  |  | 100.0 |  |  |
| Percent Non-dominant's Density | 3.3 |  |  | 0.0 |  |  | 0.0 |  |  | 5.9 |  |  | 0.0 |  |  |
| Species Richness | 7 |  |  | 8 |  |  | 7 |  |  | 6 |  |  | 4 |  |  |

$\%$ OCC - Percent Occurrence (percent of samples in which the species was observed).
STD Error - Standard Error

Table 3.3-6. Comparison of mean density $\left(\# / 100 \mathrm{~m}^{2}\right)$ of selected and non-selected taxa between stations at the Hart reef site.

| Species Name | Early Juvenile |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean Density | $\begin{gathered} \mathrm{HL1} \\ \text { STD } \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{gathered} \underline{\text { HP1 }} \\ \text { STror } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{gathered} \frac{\text { HP2 }}{\text { STD }} \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{gathered} \underset{\text { HP3 }}{\text { STD }} \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \mathrm{OCC} \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 6.4 | 6.4 | 25 | 0.0 | 0.0 | 0 |
| Archosargus probatocephalus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 6.4 | 0.8 | 3.1 | 0.0 | 0.0 | 0.0 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 28.0 | 16.2 | 50 | 530.3 | 330.2 | 50 | 1202.0 | 1054.2 | 50 | 6212.1 | 4038.2 | 50 |
| Larval fish | 3149.5 | 2703.1 | 50 | 0.0 | 0.0 | 0 | 6393.9 | 6393.9 | 25 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 272.3 | 204.5 | 50 | 2083.3 | 1203.2 | 50 | 1457.8 | 935.0 | 50 | 3712.1 | 2360.7 | 50 |
| Diplodus holbrooki | 5.6 | 5.6 | 25 | 113.6 | 113.6 | 25 | 19.2 | 19.2 | 25 | 75.8 | 75.8 | 25 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 75.8 | 75.8 | 25 | 0.0 | 0.0 | 0 | 37.9 | 37.9 | 25 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 3455.3 | 174.6 | 9.7 | 2803.0 | 117.2 | 8.3 | 9072.9 | 361.0 | 8.3 | 10037.9 | 391.0 | 8.3 |
| Total All Taxa | 3455.3 | 121.1 | 6.7 | 2803.0 | 81.7 | 5.8 | 9079.3 | 252.0 | 6.7 | 10037.9 | 273.2 | 5.8 |
| Percent Dominant's Density | 99.0 |  |  | 93.2 |  |  | 99.7 |  |  | 98.9 |  |  |
| Percent Non-dominant's Density | 1.0 |  |  | 6.8 |  |  | 0.3 |  |  | 1.1 |  |  |
| Species Richness | 4 |  |  | 4 |  |  | 5 |  |  | 4 |  |  |

Table 3.3-6. (Continued)

| Species Name | Juvenile |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean <br> Density | $\begin{aligned} & \underline{\text { HL1 }} \\ & \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{gathered} \underset{\substack{\text { HP1 } \\ \text { ETror }}}{ } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \underset{\text { HP2 }}{\text { STD }} \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{gathered} \mathrm{HP3} \\ \text { STD } \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 10.5 | 5.4 | 75 | 37.9 | 37.9 | 25 | 44.8 | 21.8 | 75 | 75.8 | 75.8 | 25 |
| Archosargus probatocephalus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 12.8 | 12.8 | 25 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 10.5 | 1.3 | 9.4 | 37.9 | 4.7 | 3.1 | 57.5 | 5.6 | 12.5 | 75.8 | 9.5 | 3.1 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 5.6 | 4.7 | 50 | 416.7 | 242.5 | 50 | 63.9 | 48.4 | 50 | 454.5 | 321.4 | 50 |
| Diplodus holbrooki | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 4.2 | 3.3 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.7 | 0.7 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 10.5 | 0.4 | 6.9 | 416.7 | 23.1 | 2.8 | 63.9 | 3.6 | 2.8 | 454.5 | 25.3 | 2.8 |
| Total All Taxa | 21.0 | 0.5 | 7.7 | 454.5 | 16.0 | 2.9 | 121.5 | 3.0 | 5.8 | 530.3 | 17.6 | 2.9 |
| Percent Dominant's Density | 96.7 |  |  | 100.0 |  |  | 100.0 |  |  | 100.0 |  |  |
| Percent Non-dominant's Density | 3.3 |  |  | 0.0 |  |  | 0.0 |  |  | 0.0 |  |  |

Table 3.3-6. (Continued)

| Species Name | Subadult |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean <br> Density | $\begin{aligned} & \frac{\text { HL1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \frac{\text { HP1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | $\begin{aligned} & \frac{\text { HP2 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \frac{\text { HP3 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 30.1 | 10.0 | 100 | 151.5 | 87.5 | 50 | 70.3 | 62.1 | 50 | 113.6 | 113.6 | 25 |
| Archosargus probatocephalus | 4.2 | 1.8 | 75 | 0.0 | 0.0 | 0 | 38.4 | 24.5 | 50 | 113.6 | 72.5 | 50 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 2.8 | 1.1 | 75 | 37.9 | 37.9 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 37.1 | 3.7 | 31.3 | 189.4 | 18.9 | 9.4 | 108.7 | 9.4 | 12.5 | 227.3 | 18.6 | 9.4 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 37.9 | 37.9 | 25 |
| Diplodus holbrooki | 1.4 | 1.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 8.4 | 8.4 | 25 | 0.0 | 0.0 | 0 | 19.2 | 19.2 | 25 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 4.2 | 1.4 | 100 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 14.0 | 0.5 | 8.3 | 0.0 | 0.0 | 0.0 | 19.2 | 1.1 | 1.4 | 37.9 | 2.1 | 1.4 |
| Total All Taxa | 51.1 | 1.2 | 15.4 | 189.4 | 6.0 | 2.9 | 127.9 | 3.1 | 4.8 | 265.2 | 6.1 | 3.8 |
| Percent Dominant's Density | 97.3 |  |  | 100.0 |  |  | 100.0 |  |  | 100.0 |  |  |
| Percent Non-dominant's Density | 2.7 |  |  | 0.0 |  |  | 0.0 |  |  | 0.0 |  |  |
| Species Richness | 6 |  |  | 2 |  |  | 3 |  |  | 3 |  |  |

Table 3.3-6. $\quad$ (Continued)

| Species Name | Adult |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean <br> Density | $\begin{aligned} & \frac{\text { HL1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\underset{\text { Error }}{\frac{\text { HP1 }}{\text { STD }}}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{aligned} & \frac{\mathrm{HP2}}{\mathrm{STD}} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \text { HP3 } \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 16.8 | 8.2 | 75 | 37.9 | 37.9 | 25 | 19.2 | 19.2 | 25 | 113.6 | 72.5 | 50 |
| Archosargus probatocephalus | 23.8 | 8.9 | 100 | 492.4 | 272.3 | 75 | 159.8 | 95.3 | 75 | 416.7 | 37.9 | 100 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 113.6 | 72.5 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 1.4 | 0.8 | 50 | 37.9 | 37.9 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 42.0 | 3.4 | 28.1 | 681.8 | 59.8 | 21.9 | 179.0 | 19.8 | 12.5 | 530.3 | 52.0 | 18.8 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplodus holbrooki | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 20.3 | 9.9 | 75 | 189.4 | 72.5 | 75 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.7 | 0.7 | 25 | 75.8 | 43.7 | 50 | 0.0 | 0.0 | 0 | 37.9 | 37.9 | 25 |
| Balistidae spp. | 1.4 | 0.8 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 6.4 | 6.4 | 25 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 37.9 | 37.9 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.7 | 0.7 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 23.1 | 1.1 | 9.7 | 303.0 | 11.1 | 8.3 | 6.4 | 0.4 | 1.4 | 37.9 | 2.1 | 1.4 |
| Total All Taxa | 65.1 | 1.3 | 15.4 | 984.8 | 20.2 | 12.5 | 185.4 | 6.2 | 4.8 | 568.2 | 16.4 | 6.7 |
| Percent Dominant's Density | 93.5 |  |  | 88.5 |  |  | 96.6 |  |  | 100.0 |  |  |
| Percent Non-dominant's Density | 6.5 |  |  | 11.5 |  |  | 3.4 |  |  | 0.0 |  |  |
| Species Richness | 7 |  |  | 7 |  |  | 3 |  |  | 3 |  |  |

\% OCC - Percent Occurrence (percent of samples in which the species was observed).
STD Error - Standard error of the observed density.

Table 3.3-7. Comparison of mean density (\#/100 $\mathrm{m}^{2}$ ) of selected and non-selected taxa between stations at the Walker reef site.

| Species Name | Early-Juvenile |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean Density | $\begin{aligned} & \hline \frac{\text { WL1 }}{} \\ & \hline \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{aligned} & \hline \text { WL2 } \\ & \hline \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | $\begin{aligned} & \frac{\text { WP1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{aligned} & \hline \underline{\text { WP2 }} \\ & \hline \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{aligned} & \hline \text { WP3 } \\ & \hline \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 1.3 | 1.3 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Archosargus probatocephalus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 2.4 | 2.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 1.3 | 0.2 | 3.1 | 2.4 | 0.3 | 3.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 2.5 | 2.5 | 25 | 0.0 | 0.0 | 0 | 189.4 | 189.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 126.3 | 124.6 | 50 | 95.4 | 55.1 | 50 | 0.0 | 0.0 | 0 | 3787.9 | 3787.9 | 25 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 0.0 | 0.0 | 0 | 21.5 | 13.7 | 50 | 4166.7 | 2505.4 | 75 | 606.1 | 396.1 | 50 | 114.5 | 114.5 | 25 |
| Diplodus holbrooki | 125.0 | 125.0 | 25 | 4.8 | 2.8 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 6.3 | 6.3 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 260.0 | 9.5 | 6.9 | 121.7 | 5.4 | 8.3 | 4356.1 | 231.1 | 5.6 | 4393.9 | 211.1 | 4.2 | 114.5 | 6.4 | 1.4 |
| Total All Taxa | 261.3 | 6.7 | 5.8 | 124.0 | 3.7 | 6.7 | 4356.1 | 160.1 | 3.8 | 4393.9 | 146.6 | 2.9 | 114.5 | 4.4 | 1.0 |
| Percent Dominant's Density | 96.2 |  |  | 94.2 |  |  | 95.7 |  |  | 100.0 |  |  | 100.0 |  |  |
| Percent Non-dominant's Density | 3.8 |  |  | 5.8 |  |  | 4.3 |  |  | 0.0 |  |  | 0.0 |  |  |
| Species Richness | 5 |  |  | 4 |  |  | 2 |  |  | 2 |  |  | 1 |  |  |

Table 3.3-7. (Continued)

| Species Name | Juvenile |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean <br> Density | $\begin{aligned} & \hline \frac{\text { WL1 }}{} \\ & \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \frac{\text { WL2 }}{} \\ \hline \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \frac{\text { WP1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \hline \underline{\text { WP2 }} \\ \hline \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \hline \text { WP3 } \\ & \hline \text { STD } \\ & \text { Error } \\ & \hline \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 35.0 | 16.5 | 75 | 57.3 | 23.7 | 100 | 340.9 | 217.6 | 50 | 947.0 | 488.5 | 75 | 248.1 | 100.4 | 100 |
| Archosargus probatocephalus | 0.0 | 0.0 | 0 | 14.3 | 8.3 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 8.8 | 8.8 | 25 | 31.0 | 11.9 | 75 | 113.6 | 37.9 | 75 | 113.6 | 37.9 | 75 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 175.0 | 118.1 | 50 | 2.4 | 2.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 57.3 | 57.3 | 25 |
| Mycteroperca microlepsis | 13.8 | 7.7 | 75 | 28.6 | 22.7 | 50 | 75.8 | 43.7 | 50 | 151.5 | 107.1 | 50 | 57.3 | 57.3 | 25 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 4.8 | 4.8 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 232.5 | 21.3 | 28.1 | 138.4 | 7.2 | 40.6 | 530.3 | 42.2 | 21.9 | 1212.1 | 115.6 | 25.0 | 362.6 | 30.4 | 18.8 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 5.0 | 3.5 | 50 | 7.2 | 4.6 | 50 | 75.8 | 43.7 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 15.0 | 9.6 | 50 | 240.9 | 108.8 | 100 | 1060.6 | 269.6 | 100 | 1174.2 | 378.2 | 100 | 1240.5 | 469.9 | 100 |
| Diplodus holbrooki | 201.3 | 183.3 | 75 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 2.4 | 2.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 55.0 | 48.6 | 50 | 4.8 | 4.8 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 19.1 | 19.1 | 25 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 31.0 | 25.1 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 19.1 | 19.1 | 25 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 276.3 | 11.4 | 12.5 | 286.3 | 13.4 | 13.9 | 1136.4 | 58.8 | 8.3 | 1174.2 | 65.2 | 5.6 | 1278.6 | 68.8 | 8.3 |
| Total All Taxa | 508.8 | 10.1 | 17.3 | 424.6 | 9.4 | 22.1 | 1666.7 | 42.2 | 12.5 | 2386.4 | 56.7 | 11.5 | 1641.2 | 48.1 | 11.5 |
| Percent Dominant's Density | 91.6 |  |  | 91.6 |  |  | 90.9 |  |  | 95.2 |  |  | 90.7 |  |  |
| Percent Non-dominant's Density | 8.4 |  |  | 8.4 |  |  | 9.1 |  |  | 4.8 |  |  | 9.3 |  |  |
| Species Richness | 8 |  |  | 11 |  |  | 5 |  |  | 4 |  |  | 6 |  |  |

Table 3.3-7. (Continued)

| Species Name | Subadult |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean <br> Density | $\begin{gathered} \frac{\text { WL1 }}{} \\ \hline \text { STD } \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \frac{\text { WL2 }}{} \\ \hline \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \frac{\text { WP1 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{gathered} \hline \underline{\text { WP2 }} \\ \hline \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \hline \text { WP3 } \\ & \hline \text { STD } \\ & \text { Error } \\ & \hline \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 68.8 | 41.9 | 50 | 202.8 | 158.7 | 50 | 1250.0 | 892.9 | 50 | 1136.4 | 901.7 | 50 | 209.9 | 129.9 | 50 |
| Archosargus probatocephalus | 8.8 | 7.2 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 15.0 | 5.8 | 100 | 97.8 | 27.4 | 100 | 189.4 | 37.9 | 100 | 151.5 | 61.9 | 75 | 114.5 | 49.3 | 75 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 7.2 | 7.2 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 22.5 | 6.6 | 100 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 5.0 | 3.5 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 19.1 | 19.1 | 25 |
| Subtotal Selected Taxa | 120.0 | 8.2 | 43.8 | 307.7 | 26.4 | 21.9 | 1439.4 | 154.7 | 18.8 | 1287.9 | 140.6 | 15.6 | 343.5 | 27.6 | 18.8 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 2.5 | 1.4 | 50 | 9.5 | 6.7 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 400.0 | 234.5 | 50 | 276.7 | 191.3 | 50 | 681.8 | 412.6 | 50 | 681.8 | 464.9 | 50 | 954.2 | 954.2 | 25 |
| Diplodus holbrooki | 15.0 | 15.0 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 2.5 | 2.5 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 21.3 | 18.1 | 50 | 11.9 | 11.9 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 95.4 | 95.4 | 25 |
| Haemulon plumieri | 1.3 | 1.3 | 25 | 4.8 | 2.8 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 2.4 | 2.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 2.4 | 2.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 442.5 | 22.1 | 12.5 | 307.7 | 15.3 | 12.5 | 681.8 | 37.9 | 2.8 | 681.8 | 37.9 | 2.8 | 1049.6 | 53.0 | 2.8 |
| Total All Taxa | 562.5 | 15.4 | 22.1 | 615.5 | 13.2 | 15.4 | 2121.2 | 53.9 | 7.7 | 1969.7 | 50.1 | 6.7 | 1393.1 | 37.3 | 7.7 |
| Percent Dominant's Density | 83.3 |  |  | 93.8 |  |  | 100.0 |  |  | 100.0 |  |  | 98.6 |  |  |
| Percent Non-dominant's Density | 16.7 |  |  | 6.2 |  |  | 0.0 |  |  | 0.0 |  |  | 1.4 |  |  |
| Species Richness | 11 |  |  | 9 |  |  | 3 |  |  | 3 |  |  | 5 |  |  |

Table 3.3-7. (Continued)

| Species Name | Adult |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean Density | $\begin{aligned} & \hline \frac{\text { WL1 }}{} \\ & \hline \text { STD } \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \frac{\text { WL2 }}{\text { STD }} \\ & \text { Error } \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean <br> Density | $\begin{gathered} \hline \text { WP1 } \\ \hline \text { STD } \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | $\begin{gathered} \hline \text { WP2 } \\ \hline \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | $\begin{aligned} & \hline \text { WP3 } \\ & \hline \text { STD } \\ & \text { Error } \\ & \hline \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 50.0 | 28.9 | 50 | 7.2 | 7.2 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Archosargus probatocephalus | 37.5 | 5.2 | 100 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 37.5 | 4.8 | 100 | 136.0 | 11.9 | 100 | 303.0 | 61.9 | 100 | 151.5 | 0.0 | 100 | 209.9 | 36.5 | 100 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 8.8 | 2.4 | 100 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 19.1 | 7.8 | 75 | 75.8 | 75.8 | 25 | 0.0 | 0.0 | 0 | 38.2 | 22.0 | 50 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 133.8 | 7.5 | 43.8 | 162.2 | 16.7 | 25.0 | 378.8 | 37.7 | 15.6 | 151.5 | 18.9 | 12.5 | 248.1 | 26.0 | 18.8 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 11.3 | 6.6 | 50 | 2.4 | 2.4 | 25 | 0.0 | 0.0 | 0 | 37.9 | 37.9 | 25 | 19.1 | 19.1 | 25 |
| Diplodus holbrooki | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 2.4 | 2.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 6.3 | 3.8 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 2.5 | 2.5 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 2.4 | 2.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 1.3 | 1.3 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 21.3 | 0.7 | 8.3 | 7.2 | 0.2 | 4.2 | 0.0 | 0.0 | 0.0 | 37.9 | 2.1 | 1.4 | 19.1 | 1.1 | 1.4 |
| Total All Taxa | 155.0 | 2.7 | 19.2 | 169.4 | 5.2 | 10.6 | 378.8 | 11.9 | 4.8 | 189.4 | 6.0 | 4.8 | 267.2 | 8.1 | 6.7 |
| Percent Dominant's Density | 93.5 |  |  | 91.5 |  |  | 100.0 |  |  | 100.0 |  |  | 100.0 |  |  |
| Percent Non-dominant's Density | 6.5 |  |  | 8.5 |  |  | 0.0 |  |  | 0.0 |  |  | 0.0 |  |  |
| Species Richness | 8 |  |  | 6 |  |  | 2 |  |  | 2 |  |  | 3 |  |  |

$\%$ OCC - Percent Occurrence (percent of samples in which the species was observed).
STD Error - Standard error of the observed densities.

Juvenile species density was greatest at station GP2 $\left(1,174.2 / 100 \mathrm{~m}^{2}\right)$ while percent occurrence ( $15.4 \%$ ) and richness ( 7 species) were greatest at station GL1. Juveniles of Lutjanus griseus were consistently observed across all stations but reached their highest density at station GP2 ( $719.7 / 100 \mathrm{~m}^{2}$ ). Archosargus probatocephalus also reached its highest density at station GP2 $\left(265.2 / 100 \mathrm{~m}^{2}\right)$, but were fairly well represented among all stations. While Menippe mercenaria juveniles reached their greatest density at station GP3 (17.1/100 $\mathrm{m}^{2}$ ), they were more frequently observed at station GL1 (100\%). Juvenile Mycteroperca microlepsis were only observed at stations GP1 and GP4. Juvenile Lagodon rhomboides were observed consistently across all five stations and reached their greatest density at station GP3 ( $302.9 / 100 \mathrm{~m}^{2}$ ). Chaetodipeterus faber was the only other species to occur at more than one station (occurring at three stations), reaching its highest density (59.7/100 $\mathrm{m}^{2}$ ) and percent occurrence ( $50 \%$ ) at station GP3.

The density of subadults was greatest at station GP1 $\left(916.0 / 100 \mathrm{~m}^{2}\right)$, while the percent occurrence (11.5\%) was greatest at station GP3 and species richness ( 6 species) was greatest at station GL1. Subadult Lutjanus griseus and Archosargus probatocephalus were consistently observed across all five stations, yet reached their greatest density at station GP1 (572.5/100 $\mathrm{m}^{2}$ and 209.9/100 $\mathrm{m}^{2}$, respectively). Subadult Menippe mercenaria were only observed at GP2 $\left(75.8 / 100 \mathrm{~m}^{2}\right)$ and subadult Mycteroperca microlepsis were only observed on rare occasion at station GL1 $\left(1.1 / 100 \mathrm{~m}^{2}\right)$. Subadult Lagodon rhomboides were observed at most stations and reached their highest density at station GP1 (95.4/100 m $\left.\mathrm{m}^{2}\right)$. Subadult Chaetodipeterus faber reached high density at stations GP3 (384.0/100 $\mathrm{m}^{2}$ ) and GP4 (181.8/100 $\mathrm{m}^{2}$ ) and were also present at station GL1 ( $48.0 / 100 \mathrm{~m}^{2}$ ). Other subadult individuals of the non-select taxa were only rarely observed.

Although adult Lutjanus griseus reached their highest density at station GP4 (90.9/100 m ${ }^{2}$ ), they were observed every event $(100 \%)$ at station GL1. Adult Archosargus probatocephalus were observed at all sites and reached their greatest density at station GP3 (85.3/100 $\mathrm{m}^{2}$ ). Adult Menippe mercenaria were also observed at all stations, but reached their greatest density at station GP2 $\left(151.5 / 100 \mathrm{~m}^{2}\right)$. Adult Mycteroperca microlepsis were
only observed during one event at one station (GP3). Adult Diplodus holbrooki were present at all five stations, reaching their greatest density at stations GP1 $\left(38.2 / 100 \mathrm{~m}^{2}\right)$ and GP2 (37.9/100 $\mathrm{m}^{2}$ ). Adult Lagodon rhomboides were present at three of the stations and reached their greatest density at station GP2 $\left(37.9 / 100 \mathrm{~m}^{2}\right)$. All other adult specimens occurred on rare occasion at individual stations.

## HART

For the Hart reef site (Table 3.3-6), the early-juvenile species richness was greatest at station HP2 (5 species), while density was greatest at site HP3 (10,037.9/100 $\mathrm{m}^{2}$ ) and percent occurrence (6.7\%) was highest at stations HP1 and HP2. Early-juveniles of selected taxa were extremely rare at the Hart site, being observed during only one event at station HP2. Haemulon aurolineatum and Lagodon rhomboides early-juveniles were observed at all stations, reaching their greatest density at station HP3 $\left(6,212.1 / 100 \mathrm{~m}^{2}\right.$ and $3,712.1 / 100 \mathrm{~m}^{2}$, respectively), while larval fish were observed at only two stations (HL1 and HP2) and reached their greatest density at station HP2 $\left(6,393.9 / 100 \mathrm{~m}^{2}\right)$. Early-juveniles of Diplodus holbrooki were present at all four stations and reached their highest density at station HP1 (113.6/100 m²).

Juveniles were relatively rare at the Hart reef site, with two species of selected taxa and three species of non-selected taxa being observed. The overall mean density of juveniles was greatest at station $\operatorname{HP} 3\left(530.3 / 100 \mathrm{~m}^{2}\right)$ while the percent occurrence $(7.7 \%)$ and species richness (4 species) were greatest at station HL1. Juveniles of Lutjanus griseus were observed at all four stations, reached their greatest density at station HP3 (75.8/100 m$\left.{ }^{2}\right)$, but were most frequently observed at stations HL1 (75\%) and HP2 (75\%). Archosargus probatocephalus was only observed at one station during one event, that being station HP2. Juveniles of Lagodon rhomboides were consistently observed at all four stations, reaching their highest density at stations HP1 (416.7/100 m ${ }^{2}$ ) and HP3 (454.5/100 m$\left.{ }^{2}\right)$.

For subadults, the species richness (6 species) and percent occurrence ( $15.4 \%$ ) were substantially greater for station HL1, while the mean density $\left(265.2 / 100 \mathrm{~m}^{2}\right)$ was greatest at
station HP3. For most species, the percent occurrence was always greatest at station HL1 while the mean density could be greater at other sites. For selected taxa, subadults of Lutjanus griseus (151.5/100 $\mathrm{m}^{2}$ ) and Mycteroperca microlepsis ( $37.9 / 100 \mathrm{~m}^{2}$ ) were most dense at station HP1, while for Archosargus probatocephalus the mean density was greatest at station HP3 $\left(113.6 / 100 \mathrm{~m}^{2}\right)$. Non-select taxa in the subadult size class were relatively rare, generally being observed during only one of the surveys and in low numbers.

For the adult size class, mean density was greatest at station HP1 (984.8 $100 \mathrm{~m}^{2}$ ), the percent occurrence ( $15.4 \%$ ) was greatest at station HL1, and species richness ( 7 species) was highest at stations HL1 and HP1. With the exception of Lutjanus griseus, which had its greatest mean density at station HP3 (113.6/100 $\left.\mathrm{m}^{2}\right)$, the mean density of all other selected taxa was greatest at station HP1. Chaetodipeterus faber was the non-selected species which occurred at the greatest density $\left(189.4 / 100 \mathrm{~m}^{2}\right)$ as an adult, being observed at stations HL1 and HP1 during three of the sampling events (75\%). The remaining non-selected taxa occurred very rarely at any of the stations.

## WALKER

The data for the stations at the Walker site are presented in Table 3.3-7. Species richness for early-juveniles was greatest at station WL1 ( 5 species), percent occurrence was greatest at station WL2 (6.7\%) and the mean density was greatest at station WP2 $\left(4,393.9 / 100 \mathrm{~m}^{2}\right)$. Early-juveniles of selected taxa were rarely observed at the Walker reef site stations, with two species being observed during one event. For the non-selected taxa, Lagodon rhomboides had the greatest early-juvenile mean density at station WP1 (4,116.7/100 m ${ }^{2}$ ) and was found at all stations except station WL1. The mean density for Haemulon aurolineatum was also greatest at station WP1 (189.4/100 $\mathrm{m}^{2}$ ), while larval fish density was greatest at station WP2 (3,787.9/100 m²). Diplodus holbrooki was only observed at stations WL1 and WL2, having its greatest mean density at station WL1 (125.0/100 m ${ }^{2}$ ).

For the juvenile size class, mean density was greatest at station WP2 $\left(2,386.4 / 100 \mathrm{~m}^{2}\right)$, while species richness ( 11 species) and a percent occurrence ( $22.1 \%$ ) were greatest at station

WL2. Juveniles of Lutjanus griseus were observed consistently at all stations, reaching their greatest mean density at station WP1 ( $340.9 / 100 \mathrm{~m}^{2}$ ). Menippe mercenaria were observed at four of the five stations, while Archosargus probatocephalus was only observed at one of the stations (WL2). Juvenile Mycteroperca microlepsis were also observed at all stations, reaching their greatest density at station WP2 $\left(151.5 / 100 \mathrm{~m}^{2}\right)$. Lagodon rhomboides were observed consistently at all five stations and reached their greatest density at station WP3 $\left(1,240.5 / 100 \mathrm{~m}^{2}\right)$. Haemulon aurolineatum were observed in low numbers at three of the five stations.

In the subadult size class, mean density was greatest at station WP1 $\left(2,121.2 / 100 \mathrm{~m}^{2}\right)$, with species richness (11 species) and percent occurrence (22.1\%) greatest at station WL1. Subadults Lutjanus griseus were consistently observed at all five stations, reaching the greatest density at station WP2 $\left(1,136.4 / 100 \mathrm{~m}^{2}\right)$. Menippe mercenaria were also observed consistently at all stations but reached their greatest density at station WP1 (189.4/100 m ${ }^{2}$ ). Three species (Archosargus probatocephalus, Mycteroperca microlepsis and Mycteroperca phenax) were only observed as subadults at station WL1. Subadults of Lagodon rhomboides were also observed consistently at all five stations, reaching their greatest density at station WP3 (954.2/100 $\mathrm{m}^{2}$ ). Subadults at all other non-selected taxa occurred rarely at only a few stations.

Results for the adult size class at the Walker reef class were consistent with that for the subadults (i.e.; density was greatest for station WP1 ( $378.8 / 100 \mathrm{~m}^{2}$ ), while percent occurrence ( $19 \%$ ) and species richness ( 8 species) were greatest at Station WL1). Adult Lutjanus griseus were only observed at the two linear transects (WL1 and WL2). Menippe mercenaria were observed consistently during every event at every station and reached their greatest density at Station WP1 ( $303.0 / 100 \mathrm{~m}^{2}$ ). Archosargus probatocephalus and Mycteroperca microlepsis adults were only observed at one station, WL1, but were observed during every event $(100 \%)$ at that station. Adults of all other selected and non-selected taxa occurred only rarely at a few stations and in low numbers.

Table 3.3-8 provides a comparison of the mean density of selected and non-selected taxa by reef material type between all three reef sites. For the early-juvenile size class, the PVC ribcage has the overall highest mean density ( $12,188.0 / 100 \mathrm{~m}^{2}$ ), with species richness being highest for reef balls ( 7 species) and the percent occurrence ( $6.7 \%$ ) being highest with both the concrete block and the concrete block with reef ball material types. Early-juveniles of Lutjanus griseus were observed on all reef material types, except the concrete block with reef balls material type, reaching their greatest density on the PVC web design (63.6/100 $\mathrm{m}^{2}$ ). Early-juveniles of Archosargus probatocephalus and Menippe mercenaria were only rarely observed, being found on a single reef material type. Early-juveniles of Haemulon aurolineatum reached their greatest densities on the PVC ribcage material type (9,396.3/100 $\mathrm{m}^{2}$ ), were consistently found on most material types, but were noticeably absent on the PVC web. Larval fish were also noticeably absent on the PVC web material, reaching their greatest density on the concrete block design $\left(6,393.9 / 100 \mathrm{~m}^{2}\right)$. Except for their absence on the reef ball in concrete material type, Lagodon rhomboides was consistently observed in high numbers on the other material types, reaching its greatest density on the PVC ribcage design (2,522.4/100 m${ }^{2}$ ). Diplodus holbrooki was generally in low numbers, reached its greatest density on the reef ball in concrete material type (125.0/100 $\mathrm{m}^{2}$ ), and was absent from the PVC ribcage and PVC web material type.

The juvenile size class greatest mean density ( $1,180.3 / 100 \mathrm{~m}^{2}$ ) and largest species richness ( 12 species) occurred on the reef ball material type, and the reef ball in concrete material type had the greatest percent occurrence (17.3\%). For selected taxa, juveniles of all species reached their greatest mean density $\left(175.0 / 100 \mathrm{~m}^{2}\right)$ on the reef ball design with the exception of Leiostomus xanthurus, which reached its greatest density on the reef ball and concrete design. Juvenile Lutjanus griseus were present across all of the material types, while Archosargus probatocephalus was present across all but the concrete block/reef ball and reef ball in concrete designs. Juvenile Menippe mercenaria reached their greatest percent occurrence (75\%) on the PVC ribcage design and greatest density on reef balls (32.3/100 $\mathrm{m}^{2}$ ), but were absent from the concrete block, concrete block and reef ball and PVC web designs. Juvenile Mycteroperca microlepsis had their greatest percent occurrence (75\%) in

Table 3.3-8. Comparison of mean density $\left(\# / 100 \mathrm{~m}^{2}\right)$ of selected and non-selected taxa between reef material types.

| Species Name | Early-Juvenile |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  CB  <br> Mean STD \% <br> Density Error OCC |  |  | CB,RB(3) |  |  | PVC RB |  |  | PVC WEB |  |  | RB |  |  | RB IN C |  |  |
|  |  |  |  | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \mathrm{OCC} \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | $\begin{gathered} \text { STD } \\ \text { Error } \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 6.4 | 6.4 | 25 | 0.0 | 0.0 | 0 | 0.5 | 0.5 | 12.5 | 63.6 | 63.6 | 25 | 4.7 | 4.7 | 3 | 1.3 | 1.3 | 25 |
| Archosargus probatocephalus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 2.1 | 2.1 | 12.5 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.3 | 0.3 | 3 | 0.0 | 0.0 | 0 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 6.4 | 0.8 | 3.1 | 0.0 | 0.0 | 0.0 | 2.7 | 0.3 | 3.1 | 63.6 | 8.0 | 3.1 | 5.0 | 0.6 | 0.8 | 1.3 | 0.2 | 3.1 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 1202.0 | 1054.2 | 50 | 28.0 | 16.2 | 50 | 9396.3 | 5515.3 | 37.5 | 0.0 | 0.0 | 0 | 866.5 | 575.9 | 16 | 2.5 | 2.5 | 25 |
| Larval fish | 6393.9 | 6393.9 | 25 | 3149.5 | 2703.1 | 50 | 266.6 | 266.6 | 12.5 | 0.0 | 0.0 | 0 | 485.4 | 473.2 | 9 | 126.3 | 124.6 | 50 |
| Lagodon rhomboides | 1457.8 | 935.0 | 50 | 272.3 | 204.5 | 50 | 2522.4 | 1692.6 | 50 | 27.3 | 17.4 | 50 | 1858.9 | 658.5 | 47 | 0.0 | 0.0 | 0 |
| Diplodus holbrooki | 19.2 | 19.2 | 25 | 5.6 | 5.6 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 24.3 | 16.8 | 13 | 125.0 | 125.0 | 25 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 14.2 | 10.4 | 6 | 6.3 | 6.3 | 25 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 9072.9 | 361.0 | 8.3 | 3455.3 | 174.6 | 9.7 | 12185.4 | 531.6 | 5.6 | 27.3 | 1.5 | 2.8 | 3249.3 | 112.3 | 5.0 | 260.0 | 9.5 | 6.9 |
| Total All Taxa | 9079.3 | 252.0 | 6.7 | 3455.3 | 121.1 | 6.7 | 12188.0 | 370.1 | 4.8 | 90.9 | 2.6 | 2.9 | 3254.3 | 78.8 | 3.7 | 261.3 | 6.7 | 5.8 |
| Percent Dominant's Density | 99.7 |  |  | 99.0 |  |  | 97.8 |  |  | 100.0 |  |  | 98.7 |  |  | 96.2 |  |  |
| Percent Non-dominant's Density | 0.3 |  |  | 1.0 |  |  | 2.2 |  |  | 0.0 |  |  | 1.3 |  |  | 3.8 |  |  |
| Species Richness | 5 |  |  | 4 |  |  | 5 |  |  | 2 |  |  | 7 |  |  | 5 |  |  |

Table 3.3-8. (Continued)

| Species Name | Juvenile |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CB |  |  | CB,RB(3) |  |  | PVC RB |  |  | PVC WEB |  |  | RB |  |  | RB IN C |  |  |
|  | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | $\begin{gathered} \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 44.8 | 21.8 | 75 | 10.5 | 5.4 | 75 | 158.4 | 74.7 | 87.5 | 154.5 | 52.2 | 75 | 391.6 | 99.2 | 69 | 35.0 | 16.5 | 75 |
| Archosargus probatocephalus | 12.8 | 12.8 | 25 | 0.0 | 0.0 | 0 | 12.8 | 8.3 | 50 | 45.5 | 17.4 | 75 | 65.9 | 27.9 | 25 | 0.0 | 0.0 | 0 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 11.7 | 6.0 | 75 | 0.0 | 0.0 | 0 | 32.3 | 10.5 | 28 | 8.8 | 8.8 | 25 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 7.5 | 7.2 | 6 | 175.0 | 118.1 | 50 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 27.3 | 17.4 | 50 | 51.1 | 18.2 | 28 | 13.8 | 7.7 | 75 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.6 | 0.6 | 3 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 57.5 | 5.6 | 12.5 | 10.5 | 1.3 | 9.4 | 182.9 | 19.5 | 26.6 | 227.3 | 19.0 | 25.0 | 548.9 | 47.0 | 19.9 | 232.5 | 21.3 | 28.1 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 10.4 | 6.6 | 13 | 5.0 | 3.5 | 50 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 63.9 | 48.4 | 50 | 5.6 | 4.7 | 50 | 179.7 | 125.9 | 50 | 27.3 | 17.4 | 50 | 611.4 | 117.9 | 75 | 15.0 | 9.6 | 50 |
| Diplodus holbrooki | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.5 | 0.5 | 12.5 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 201.3 | 183.3 | 75 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 33.6 | 21.2 | 50 | 27.3 | 27.3 | 25 | 0.3 | 0.3 | 3 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.6 | 0.6 | 3 | 55.0 | 48.6 | 50 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 4.2 | 3.3 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 2.4 | 2.4 | 3 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.5 | 0.5 | 12.5 | 0.0 | 0.0 | 0 | 6.3 | 4.0 | 9 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.7 | 0.7 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 63.9 | 3.6 | 2.8 | 10.5 | 0.4 | 6.9 | 214.4 | 10.0 | 6.9 | 54.5 | 2.1 | 4.2 | 631.3 | 33.9 | 5.9 | 276.3 | 11.4 | 12.5 |
| Total All Taxa | 121.5 | 3.0 | 5.8 | 21.0 | 0.5 | 7.7 | 397.3 | 9.0 | 13.0 | 281.8 | 6.2 | 10.6 | 1180.3 | 27.2 | 10.2 | 508.8 | 10.1 | 17.3 |
| Percent Dominant's Density | 100.0 |  |  | 96.7 |  |  | 93.6 |  |  | 100.0 |  |  | 90.6 |  |  | 91.6 |  |  |
| Percent Non-dominant's Density | 0.0 |  |  | 3.3 |  |  | 6.4 |  |  | 0.0 |  |  | 9.4 |  |  | 8.4 |  |  |
| Species Richness | 3 |  |  | 4 |  |  | 7 |  |  | 5 |  |  | 12 |  |  | 8 |  |  |

Table 3.3-8. (Continued)

| Species Name | Subadult |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | CB |  | CB,RB(3) |  |  | PVC RB |  |  | PVC WEB |  |  | RB |  |  | RB IN C |  |  |
|  | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \mathrm{OCC} \\ \hline \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{gathered} \hline \text { STD } \\ \text { Error } \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \mathrm{OCC} \\ \hline \end{gathered}$ | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \mathrm{OCC} \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \mathrm{OCC} \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 70.3 | 62.1 | 50 | 30.1 | 10.0 | 100 | 90.1 | 29.7 | 100 | 272.7 | 90.9 | 100 | 511.4 | 166.2 | 50 | 68.8 | 41.9 | 50 |
| Archosargus probatocephalus | 38.4 | 24.5 | 50 | 4.2 | 1.8 | 75 | 27.7 | 9.7 | 75 | 36.4 | 14.8 | 75 | 54.6 | 24.7 | 16 | 8.8 | 7.2 | 50 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 78.6 | 16.7 | 50 | 15.0 | 5.8 | 100 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.9 | 0.9 | 3 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 2.8 | 1.1 | 75 | 0.5 | 0.5 | 12.5 | 0.0 | 0.0 | 0 | 4.7 | 4.7 | 3 | 22.5 | 6.6 | 100 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 5.0 | 3.5 | 50 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 2.4 | 2.4 | 3 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 108.7 | 9.4 | 12.5 | 37.1 | 3.7 | 31.3 | 118.4 | 11.3 | 23.4 | 309.1 | 33.7 | 21.9 | 652.7 | 62.3 | 15.6 | 120.0 | 8.2 | 43.8 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 1.2 | 0.9 | 6 | 2.5 | 1.4 | 50 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 18.7 | 14.6 | 37.5 | 9.1 | 9.1 | 25 | 341.0 | 142.2 | 31 | 400.0 | 234.5 | 50 |
| Diplodus holbrooki | 0.0 | 0.0 | 0 | 1.4 | 1.4 | 25 | 1.1 | 1.1 | 12.5 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 15.0 | 15.0 | 25 |
| Chaetodipterus faber | 19.2 | 19.2 | 25 | 8.4 | 8.4 | 25 | 216.0 | 122.4 | 50 | 181.8 | 181.8 | 25 | 0.0 | 0.0 | 0 | 2.5 | 2.5 | 25 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 13.4 | 12.0 | 6 | 21.3 | 18.1 | 50 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 4.2 | 1.4 | 100 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 5.4 | 4.8 | 9 | 1.3 | 1.3 | 25 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 9.1 | 9.1 | 25 | 0.3 | 0.3 | 3 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.3 | 0.3 | 3 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Non-selected Taxa | 19.2 | 1.1 | 1.4 | 14.0 | 0.5 | 8.3 | 235.7 | 12.0 | 5.6 | 200.0 | 10.1 | 4.2 | 361.6 | 18.9 | 3.3 | 442.5 | 22.1 | 12.5 |
| Total All Taxa | 127.9 | 3.1 | 4.8 | 51.1 | 1.2 | 15.4 | 354.1 | 8.9 | 11.1 | 509.1 | 12.3 | 9.6 | 1014.3 | 23.1 | 7.1 | 562.5 | 15.4 | 22.1 |
| Percent Dominant's Density | 100.0 |  |  | 97.3 |  |  | 99.5 |  |  | 96.4 |  |  | 97.2 |  |  | 83.3 |  |  |
| Percent Non-dominant's Density | 0.0 |  |  | 2.7 |  |  | 0.5 |  |  | 3.6 |  |  | 2.8 |  |  | 16.7 |  |  |
| Species Richness | 3 |  |  | 6 |  |  | 6 |  |  | 5 |  |  | 12 |  |  | 11 |  |  |

Table 3.3-8. (Continued)

| Species Name | Adult |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CB |  |  | CB,RB(3) |  |  | PVC RB |  |  | PVC WEB |  |  | RB |  |  | RB IN C |  |  |
|  | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \text { \% } \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean Density | $\begin{aligned} & \text { STD } \\ & \text { Error } \\ & \hline \end{aligned}$ | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ | Mean <br> Density | STD <br> Error | $\begin{gathered} \% \\ \text { OCC } \\ \hline \end{gathered}$ |
| Selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lutjanus griseus | 19.2 | 19.2 | 25 | 16.8 | 8.2 | 75 | 52.3 | 23.3 | 87.5 | 90.9 | 52.5 | 50 | 31.7 | 13.5 | 19 | 50.0 | 28.9 | 50 |
| Archosargus probatocephalus | 159.8 | 95.3 | 75 | 23.8 | 8.9 | 100 | 49.6 | 19.0 | 100 | 18.2 | 18.2 | 25 | 127.9 | 46.6 | 31 | 37.5 | 5.2 | 100 |
| Menippe mercenaria | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 9.1 | 6.4 | 37.5 | 36.4 | 14.8 | 75 | 138.0 | 22.9 | 69 | 37.5 | 4.8 | 100 |
| Leiostomus xanthurus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Mycteroperca microlepsis | 0.0 | 0.0 | 0 | 1.4 | 0.8 | 50 | 2.1 | 2.1 | 12.5 | 0.0 | 0.0 | 0 | 4.7 | 4.7 | 3 | 8.8 | 2.4 | 100 |
| Callinectes sapidus | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 16.6 | 9.9 | 19 | 0.0 | 0.0 | 0 |
| Mycteroperca phenax | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lutjanus synagris | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Subtotal Selected Taxa | 179.0 | 19.8 | 12.5 | 42.0 | 3.4 | 28.1 | 113.1 | 8.1 | 29.7 | 145.5 | 11.4 | 18.8 | 318.9 | 20.7 | 17.6 | 133.8 | 7.5 | 43.8 |
| Non-selected Taxa |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Haemulon aurolineatum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Larval fish | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Lagodon rhomboides | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 4.8 | 2.7 | 37.5 | 0.0 | 0.0 | 0 | 12.2 | 6.9 | 13 | 11.3 | 6.6 | 50 |
| Diplodus holbrooki | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 13.3 | 6.4 | 50 | 9.1 | 9.1 | 25 | 9.5 | 6.6 | 6 | 0.0 | 0.0 | 0 |
| Chaetodipterus faber | 0.0 | 0.0 | 0 | 20.3 | 9.9 | 75 | 6.4 | 5.8 | 25 | 0.0 | 0.0 | 0 | 23.7 | 13.8 | 9 | 0.0 | 0.0 | 0 |
| Orthopristis chrysoptera | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 2.7 | 2.4 | 6 | 0.0 | 0.0 | 0 |
| Haemulon plumieri | 0.0 | 0.0 | 0 | 0.7 | 0.7 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 16.6 | 8.1 | 13 | 0.0 | 0.0 | 0 |
| Balistidae spp. | 0.0 | 0.0 | 0 | 1.4 | 0.8 | 50 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Eucinostomus spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Blenniidae spp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 3.7 | 2.0 | 37.5 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Gobiidae sp. | 6.4 | 6.4 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 6.3 | 3.8 | 50 |
| Opsanus beta | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 7.1 | 5.2 | 6 | 2.5 | 2.5 | 25 |
| Diplectrum formosum | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 7.1 | 5.2 | 6 | 0.0 | 0.0 | 0 |
| Prionotus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 9.5 | 6.6 | 6 | 0.0 | 0.0 | 0 |
| Acanthostracion quadricornis | 0.0 | 0.0 | 0 | 0.7 | 0.7 | 25 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Calamus sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 |
| Chilomycterus schoepfi | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.3 | 0.3 | 3 | 0.0 | 0.0 | 0 |
| Scorpaenidae sp. | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 | 0 | 1.3 | 1.3 | 25 |
| Subtotal Non-selected Taxa | 6.4 | 0.4 | 1.4 | 23.1 | 1.1 | 9.7 | 28.3 | 0.8 | 8.3 | 9.1 | 0.5 | 1.4 | 88.6 | 1.7 | 3.8 | 21.3 | 0.7 | 8.3 |
| Total All Taxa | 185.4 | 6.2 | 4.8 | 65.1 | 1.3 | 15.4 | 141.3 | 2.7 | 14.9 | 154.5 | 3.7 | 6.7 | 407.5 | 7.0 | 8.1 | 155.0 | 2.7 | 19.2 |
| Percent Dominant's Density | 96.6 |  |  | 93.5 |  |  | 87.9 |  |  | 100.0 |  |  | 78.8 |  |  | 93.5 |  |  |
| Percent Non-dominant's Density | 3.4 |  |  | 6.5 |  |  | 12.1 |  |  | 0.0 |  |  | 21.2 |  |  | 6.5 |  |  |
| Species Richness | 3 |  |  | 7 |  |  | 8 |  |  | 4 |  |  | 14 |  |  | 8 |  |  |

\% OCC - Percent Occurrence (percent of samples in
CB - Concrete Block, CB, RB (3) - Concrete Block/Reef Balls, PVC RB - PVC Ribcage, PVC - PVC WEB - Web Design, RB - Reef Balls, RB IN C - Reef Balls in Concrete
the reef ball in concrete design and were absent from the PVC ribcage, concrete block and concrete block and reef ball design types. Juveniles of Lagodon rhomboides were consistently observed across all material types, reaching their greatest density (611.4/100 $\mathrm{m}^{2}$ ) and percent occurrence ( $75 \%$ ) on the reef ball designs. Juvenile Diplodus holbrooki were at a substantially higher density $\left(201.3 / 100 \mathrm{~m}^{2}\right)$ on the reef ball in concrete design and only occurred on one other reef material type (PVC ribcage) in low numbers. Juveniles of all other non-selected and select taxa generally were in low numbers and sporadically spread through material types.

The subadult size class reached its greatest density $\left(1,014.3 / 100 \mathrm{~m}^{2}\right)$ as well as its greatest richness ( 12 species) on reef balls, while the highest percent occurrence ( $22.1 \%$ ) occurred in the reef ball and concrete material type. Subadult Lutjanus griseus were commonly found on all material types (50-75\%), were most dense on the reef ball habitat type ( $511.4 / 100 \mathrm{~m}^{2}$ ). Archosargus probatocephalus also had their greatest density on the reef ball material type ( $54.6 / 100 \mathrm{~m}^{2}$ ), but had a lower percent occurrence on this material type ( $16 \%$ ) than on other material types. Subadults of Menippe mercenaria were only found on the reef ball and reef ball in concrete material types, being found most frequently on the reef ball in concrete material $(100 \%)$ but having a greater density on the reef ball material $\left(78.6 / 100 \mathrm{~m}^{2}\right)$. Mycteroperca microlepsis reached its greatest density $\left(22.5 / 100 \mathrm{~m}^{2}\right)$ and percent occurrence ( $100 \%$ ) on the reef ball in concrete material type and was not observed in association with the concrete block and PVC web materials. Subadults of Lagodon rhomboides reached their greatest density ( $400.0 / 100 \mathrm{~m}^{2}$ ) and percent occurrence ( $50 \%$ ) on the reef ball in concrete material type and were noticeably absent from the concrete block and concrete block/reef ball material type. Subadults of Chaetodipeterus faber reached their greatest mean density (216.0/100 $\mathrm{m}^{2}$ ) and percent occurrence ( $50 \%$ ) at the PVC ribcage design, and were noticeably absent from the reef ball material type.

The mean density ( $407.5 / 100 \mathrm{~m}^{2}$ ) and species richness ( 14 species) of adults were greatest on the reef ball habitat type, while species occurrence (19.2\%) was greatest on the reef ball in concrete material type. Adults of Lutjanus griseus were present at all sites, reaching their
greatest density on the PVC web material type ( $90.9 / 100 \mathrm{~m}^{2}$ ) and the greatest percent occurrence ( $87.5 \%$ ) on the PVC ribcage design. Archosargus probatocephalus adults were consistently found at almost all material types and reached their greatest density (159.8/100 $\mathrm{m}^{2}$ ) on the concrete block material type, with low numbers being found at the PVC web material type. Adult Menippe mercenaria reached their greatest mean density (138.0/100 $\mathrm{m}^{2}$ ) on the reef ball material type and reach their greatest occurrence $(100 \%)$ on the reef ball in concrete material types, but were noticeably absent from the concrete block and concrete block/reef ball material types. Mycteroperca microlepsis adults were noticeably missing from the concrete block and PVC web designs and occurred in low numbers on the other four material types. However, this species occurred during every survey ( $100 \%$ ) of the reef ball in concrete material type. Adult non-selected taxa were generally in low numbers and infrequently encountered on all material types, most obvious being the concrete block (1.4\%) where only one non-selected taxa was observed during one event.

ANOVA testing indicated that the selected taxa mean density was not significantly different for early juvenile, subadult or adult size classes between reef sites. The mean density of the selected taxa juvenile size class was significantly different $(\mathrm{p}=0.000)$ among three reef sites. Among juvenile selected taxa at the three reefs, mean density was significantly different among the Gerkin - Hart ( $\mathrm{p}=0.001$ ) and Hart - Walker ( $\mathrm{p}=0.000$ ) reefs.

The mean density of selected taxa was not significantly different among early juvenile, juvenile, subadult and adult selected taxa size classes among stations within each reef site. The mean density was not also significantly different among early juvenile, juvenile, subadult and adult selected taxa size classes among reef material types.

The mean density was not significantly different for the early juvenile size class of selected taxa among reef sites for reef balls. However, the mean density of juvenile, subadult and adult selected taxa was significantly different ( $\mathrm{p}=0.020,0.019,0.000$, respectively) on reef balls located at the three reef sites.

The log transformation of the density data appeared to decrease the likelihood of observing statistically significant differences among the data sets. This resulted in order of magnitude differences in untransformed data not being significantly different.

### 4.0 DISCUSSION

## GIS Habitat Mapping Effort

The habitat mapping and GIS database effort characterized the Sarasota Bay National Estuary Program shoreline in terms of geomorphology, shoreline morphology and intertidal vegetation tiers, and also refined the habitat map of the SBNEP study area. A total length of nearly 410 miles of shoreline was determined for the study area. The dominant shoreline type for the geomorphology tier was altered, which is consistent with the shore morphology tier, which showed bulkheads as the dominant shore morphology type, and the vegetative tier that indicated $51 \%$ of the shoreline was deemed to have no vegetation. Twenty-five percent $(25 \%)$ of the shoreline length was characterized as having a deep wetland fringe and this vegetative fringe was dominated by mangrove vegetation at $42 \%$.

Refinement of the habitat map for the SBNEP study area revealed that open water features, (i.e., bays and estuaries, were the dominant feature class at over $91 \%$ coverage. Of the tidal wetland features mapped, mangrove swamps at $6.7 \%$ was the next most common feature observed. The tidal features class and shoreline tier class followed the trends of development within the SBNEP study area.

During February and June 2002, the SBNEP fishery sampling effort in Sarasota Bay constituted 160 fixed-station seine samples in which 315,208 animals were collected. These samples ranged between 12 to 93 samples collected using one of five gear deployment techniques at 40 stations located in natural or restored habitats. The total number of animals collected for each gear deployment technique ranged from 465 to 196,333 animals.

Interpreting the community structure of the fauna collected in Sarasota Bay during 2002 is complicated by the finding that species composition and numerical abundance varied in relation to the type of gear deployment technique used at fishery sampling stations. Differences in sample size (number of samples) or sample area (area swept by net) between various gear deployment techniques yielded non-uniform results for characterizing species occurrence and relative abundance. Simply stated, the Sarasota Bay fishery community was not equally accessible to all gear deployment techniques used in this study. The study design required assessment of a variety of habitat types over a large spatial area; consequently, limitations due to selected gear types were hard to overcome. In anticipation of this result, the study design incorporated additional measures or analyses of community structure, beyond merely characterizing relative abundance and species occurrence, which could help to overcome the bias generated by selected gear types. Equally challenging was the effort to detect significant differences in a fishery community exhibiting high dominance.

## Comparison with Nearby Systems

The fishery sampling methodology employed in Sarasota Bay in 2002 was based on the protocols established by the Florida Marine Research Institute (FMRI) in several estuaries throughout Florida. Therefore, it is instructive to compare historical catch statistics for the estuarine regions adjacent to Sarasota Bay wherein FIM-FMRI monitoring has been conducted. Through 2001, FMRI has completed thirteen years of fishery independent monitoring (FIM) in several of Florida's estuarine systems using a variety of different gears and deployment techniques. For the present study, a review of FIM-FMRI catch statistics was made for the most recent three-year period with published data (i.e., 1999, 2000 \& 2001). Catch and effort data for $21.3-\mathrm{m}$ seine collections in Tampa Bay and Charlotte Harbor are reported in FIM-FMRI annual data summary reports (FMRI, 1999; FMRI, 2000; FMRI, 2001).

During the three year period 1999 through 2001 of FIM-FMRI monitoring, Tampa Bay annual catch and effort varied from 404,637 to 429,891 animals collected in 564 seine hauls
each year using a $21.3-\mathrm{m}$ seine in shore-fringing habitats. In Charlotte Harbor, annual catch and effort varied from 106,974 to 393,924 animals collected in 384 seine hauls each year using a $21.3-\mathrm{m}$ seine in shore-fringing habitats. For Sarasota Bay during 2002, a total catch of 315,208 animals was collected with less than half the effort (i.e., just 160 samples) than in adjacent estuaries and using equal or smaller sample areas over all gear deployment techniques in shore-fringing habitats. By comparison, during a 1990 SBNEP study, Mote Marine Laboratory collected 32,180 animals in $12015.2-\mathrm{m}$ seine samples but used a net mesh of $6.4-\mathrm{mm}$ (Edwards and Heuter, 1992)--twice the size of the $3.2-\mathrm{mm}$ net mesh used in FIM-FMRI monitoring and the present study.

The species richness (i.e., total number of taxa collected) for all $21.3-\mathrm{m}$ seine hauls during 1999 through 2001 FIM-FMRI monitoring varied from 110 to 119 taxa in Tampa Bay, and from 100 to 121 taxa in Charlotte Harbor. For Sarasota Bay during 2002, the total catch over all gear deployment techniques included 80 taxa in two sample seasons, winter and summer. By comparison, during a 1990 SBNEP study, Mote Marine Laboratory collected 74 taxa in bimonthly seine samples at 20 stations. Bird (1980) collected 119 taxa in bimonthly samples during 1979 at 10 stations in Sarasota Bay using a variety of gears and sampling techniques. The fewer taxa reported for Sarasota Bay than adjacent Tampa Bay and Charlotte Harbor in recent SBNEP-sponsored studies may be attributed to gear type limitations (e.g., different net mesh size in 1990 SBNEP study, or smaller sample areas) and seasonality limitations (e.g., only two seasons sampled in 2002 SBNEP study).

The seine samples collected in Sarasota Bay during 2002 were dominated by bait and forage fishes such as Anchoa mitchilli, Lucania parva, Gambusia holbrooki, Menidia spp., Harengula jaguana, Poecilia latipinna, Lagodon rhomboides, Eucinostomus spp., and Cyprinodon variegatus. This same pattern of dominance by a few taxa is exhibited for Tampa Bay and Charlotte Harbor during prior years. Edwards and Heuter (1992) indicated that this species abundance pattern appears "normal" for Sarasota Bay and adjacent estuaries.

Selected taxa were common in Sarasota Bay samples during 2002 when compared to dominant taxa and were typically among the top ten most abundant taxa in samples taken with various gear deployment techniques. A total of 31,088 animals (19 taxa) designated as selected taxa were collected in Sarasota Bay during 2002. Over all gear deployment techniques, selected taxa comprised from 1.5 to $26.6 \%$ of the total catch from Sarasota Bay during 2002. Leiostomus xanthurus and Mugil spp. (likely Mugil cephalus) typically comprised from 7 to $25 \%$ of the total catch from Sarasota Bay during 2002. In addition, Farfantepenaeus duorarum and Callinectes sapidus exhibited lesser abundance but were frequently collected.

A similar pattern of dominance among selected taxa collected with various gear deployment techniques was exhibited for Tampa Bay and Charlotte Harbor during prior years. For example, species richness varied from 13 to 22 taxa designated as Selected Taxa for Tampa Bay and Charlotte Harbor during 1999, 2000 and 2001 in collections with a $21.3-\mathrm{m}$ seine. However, annual variations occurred in the rank of specific taxa comprising the numerically dominant selected taxa over all gear deployment techniques. For example, either Farfantepenaeus duorarum or Leiostomus xanthurus were numerically dominant among selected taxa in Tampa Bay in 1999 and 2000; but this pattern shifted to Leiostomus xanthurus and Mugil cephalus as numerically dominant during 2001. Similarly, Farfantepenaeus duorarum and Cynoscion arenarius were numerically dominant among selected taxa in Charlotte Harbor in 1999 and 2000; but this pattern shifted to Leiostomus xanthurus and Farfantepenaeus duorarum as numerically dominant during 2001. A noTable difference in dominance by selected taxa between Sarasota Bay and adjacent estuaries is the lack of occurrence of Elops saurus or Cynoscion arenarius in Sarasota Bay samples during 2002. Edwards and Heuter (1992) reported a similar result of few or none Elops saurus and Cynoscion arenarius in seine samples from Sarasota Bay in 1990. These taxa typically occurred in the top five most abundant selected taxa in Tampa Bay and Charlotte Harbor during prior years; although the greatest frequencies of occurrence were related to collections made with larger gears (e.g., 183-m haul or purse seines). This picture of differences in selected taxa dominance between Sarasota Bay and adjacent estuaries may
be attributed to limitations due to gear deployment techniques that reflect differences in microhabitats sampled between estuaries.

A comparison reported by Edwards and Heuter (1992) related to differences in abundance of juvenile groupers between 1979 and 1990 in Sarasota Bay appears to hold for the present study. Bird (1980) collected 171 juvenile gag (Mycteroperca microlepis) and 37 juvenile red grouper (Epinephelus morio) from Sarasota Bay in 1979. Edwards and Heuter (1992) collected only 5 juvenile gag and no juvenile red grouper from Sarasota Bay in 1990. During the present study in Sarasota Bay, just 10 juvenile gag and no juvenile red grouper were collected in 2002. Compared to Sarasota Bay, FIM-FMRI findings suggest that juvenile gag exhibits similar abundance in Tampa Bay but is 2 to 5 times more abundant in Charlotte Harbor than in Sarasota Bay. Similarly, juvenile red grouper is rarely, if ever, collected by FIM-FMRI monitoring in Tampa Bay or Charlotte Harbor.

The catch per unit area (mean density) was also compared between nearby bay systems for 2001 with this study's results. The overall mean density of taxa collected from the bay seine in Tampa Bay and Charlotte Harbor was much lower than that observed for Sarasota Bay. The overall mean density for all species was $2,223.3 / 100 \mathrm{~m}^{2}$ for the Sarasota Bay area, compared to $239.6 / 100 \mathrm{~m}^{2}$ for Tampa Bay and $318.1 / 100 \mathrm{~m}^{2}$ for Charlotte Harbor. The 10 dominant species for Sarasota Bay all were observed at a greater mean density than the top 10 dominant species within Tampa Bay, while the top 4 dominant species within Sarasota Bay exceeded the mean density of the dominant species in Charlotte Harbor. The greater densities of species within Sarasota Bay were most noticeable for baitfish species (Anchoa mitchilli, Menidia spp. and Harengula jaguana) and for several of the killifish species. Lagodon rhomboides was the only Sarasota Bay dominant species observed within the range between Tampa Bay and Charlotte Harbor (54.9/100 $\mathrm{m}^{2}$ for Tampa Bay, 94.6/100 $\mathrm{m}^{2}$ for Sarasota Bay and 160.6/100 $\mathrm{m}^{2}$ for Charlotte Harbor).

Similar trends of mean density also held true for the selected taxa. The overall density of selected taxa within Sarasota Bay ( $220.8 / 100 \mathrm{~m}^{2}$ ) was nearly five times that of Tampa Bay
(48.3/100 $\mathrm{m}^{2}$ ) and eight times that of Charlotte Harbor $\left(29.5 / 100 \mathrm{~m}^{2}\right)$. This difference in density was primarily due to the high mean density observed for Leiostomus xanthurus (163.7/100 $\mathrm{m}^{2}$ ) within the Sarasota Bay, as compared to $42.2 / 100 \mathrm{~m}^{2}$ for Tampa Bay and $15.3 / 100 \mathrm{~m}^{2}$ for Charlotte Harbor. The combined mean density for all species of mullet (Mugil spp.) is also substantially higher in Sarasota Bay ( $52.0 / 100 \mathrm{~m}^{2}$ ) when compared to Tampa Bay $\left(0.2 / 100 \mathrm{~m}^{2}\right)$ and Charlotte Harbor $\left(1.8 / 100 \mathrm{~m}^{2}\right)$. The mean density of the pink shrimp (Farfantepenaeus duorarum) in Sarasota Bay (3.3/100 m${ }^{2}$ ) however, was below that of Tampa Bay (3.4/100 $\mathrm{m}^{2}$ ) and Charlotte Harbor (8.8/100 m ${ }^{2}$ ).

When interpreting the results of the mean density data it is important to remember that the mean density calculated for Tampa Bay and Charlotte Harbor represents the density across all 12 months of sampling, whereas the Sarasota Bay data represents sampling conducted in February and June only. February was the month with the highest overall abundance for taxa collected in Tampa Bay and the second highest month of total abundance for Charlotte Harbor, whereas June for both locations falls at the lower end of the abundance range.

## Comparison with Northern Indian River Lagoon

The mean density of taxa for the 21.3-meter bay seine for the northern Indian River for 2001 was also compared with the data collected for Sarasota Bay. The total mean density of all species collected for Sarasota Bay $(2,223.3 / 100 \mathrm{~m} 2)$ was about three times higher than the mean density for the northern Indian River (714.7/100 m2). Anchoa mitchilli was the dominant taxa within both bay systems, but the species mean density in Sarasota Bay (714.6/100 m2) was greater than the mean density within northern Indian River (449.2/100 $\mathrm{m} 2)$. Similar large differences in mean density were observed for the other dominant species within the Sarasota Bay system as compared to the northern Indian River. Interestingly, the species assemblage of the top 10 dominant taxa for the northern Indian River varies substantially from that of Sarasota Bay, with only five species being dominant in both bay systems.

The selected taxa also reflected greater mean density within the Sarasota Bay system ( $220.8 / 100 \mathrm{~m} 2$ ) when compared to the northern Indian River $(48.9 / 100 \mathrm{~m} 2)$. The mean density of Leiostomus xanthurus is significantly different between the two bay systems, with a mean density of $163.7 / 100 \mathrm{~m} 2$ observed in Sarasota Bay compared to a density of 1.7/100 m 2 observed within the northern Indian River. Conversely, the overall mean density of mullet was relatively similar, with $37.6 / 100 \mathrm{~m} 2$ in the northern Indian River and 52.0/100 m 2 observed within Sarasota Bay. The mean densities of Farfantepenaeus duorarum were similar between the two bay systems. It appears that the northern Indian River, located on the east coast of Florida, is a different system than Sarasota Bay, Tampa Bay and Charlotte Harbor.

## Comparison within Sarasota Bay

The observed mean density of the taxa collected varied widely between the various sampling stations. The highest overall density was observed for a deep water Bowlees Creek station $\left(23,801 / 100 \mathrm{~m}^{2}\right)$. The stations with the lowest observed density generally appear to occur in the small shallow creek systems with a strong freshwater influence, such as Philippi Creek and Hog Creek. In general, the highest species diversity was observed at the restored stations. This may be due to the presence several microhabitats and adjacent habitat types that occur in close proximity at these sampling stations, whereas the natural site sampled are generally more uniformed in nature with fewer habitat types adjacent. Average species diversity does not appear to follow a consistent trend, with an equal number of restored and natural habitats scoring high and low.

Selected taxa represented a small amount of the catch. A total of 19 species of selected taxa were observed out of the 80 species collected. The number of selected taxa observed at each site varied widely (from zero to eight species) with no consistent trend being observed between restored and natural sites. Although Leiostomus xanthurus and Mugil spp. were the most frequently encountered selected taxa, they were not equally represented at all stations. Their density also varied widely and likely reflects the various microhabitats present at those locations as well as physiographic conditions.

Pooling the selected taxa for both natural and restored sites reveals that the two types of sites are very similar in terms of selected taxa composition. The mean density of Leiostomus xanthurus was slightly higher for the restored sites ( $170.1 / 100 \mathrm{~m}^{2}$ ) when compared to the natural sites $\left(158.5 / 100 \mathrm{~m}^{2}\right)$. The overall mean density for the restored sites is slightly higher, at $230.7 / 100 \mathrm{~m}^{2}$, than the natural sites $\left(212.2 / 100 \mathrm{~m}^{2}\right)$.

Comparing the mean density of taxa by guild reveals that the natural sites have substantially higher mean density than that of the restored sites. Natural sites had higher mean density for both resident and transient species, however the restored sites had a slightly higher mean density of nursery species. The primary difference among guilds results from the high numbers of transient species found at a few of the natural sites swaying the overall density numbers toward that of the natural sites.

Not unexpectedly, the transient species guild shows the widest range in mean density between all sites varying widely between each microhabitat type and alternating back and forth between restored and natural habitat types. This reflects the transient nature and schooling behavior of these species, which often occur in high numbers over small areas. The natural sites did, however, have higher total species richness than that of the restored sites. Although the total mean densities fluctuated back and forth by microhabitat type for the resident species, the natural microhabitat types were generally higher in total species richness. Again, this could reflect the age and/or maturity of the restored system.

To summarize, it appears that the restored microhabitat types are faring fairly well compared to that of the natural habitat types. Other than the density of transient species, the species composition and density of the restored sites is comparable to and in some cases better than that of the natural sites. However, the natural sites do appear to generally be more diverse and there may be a lag affect related to species richness.

The ability to capture some taxa may have been limited by the use of the FMRI FIM methodology with it standard depth of deployment. The depth requirement, as well as the inability to get underneath overhanging vegetation, may have limited the ability to catch some species. Refinement of this methodology or adoption of an alternate methodology to sample in these habitat areas could provide information on the use of other microhabitat types by the selected taxa.

Assessment surveys of the three artificial reef sites also occurred during February and June 2002. A total of 59,027 individuals were observed within 27 species or species groups during the 56 sampling events. Early-juveniles of Clupeidae spp., Haemulon aurolineatum, larval fish and Lagodon rhomboides had noticeable seasonal variance between the February and June events. The Haemulon aurolineatum and Lagodon rhomboides early-juveniles were dominant during the February sampling event whereas the Clupeidae spp. was present only in the June event. The observed density for Lagodon rhomboides during the February event for early-juveniles may have been enhanced by the presence of large accumulations of macroalgae adjacent to the reef structures. This species was also observed in high numbers at many of the fishery sampling stations as noted earlier, whereas Haemulon aurolineatum was not observed in any of the fishery sampling stations. These early-juvenile stages were the dominant taxa, representing over $97 \%$ of the total individuals observed.

Overall, the monitoring of the reef sites reveals a distinct species assemblage from that of other habitat types within the Bay system. The diversity of species tended to increase with the variety of material types and with the increasing number of replicate samples collected. Many reef-dependent species such as Haemulon aurolineatum, Lutjanus griseus and Mycteroperca microlepsis were observed on the artificial reef material in relatively high numbers. Some species that are present in other habitat types within the Bay system were also found in generally higher numbers on the artificial reef material, including the commercially important stone crab (Menippe mercenaria).

Juveniles of Mycteroperca microlepsis were only observed during the June surveys and had a very high affinity for positioning themselves within the center of reef ball units. Subadult and adult Mycteroperca microlepsis preferred the large open spaces provided by the reef balls for hiding and were only found on structures with this type of material. In contrast, free ranging species, such as Chaetodipterus faber and Archosargus probatocephalus, were observed swimming over most of the material types. The selected taxa Lutjanus griseus was present in many life stages across almost all of the reef material types, but the subadults and juveniles of this species tended to prefer the interior of the reef ball structures whereas the larger adults would swim more openly away from and above the material. Menippe mercenaria were observed across many of the reef types, being noticeably absent from the concrete block type reefs, and appeared to prefer the interface between the hard reef structure and the soft substrate where they could burrow.

When comparing the reef balls at different sites, it appears that the time since deployment may be a factor in determining species composition. The newly deployed reef balls at the Walker reef site generally had low diversity and were heavily utilized by juvenile and subadult size classes. The new reef balls at the Walker sites were also heavily colonized by Menippe mercenaria of all size classes. In contrast, the older reef ball units on the Walker and Hart reef sites generally contained adult Mycteroperca microlepsis and adult Lutjanus griseus, but fewer of the juvenile and subadult size classes. It was also noted that the presence of an adult Mycteroperca microlepsis tended to limit the number and type of smaller fish present on the reef. This may reflect the predation rate of smaller individuals by adults of this species.

The complexity of the reef site influenced the type and number of species present. The lower complexity sites, such as concrete block, generally did not provide the number of protected spaces needed for some of the species to successfully habitate. This material type generally had low numbers and low species diversity. However, increasing the complexity of a concrete block pile with the presence of reef balls greatly improved the diversity and numbers of fish species present. The reef ball and concrete block pile habitat type supported
adults of many species including Mycteroperca microlepsis, but again had a noticeable lower abundance of early-juvenile and juvenile species. This may reflect predation influences or an adaptation of the early-juveniles and juveniles to avoid these large predatory species.

The reef ball units provided a habitat for a wide variety of different species, having the highest species richness for all material types. This material type was observed to have all four-size classes, from the early-juvenile stage of some up to the largest adult stage, of the Mycteroperca microlepsis. Again, Menippe mercenaria heavily utilized the reef balls both along the substrate and internally. Although this material type has the complexity of two, does provide good cover and structure for a variety of species.

The PVC reefs had different results depending on the type of unit constructed. The PVC web material generally had lower diversity and limited suitability for a variety of species. Mycteroperca microlepsis subadults and adults were noticeably absent from both of the PVC material types. Lutjanus griseus, on the other hand, were very abundant on this material type, especially the PVC ribcage design with several adults being observed. The free-swimming species, Archosargus probatocephalus and Chaetodipterus faber, also showed high utilization of the PVC ribcage design. This PVC material type is ideal for sites such as the Gerkin site where there is a soft muck layer in which typical, heavy concrete material would rapidly sink.

### 5.0 RECOMMENDATIONS

Development of detailed recommendations on restoration of juvenile fisheries habitat is difficult due to inherent limitations of the sampling gear and methodology. The observed species composition and abundance varied by gear deployment technique and gear type. The gear type and deployment technique were dictated by the conditions associated with the habitat areas to be sampled and so the fisheries community was not equally accessible to all gear deployment techniques. Deployment techniques also limited the ability to sample
"close-in" to specific habitat types and overhanging vegetation. The effort to detect differences was further exacerbated by a fish community that exhibited high dominance.

Despite these limitations, useful recommendations to guide juvenile fisheries habitat restoration can be made. The GIS shoreline mapping and tidal wetland features mapping are instructive in developing recommendations. Bulkheads ( 185.8 miles) and riprap ( 38.5 miles) represent $45.3 \%$ and $9.4 \%$, respectively, of the total shoreline length within the SBNEP study area. Most of the 207.3 miles (50.5\%) of the unvegetated shoreline occurs along these altered shorelines.

## Shoreline Enhancement

While these distances of altered shoreline are sobering as to the changes wrought to the estuary, they also represent a vast area in which restoration may occur. Due to its sloping nature and ability to provide some structural habitat, the riprap areas are of less concern/opportunity. However, tremendous opportunity exists to enhance the bulkhead shorelines through a seawall enhancement program. Seawall enhancement can be done with a variety of structures and materials, with no one type being necessarily preferred or recommended for all situations. Development of different enhancement units and techniques should continue.

Critical to large-scale enhancement of seawalls will be the streamlining of the permitting process. Neither the state nor federal government permitting agencies presently have either exemptions or general permits for the deployment of seawall enhancement materials. The expertise, time, and expense currently needed to obtain the required permits discourage the widespread use of seawall enhancements in this region. A combined effort of the local governments, NEP's and other agencies is needed to encourage the permitting agencies to streamline their processes.

## Exotics Removal

The GIS shoreline mapping also reveals that 7.6 miles ( $1.8 \%$ ) of the shoreline is dominated by nuisance or exotic vegetation (Australian pine, Brazilian pepper, and cattail). These vegetation types represent only a small proportion of the total shoreline, but provide relatively easy enhancement opportunities. These shoreline types occur in many bay segments, but are concentrated in the southern portion (Sections 13, 14, and 16) of the study area, including Philippi Creek.

## Habitat Diversity

The shoreline mapping also reveals very low amounts of Juncus ( 0.6 miles, $0.1 \%$ ), leatherfern ( 1.5 miles, $0.4 \%$ ) and Spartina ( 2.3 miles, $0.5 \%$ ) along the shorelines of the SBNEP study area. Although sparsely spread through many of the bay segments, these three shoreline types are concentrated in Sections 13, 14, and 16 in the southern portion of the study area. Philippi Creek, located in segment 13, contains a large proportion of these habitat types, as well as the nuisance and exotic habitat types mentioned earlier. The low proportion of these habitat types would indicate that these habitat types should be targeted for creation. However, one potential reason for the low percentage of the shoreline represented by these habitat types is that they represent an early successional habitat stage. These habitat types often proceed in development to a mangrove community, which is the climax community in our area.

Due to the limitations in the sampling program, as noted above, definitive direction as to the types of habitats to be restored cannot be provided. However, because on the apparent success of the restored habitats for several juvenile species, it is recommended that habitat restoration continue. Based on the variable nature of where the selected species were captured, it appears prudent to continue to create restoration sites with a mosaic of the various microhabitat types sampled.

As noted, the water quality data reveal limited areas of oligohaline and mesohaline water quality conditions. Within these areas, native habitats are poorly represented. These areas
are well documented to support many selected fisheries species that aggregate during portions of their life history within these salinity regimes. As such, it is recommended that these salinity regimes be targeted for habitat restoration and creation in the future.

## Submerged Habitats

The tidal wetlands shoreline mapping reveals $91.1 \%$ of the SBNEP acreage is contained within the bays and estuaries classification. Although this classification contains seagrass habitat, large areas of unvegetated bottom and open water occur within the study area. As for the altered shorelines noted above, these open areas of bay represent large areas in which enhancement could potentially occur. One opportunity for enhancement would be the spread of seagrass into deeper areas through water quality improvement. Another enhancement opportunity involves the permitting and placement of additional artificial reefs within the deeper portions of the bays and estuaries.

Of the emergent habitat types, salt marshes with 68.2 acres ( $0.2 \%$ ) and oyster bars with 83.5 acres $(0.2 \%)$ are only marginally represented within SBNEP study area. In addition, 552.3 acres $(1.6 \%)$ of the study area is unvegetated bottom. Although the unvegetated bottom areas are likely unvegetated for reasons of wave action or unsuitable depths (too shallow), slight increases in the depth of these areas could transform them into seagrass meadows, while a decrease in the depth could allow salt marsh development. Placement of oyster shell in some of these locations could also lead to oyster bar development.

Artificial reef materials in the SBNEP study area provide a unique habitat type for a variety of reef-dependent fish species, as well as the commercially important Menippe mercenaria. The ability of artificial reef material to enhance the species composition and abundance on otherwise poorly inhabited bottoms, such as dredge holes and open sand areas, greatly increases the ability to improve the fishery of some of these species. In general, the more complex the reef design and the more material types that can be used, the better the reef will be for a variety of species types. Enhancement of existing artificial reefs is possible and recommended through the placement of additional material, and stacking or placing of new
material types around existing material (i.e., placement of concrete blocks around existing reef balls).

The use of prefabricated PVC reef unit designs is recommended within those dredge hole areas where typical artificial reef material would rapidly settle. Again, a variety of design and shapes should be used to increase and improve the overall efficiency of these sites. However, due to the light weight of this material type, it is recommended that the PVC reefs be limited to large dredge type holes where the water current are low and physical disturbances would be limited.

The reef ball units continue to make an excellent artificial reef material type. Future deployment should consider enhancement to existing reef ball units to improve complexity, however. It may be advantageous to create reef ball designs that reduce or eliminate the large open spaces in the interior of the reef ball that could be utilized by adult Mycteroperca microlepsis. This would allow the smaller juvenile and early-juvenile species to successfully hide within the reef ball units without fear of predation.

## Future Work

The Habitat Restoration Working Group (HRWG) of the SBNEP determined that a five-year restoration plan document should be prepared. It was felt that this document could be used to develop and improve restoration targets and may assist local agencies and organizations in obtaining additional funds for restoration projects. A similar guiding document has been prepared for other areas such as Tampa and Chesapeake Bays. As a precursor to the development of this plan, further research is recommended to fully evaluate habitat availability in Sarasota Bay. Members of the Sarasota Bay technical community should review existing research in specific areas (including essential fish habitat, wetland restoration, artificial reefs, and seawall enhancement) and provide written habitat restoration recommendations. These recommendations should be summarized in a report and presented at a workshop to the full SBNEP TAC for comment.

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## Appendix A












Figure A-10. Percentage of bays and estuaries within the Sarasota Bay study area by bay segment.


Figure A-11. Percentage of mangrove swamps within the Sarasota Bay study area by bay segment.


Figure A-12. Percentage of oyster bars within the Sarasota Bay study area by bay segment.


Figure A-13. Percentage of saltwater marshes within the Sarasota Bay study area by bay segment.


Figure A-14. Percentage of salt barren within the Sarasota Bay study area by bay segment.


Figure A-15. Percentage of unvegetated bottoms within the Sarasota Bay study area by bay segment.


Figure A-16. Percentage of beaches other than swimming within the Sarasota Bay study area by bay segment.

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## Appendix B

Table B-1. Mean density ( $\# / 100 \mathrm{~m}^{2}$ ) and standard error ( $\pm$ SE) by taxa at each sampling station for all sampling events.

|  |  | Tidy Island | Tidy Island | Tidy Island | Leffis Key | Sister Keys |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Sister Keys |  |  |  |  |  |  |
| Site 3 |  |  |  |  |  |  |

Table B-1. (Continued).

| Species name | Common name | Tidy Island Site C 1 <br> Mean $\pm$ SE | Tidy Island <br> Site C 2 <br> Mean $\pm$ SE | Tidy Island Site C 3 Mean $\pm$ SE | Leffis Key Site C 4 Mean $\pm$ SE | Sister Keys <br> Site C 5 <br> Mean $\pm$ SE | Sister Keys <br> Site C 6 <br> Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Oligoplites saurus | Leatherjacket | $2.0 \pm 1.7$ | $0.2 \pm 0.2$ | $0.4 \pm 0.3$ | $2.5 \pm 2.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $25 \pm 25$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $1.4 \pm 1.4$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Synodus foetens | Inshore lizardfish | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $1.1 \pm 0.9$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $1.3 \pm 0.7$ | $3.2 \pm 3.2$ | $2.1 \pm 2.1$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $7.1 \pm 7.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timucu | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ |
| Chilomycterus schoepfi | Striped burrfish | $0.0 \pm 0.0$ | $4.1 \pm 4.1$ | $0.0 \pm 0.0$ | $0.9 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma robustum | Code goby | $0.0 \pm 0.0$ | $2.9 \pm 2.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $2.9 \pm 2.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diplodus holbrooki | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.9 \pm 2.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $3.7 \pm 3.7$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $2.9 \pm 2.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $0.5 \pm 0.5$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.8 \pm 1.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | Tidy Island Site C 1 Mean $\pm$ SE | Tidy Island Site C 2 <br> Mean $\pm$ SE | Tidy Island Site C 3 Mean $\pm$ SE | Leffis Key <br> Site C 4 <br> Mean $\pm$ SE | Sister Keys Site C 5 <br> Mean $\pm$ SE | Sister Keys <br> Site C 6 <br> Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $1319.3 \pm 379.1$ | $1754.6 \pm 731.7$ | $1046.4 \pm 372.3$ | $5286.6 \pm 3608.9$ | $1350 \pm 739.7$ | $1086.8 \pm 314.8$ |
| Richness |  | $11.0 \pm 1.3$ | $10.3 \pm 1.7$ | $9.8 \pm 2.6$ | $13.8 \pm 1.4$ | $3 \pm 1.3$ | $9.3 \pm 1$ |
| Diversity |  | $1.4 \pm 0.2$ | $1.1 \pm 0.2$ | $1.1 \pm 0.2$ | $1.0 \pm 0.3$ | $0.5 \pm 0.2$ | $1.4 \pm 0.1$ |

Table B-1. (Continued).

| Species name | Common name | Sister Keys <br> Site C 7 <br> Mean $\pm$ SE | El Conquistador Site C 8 <br> Mean $\pm$ SE | El Conquistador Site C 9 <br> Mean $\pm$ SE | Bowlees Creek Site C 10 <br> Mean $\pm$ SE | Bowlees Creek <br> Site C 11 <br> Mean $\pm$ SE | Bowlees Creek Site C 12 Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchoa mitchilli | Bay anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $14774.3 \pm 7441.3$ | $400 \pm 400$ |
| Lucania parva | Rainwater killifish | $2935.2 \pm 1659.6$ | $2.2 \pm 2.2$ | $323.0 \pm 147.4$ | $28.6 \pm 26.9$ | $11.8 \pm 11.8$ | $16.1 \pm 6.8$ |
| Gambusia holbrooki | Eastern mosquito fish | $0.0 \pm 0.0$ | $92.4 \pm 63.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Leiostomus xanthurus | Spot | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $82.1 \pm 82.1$ | $2.0 \pm 1.1$ | $2082.4 \pm 2066.7$ | $74.8 \pm 74.2$ |
| Menidia spp. | Assorted silverside | $13.0 \pm 12.6$ | $1.1 \pm 1.1$ | $102.5 \pm 40.4$ | $43.7 \pm 29.5$ | $150 \pm 93.8$ | $351.6 \pm 332.4$ |
| Harengula jaguana | Scaled sardine | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5058.8 \pm 5027.5$ | $0.0 \pm 0.0$ |
| Poecilia latipinna | Sailfin molly | $0.0 \pm 0.0$ | $127.2 \pm 51.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lagodon rhomboides | Pinfish | $209.3 \pm 109.1$ | $5.4 \pm 4.1$ | $47.1 \pm 27.7$ | $39.3 \pm 36.7$ | $202.9 \pm 183.9$ | $69.1 \pm 13.3$ |
| Eucinostomus spp. | Assorted mojarra | $1.1 \pm 1.1$ | $0.0 \pm 0.0$ | $732.9 \pm 582.3$ | $100.9 \pm 63.1$ | $121 \pm 88.7$ | $11.1 \pm 6.9$ |
| Cyprinodon variegatus | Sheepshead minnow | $3 \pm 3$ | $1.1 \pm 1.1$ | $108.3 \pm 100.8$ | $1.3 \pm 1.3$ | $117.6 \pm 89.1$ | $0.0 \pm 0.0$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $350 \pm 267.5$ | $0.0 \pm 0.0$ |
| Fundulus grandis | Gulf killifish | $48.9 \pm 43.8$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $94.1 \pm 94.1$ | $0.0 \pm 0.0$ |
| Mugil cephalus | Striped mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Anchoa cubana | Cuban anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $535.3 \pm 313.5$ | $0.0 \pm 0.0$ |
| Floridichthys carpio | Goldspotted killifish | $17.1 \pm 13.6$ | $1.1 \pm 1.1$ | $23.9 \pm 19.5$ | $2.0 \pm 1.7$ | $50 \pm 46.2$ | $4.1 \pm 2.5$ |
| Eucinostomus harengulus | Tidewater mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $1.3 \pm 0.6$ | $97.8 \pm 56.5$ | $0.5 \pm 0.5$ |
| Microgobius gulosus | Clown goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus majalis | Striped killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.9 \pm 5.9$ | $0.0 \pm 0.0$ |
| Clupeidae spp. | Assorted herring | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.6 \pm 1.6$ | $0.2 \pm 0.2$ | $2.9 \pm 2.9$ | $0.0 \pm 0.0$ |
| Eucinostomus gula | Silver jenny | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $17.3 \pm 17.8$ | $2.5 \pm 2.3$ | $0.0 \pm 0.0$ | $2.9 \pm 2.9$ |
| Anchoa hepsetus | Striped anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $8.4 \pm 8.4$ | $135.3 \pm 135.3$ | $0.0 \pm 0.0$ |
| Brevoortia spp. | Assorted menhaden | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $2.0 \pm 1.3$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $1.1 \pm 0.7$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Strongylura notata | Redfin needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.2 \pm 0.2$ | $2.9 \pm 2.9$ | $0.4 \pm 0.4$ |
| Syngnathus scovelli | Gulf pipefish | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.9 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphoeroides nephelus | Southern puffer | $0.5 \pm 0.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $4.5 \pm 3.8$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $20.5 \pm 20.5$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lophogobius cyprinoides | Crested goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Heterandria formosa | Least killifish | $0.0 \pm 0.0$ | $33.7 \pm 26.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | Sister Keys <br> Site C 7 <br> Mean $\pm$ SE | $\begin{gathered} \hline \hline \text { El Conquistador } \\ \text { Site C } 8 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | ```El Conquistador Site C 9 Mean + SE``` | $\begin{gathered} \hline \hline \text { Bowlees Creek } \\ \text { Site C } 10 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Bowlees Creek <br> Site C 11 <br> Mean $\pm$ SE | Bowlees Creek Site C 12 Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Oligoplites saurus | Leatherjacket | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.3 \pm 0.8$ | $0.7 \pm 0.5$ | $1.1 \pm 1.1$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $23.9 \pm 14.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Opsanus beta | Gulf toadfish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Synodus foetens | Inshore lizardfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $0.2 \pm 0.2$ | $2.0 \pm 1.3$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timиси | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $5.9 \pm 5.9$ | $0.0 \pm 0.0$ |
| Chilomycterus schoepfi | Striped burrfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $6.5 \pm 3.8$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma robustum | Code goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diplodus holbrooki | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | $\begin{gathered} \hline \hline \text { Sister Keys } \\ \text { Site C } 7 \\ \text { Mean } \pm \text { SE } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { El Conquistador } \\ \text { Site C } 8 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | $\begin{gathered} \hline \text { El Conquistador } \\ \text { Site C } 9 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Bowlees Creek <br> Site C 10 <br> Mean $\pm$ SE | Bowlees Creek <br> Site C 11 <br> Mean $\pm$ SE | Bowlees Creek <br> Site C 12 <br> Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ |
| Chaetodipterus faber | Atlantic spadefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $3232.3 \pm 1608.7$ | $296.7 \pm 150.3$ | $1462.2 \pm 540.2$ | $242.3 \pm 44.6$ | $23800.7 \pm 12321$ | $931.6 \pm 462.5$ |
| Richness |  | $5.8 \pm 0.3$ | $5.3 \pm 1.7$ | $6.8 \pm 1.8$ | $11.5 \pm 2.5$ | $8.8 \pm 2$ | $6.8 \pm 1.4$ |
| Diversity |  | $0.6 \pm 0.2$ | $1 \pm 0.3$ | $1.0 \pm 0.3$ | $0.9 \pm 0.2$ | $0.7 \pm 0.3$ | $0.9 \pm 0.3$ |

Table B-1. (Continued).

| Species name | Common name | $\begin{gathered} \hline \hline \text { S Lido Beach } \\ \text { Site C } 13 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | $\begin{gathered} \hline \hline \text { S Lido Lagoon } \\ \text { Site C } 14 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Phillippi Creek Site C 15 Mean $\pm$ SE | Phillippi Creek <br> Site C 16 <br> Mean $\pm$ SE | ```Heron Lagoon Site C 17 Mean }\pm\mathrm{ SE``` | Heron Lagoon Site C 18 Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchoa mitchilli | Bay anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $13.4 \pm 13.2$ | $0.0 \pm 0.0$ | $273.2 \pm 252.5$ | $47.1 \pm 47.1$ |
| Lucania parva | Rainwater killifish | $216.8 \pm 137$ | $2566.1 \pm 831.8$ | $3.2 \pm 3.2$ | $8.2 \pm 4.3$ | $0.0 \pm 0.0$ | $5867.6 \pm 1577.5$ |
| Gambusia holbrooki | Eastern mosquito fish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $3.2 \pm 3$ | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $2861.8 \pm 1928.1$ |
| Leiostomus xanthurus | Spot | $70.9 \pm 52.6$ | $4.3 \pm 4.3$ | $1.1 \pm 1.1$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ |
| Menidia spp. | Assorted silverside | $17.1 \pm 17.1$ | $40.9 \pm 34.3$ | $1351.6 \pm 771.9$ | $0.5 \pm 0.5$ | $11 \pm 5.6$ | $1000 \pm 540.9$ |
| Harengula jaguana | Scaled sardine | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Poecilia latipinna | Sailfin molly | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.2 \pm 3.2$ | $0.4 \pm 0.2$ | $0.4 \pm 0.4$ | $655.9 \pm 407.4$ |
| Lagodon rhomboides | Pinfish | $269.1 \pm 129.6$ | $97.5 \pm 62.5$ | $12.3 \pm 11.6$ | $3.9 \pm 0.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus spp. | Assorted mojarra | $4.5 \pm 2.9$ | $5.7 \pm 3.3$ | $4.4 \pm 4.4$ | $4.6 \pm 4.2$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Cyprinodon variegatus | Sheepshead minnow | $5.0 \pm 4.1$ | $7.7 \pm 5.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.9 \pm 2.9$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus grandis | Gulf killifish | $1.4 \pm 1.4$ | $4.3 \pm 4.3$ | $0.4 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil cephalus | Striped mullet | $1.4 \pm 1.4$ | $6.8 \pm 6.8$ | $1.1 \pm 1.1$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $5.9 \pm 5.9$ |
| Anchoa cubana | Cuban anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Floridichthys carpio | Goldspotted killifish | $1.4 \pm 1.4$ | $23.0 \pm 8.4$ | $2.3 \pm 2.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $35.3 \pm 28$ |
| Eucinostomus harengulus | Tidewater mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $32.7 \pm 21.4$ | $0.4 \pm 0.2$ | $5.9 \pm 3.8$ | $0.0 \pm 0.0$ |
| Microgobius gulosus | Clown goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $27.4 \pm 26.2$ | $6.1 \pm 2.8$ | $0.4 \pm 0.4$ | $126.5 \pm 79.4$ |
| Fundulus majalis | Striped killifish | $0.0 \pm 0.0$ | $4.3 \pm 4.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Clupeidae spp. | Assorted herring | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus gula | Silver jenny | $1.1 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Anchoa hepsetus | Striped anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Brevoortia spp. | Assorted menhaden | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura notata | Redfin needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $3.7 \pm 3.7$ | $0.0 \pm 0.0$ |
| Syngnathus scovelli | Gulf pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $8.5 \pm 8.5$ |
| Sphoeroides nephelus | Southern puffer | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lophogobius cyprinoides | Crested goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $25.0 \pm 19.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $3.7 \pm 2.8$ |
| Heterandria formosa | Least killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | $\begin{gathered} \hline \hline \text { S Lido Beach } \\ \text { Site C } 13 \\ \text { Mean } \pm \text { SE } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { S Lido Lagoon } \\ \text { Site C } 14 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Phillippi Creek <br> Site C 15 <br> Mean $\pm$ SE | Phillippi Creek <br> Site C 16 <br> Mean $\pm$ SE | Heron Lagoon Site C 17 Mean $\pm$ SE | Heron Lagoon Site C 18 Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Oligoplites saurus | Leatherjacket | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.5 \pm 0.3$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $11.8 \pm 6.8$ |
| Synodus foetens | Inshore lizardfish | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $0.2 \pm 0.2$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timucu | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chilomycterus schoepfi | Striped burrfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma robustum | Code goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ |
| Diplodus holbrooki | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | $\begin{gathered} \hline \hline \text { S Lido Beach } \\ \text { Site C } 13 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | $\begin{gathered} \hline \hline \text { S Lido Lagoon } \\ \text { Site C } 14 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Phillippi Creek <br> Site C 15 <br> Mean $\pm$ SE | Phillippi Creek <br> Site C 16 <br> Mean $\pm$ SE | $\begin{gathered} \hline \hline \text { Heron Lagoon } \\ \text { Site C } 17 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | $\begin{gathered} \hline \hline \text { Heron Lagoon } \\ \text { Site C } 18 \\ \text { Mean } \pm \text { SE } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ |
| Calamus arctifrons | Grass porgy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $591.8 \pm 299.6$ | $2760.7 \pm 863.1$ | $1484.7 \pm 828.9$ | $25.9 \pm 8.2$ | $296.3 \pm 25.6$ | $10628.3 \pm 2720.9$ |
| Richness |  | $6.5 \pm 1.8$ | $5.5 \pm 0.5$ | $8.3 \pm 0.6$ | $6.3 \pm 0.5$ | $3.5 \pm 0.6$ | $7.5 \pm 1.5$ |
| Diversity |  | $0.8 \pm 0.2$ | $0.4 \pm 0.8$ | $0.9 \pm 0.3$ | $1.4 \pm 0.1$ | $0.5 \pm 0.2$ | $0.9 \pm 0.1$ |

Table B-1. (Continued).

| Species name | Common name | N Catfish <br> Site C 19 <br> Mean $\pm$ SE | Bird Key <br> Site C 20 <br> Mean $\pm$ SE | $\begin{gathered} \hline \hline \text { South Creek } \\ \text { Site C } 21 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | $\begin{gathered} \hline \hline \text { South Creek } \\ \text { Site C } 22 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Leffis Key Site R 1 $\text { Mean } \pm \text { SE }$ | Leffis Key Site R 2 $\text { Mean } \pm \text { SE }$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchoa mitchilli | Bay anchovy | $7.3 \pm 7.1$ | $1.4 \pm 1.4$ | $217.1 \pm 217.1$ | $19.5 \pm 19.5$ | $2822.9 \pm 2822.9$ | $0.0 \pm 0.0$ |
| Lucania parva | Rainwater killifish | $0.4 \pm 0.2$ | $518.6 \pm 346.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $107.7 \pm 79.1$ | $116.7 \pm 116.7$ |
| Gambusia holbrooki | Eastern mosquito fish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Leiostomus xanthurus | Spot | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $289.8 \pm 238$ | $25 \pm 16$ |
| Menidia spp. | Assorted silverside | $1.3 \pm 1.3$ | $92.3 \pm 83.3$ | $5.5 \pm 4.0$ | $7.3 \pm 4.6$ | $81.8 \pm 39$ | $383.3 \pm 383.3$ |
| Harengula jaguana | Scaled sardine | $222.9 \pm 222.9$ | $14.3 \pm 14.3$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Poecilia latipinna | Sailfin molly | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $84.3 \pm 84.3$ | $58.3 \pm 58.3$ |
| Lagodon rhomboides | Pinfish | $0.5 \pm 0.3$ | $78.2 \pm 47.4$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $380.9 \pm 172.7$ | $1166.7 \pm 77.9$ |
| Eucinostomus spp. | Assorted mojarra | $8.9 \pm 8.9$ | $3.6 \pm 2.1$ | $5.0 \pm 5.0$ | $0.0 \pm 0.0$ | $120 \pm 35.5$ | $50 \pm 31.9$ |
| Cyprinodon variegatus | Sheepshead minnow | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $184.1 \pm 103.7$ | $8.3 \pm 8.3$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $267.3 \pm 154.7$ | $0.0 \pm 0.0$ |
| Fundulus grandis | Gulf killifish | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $17 \pm 12.8$ | $0.0 \pm 0.0$ |
| Mugil cephalus | Striped mullet | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Anchoa cubana | Cuban anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Floridichthys carpio | Goldspotted killifish | $0.0 \pm 0.0$ | $47.3 \pm 39.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus harengulus | Tidewater mojarra | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Microgobius gulosus | Clown goby | $3.0 \pm 1.2$ | $0.0 \pm 0.0$ | $25.4 \pm 7.1$ | $16.8 \pm 12.8$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Fundulus majalis | Striped killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.4 \pm 2.9$ | $0.0 \pm 0.0$ |
| Clupeidae spp. | Assorted herring | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $17.9 \pm 17.9$ | $0.0 \pm 0.0$ |
| Eucinostomus gula | Silver jenny | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $6.4 \pm 3.8$ | $0.0 \pm 0.0$ |
| Anchoa hepsetus | Striped anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Brevoortia spp. | Assorted menhaden | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.0 \pm 0.0$ | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $13.2 \pm 7.3$ | $33.3 \pm 19.2$ |
| Strongylura notata | Redfin needlefish | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ |
| Syngnathus scovelli | Gulf pipefish | $0.0 \pm 0.0$ | $1.1 \pm 0.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $16.7 \pm 9.6$ |
| Sphoeroides nephelus | Southern puffer | $0.2 \pm 0.2$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $3.6 \pm 2.7$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $8.3 \pm 8.3$ |
| Lophogobius cyprinoides | Crested goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.7 \pm 3.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Heterandria formosa | Least killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | N Catfish Site C 19 Mean $\pm$ SE | Bird Key <br> Site C 20 <br> Mean $\pm$ SE | $\begin{gathered} \hline \text { South Creek } \\ \text { Site C } 21 \\ \text { Mean } \pm \text { SE } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { South Creek } \\ \text { Site C } 22 \\ \text { Mean } \pm \text { SE } \\ \hline \end{gathered}$ | ```Leffis Key Site R 1 Mean }\pm\mathrm{ SE``` | $\begin{gathered} \hline \hline \text { Leffis Key } \\ \text { Site R } 2 \\ \text { Mean } \pm \text { SE } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Oligoplites saurus | Leatherjacket | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.8 \pm 1.4$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $23.6 \pm 23.6$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.4 \pm 0.8$ | $8.3 \pm 8.3$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Synodus foetens | Inshore lizardfish | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $8.3 \pm 8.3$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.7 \pm 2.0$ | $0.0 \pm 0.0$ |
| Gobiosoma spp. | Assorted goby | $0.4 \pm 0.2$ | $0.0 \pm 0.0$ | $2.9 \pm 1.3$ | $2.1 \pm 2.1$ | $3.9 \pm 2.7$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $8.3 \pm 8.3$ |
| Mugil curema | White mullet | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timucu | Timucu | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chilomycterus schoepfi | Striped burrfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma robustum | Code goby | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.9 \pm 0.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diplodus holbrooki | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.4 \pm 0.2$ | $0.9 \pm 0.9$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.1 \pm 1.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | $\begin{gathered} \hline \hline \text { N Catfish } \\ \text { Site C } 19 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | $\begin{gathered} \hline \hline \text { Bird Key } \\ \text { Site C } 20 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | $\begin{gathered} \hline \hline \text { South Creek } \\ \text { Site C } 21 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | South Creek <br> Site C 22 <br> Mean $\pm$ SE | Leffis Key <br> Site R 1 <br> Mean $\pm$ SE | Leffis Key <br> Site R 2 <br> Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.5 \pm 0.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.9 \pm 0.9$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $248 \pm 226$ | $759.1 \pm 408.4$ | $267.7 \pm 230.4$ | $47.5 \pm 18.7$ | $4441.2 \pm 2543.5$ | $1891.7 \pm 703.7$ |
| Richness |  | $6 \pm 1.9$ | $5.0 \pm 1.2$ | $6.3 \pm 0.9$ | $3.3 \pm 0.9$ | $14.8 \pm 0.8$ | $4.5 \pm 1.3$ |
| Diversity |  | $1 \pm 0.5$ | $0.4 \pm 0.2$ | $0.8 \pm 0.2$ | $0.4 \pm 0.1$ | $1.3 \pm 0.4$ | $0.5 \pm 0.1$ |

Table B-1. (Continued).

| Species name | Common name | Leffis Key Site R 3 $\text { Mean } \pm \text { SE }$ | Durante Park Site R 4 <br> Mean $\pm$ SE | Durante Park Site R 5 <br> Mean $\pm$ SE | $\begin{gathered} \hline \hline \text { Durante Park } \\ \text { Site R } 6 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Sarasota Baywalk Site R 7 <br> Mean $\pm$ SE | Sarasota Baywalk <br> Site R 8 <br> Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchoa mitchilli | Bay anchovy | $18.6 \pm 18.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.6 \pm 0.6$ |
| Lucania parva | Rainwater killifish | $3.7 \pm 3.1$ | $0.6 \pm 0.6$ | $258.3 \pm 216.2$ | $511.2 \pm 359.6$ | $1.9 \pm 1.9$ | $0.6 \pm 0.6$ |
| Gambusia holbrooki | Eastern mosquito fish | $0.0 \pm 0.0$ | $665.4 \pm 68.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Leiostomus xanthurus | Spot | $38.7 \pm 21.8$ | $0.0 \pm 0.0$ | $991.7 \pm 98.6$ | $256.4 \pm 253.6$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Menidia spp. | Assorted silverside | $155.0 \pm 96.7$ | $13.5 \pm 9.4$ | $741.7 \pm 443.5$ | $30.0 \pm 17.6$ | $45.5 \pm 20.9$ | $7.1 \pm 6.2$ |
| Harengula jaguana | Scaled sardine | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Poecilia latipinna | Sailfin molly | $8.9 \pm 8.9$ | $287.8 \pm 280.2$ | $25.0 \pm 25.0$ | $0.0 \pm 0.0$ | $1.9 \pm 1.2$ | $159.6 \pm 58.2$ |
| Lagodon rhomboides | Pinfish | $19.8 \pm 11.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $377.9 \pm 173.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus spp. | Assorted mojarra | $53.0 \pm 25.8$ | $0.0 \pm 0.0$ | $683.3 \pm 394.5$ | $144.3 \pm 109.4$ | $5.3 \pm 4.5$ | $0.0 \pm 0.0$ |
| Cyprinodon variegatus | Sheepshead minnow | $32.0 \pm 18.1$ | $86.5 \pm 35.6$ | $341.7 \pm 341.7$ | $0.0 \pm 0.0$ | $35.9 \pm 22.4$ | $259 \pm 169.7$ |
| Mugil spp. | Assorted mullet | $1.3 \pm 1.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus grandis | Gulf killifish | $3.2 \pm 1.7$ | $2.6 \pm 2.6$ | $166.7 \pm 166.7$ | $0.5 \pm 0.5$ | $4.7 \pm 2.9$ | $17.3 \pm 12.5$ |
| Mugil cephalus | Striped mullet | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $23.7 \pm 15.5$ | $25.6 \pm 24$ |
| Anchoa cubana | Cuban anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Floridichthys carpio | Goldspotted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.1 \pm 4.3$ | $5.8 \pm 5.8$ |
| Eucinostomus harengulus | Tidewater mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $25.0 \pm 25.0$ | $0.0 \pm 0.0$ | $51.6 \pm 46.1$ | $0.0 \pm 0.0$ |
| Microgobius gulosus | Clown goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $13.9 \pm 13.0$ | $0.0 \pm 0.0$ |
| Fundulus majalis | Striped killifish | $10.2 \pm 6.1$ | $0.6 \pm 0.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $12.3 \pm 4.1$ | $42.9 \pm 29.1$ |
| Clupeidae spp. | Assorted herring | $2.1 \pm 2.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus gula | Silver jenny | $1.1 \pm 0.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.6 \pm 1$ | $46.2 \pm 45.3$ | $0.0 \pm 0.0$ |
| Anchoa hepsetus | Striped anchovy | $0.9 \pm 0.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Brevoortia spp. | Assorted menhaden | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $4.5 \pm 2.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $30.2 \pm 29.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura notata | Redfin needlefish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus scovelli | Gulf pipefish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $3.4 \pm 2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphoeroides nephelus | Southern puffer | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.3 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $1.3 \pm 0.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Lophogobius cyprinoides | Crested goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Heterandria formosa | Least killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | Leffis Key Site R 3 $\text { Mean } \pm \text { SE }$ | $\begin{gathered} \text { Durante Park } \\ \text { Site R } 4 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Durante Park Site R 5 <br> Mean $\pm$ SE | $\begin{gathered} \text { Durante Park } \\ \text { Site R } 6 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Sarasota Baywa Site R 7 <br> Mean $\pm$ SE | Sarasota Baywalk <br> Site R 8 <br> Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Oligoplites saurus | Leatherjacket | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.6 \pm 0.6$ | $0.0 \pm 0.0$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Synodus foetens | Inshore lizardfish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.7 \pm 2.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.1 \pm 1.9$ | $2.6 \pm 2.6$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timucu | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chilomycterus schoepfi | Striped burrfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $4.5 \pm 2.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma robustum | Code goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diplodus holbrooki | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.1 \pm 1.1$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | $\begin{gathered} \hline \hline \text { Leffis Key } \\ \text { Site R 3 } \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Durante Park Site R 4 Mean $\pm$ SE | $\begin{gathered} \hline \hline \text { Durante Park } \\ \text { Site R } 5 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | $\begin{gathered} \hline \hline \text { Durante Park } \\ \text { Site R } 6 \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Sarasota Baywalk Site R 7 <br> Mean $\pm$ SE | $\begin{aligned} & \hline \hline \text { Sarasota Baywalk } \\ & \text { Site R 8 } \\ & \text { Mean } \pm \text { SE } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $356.4 \pm 77.1$ | $1061.5 \pm 576.7$ | $3233.3 \pm 939.6$ | $1362.7 \pm 270.5$ | $253.7 \pm 133.8$ | $521.2 \pm 253.3$ |
| Richness |  | $12.3 \pm 0.9$ | $4.3 \pm 0.3$ | $3.0 \pm 0.8$ | $9.0 \pm 1.8$ | $9.5 \pm 2.9$ | $5.3 \pm 1.4$ |
| Diversity |  | $1.5 \pm 0.2$ | $0.6 \pm 0.2$ | $0.7 \pm 0.3$ | $0.8 \pm 0.3$ | $1.4 \pm 0.5$ | $1.1 \pm 0.2$ |

Table B-1. (Continued).

| Species name | Common name | Sarasota Baywalk Site R 9 $\text { Mean } \pm \text { SE }$ | Sixth St Canal <br> Site R 10 <br> Mean $\pm$ SE | Sixth St Canal Site R 11 <br> Mean $\pm$ SE | $\begin{aligned} & \text { Hog Creek } \\ & \text { Site R } 12 \\ & \text { Mean } \pm \text { SE } \end{aligned}$ | Hog Creek Site R 13 <br> Mean $\pm$ SE | Selby Shoreline Site R 14 <br> Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Anchoa mitchilli | Bay anchovy | $0.4 \pm 0.4$ | $43.0 \pm 27.9$ | $3879.3 \pm 3669.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania parva | Rainwater killifish | $1.3 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.0 \pm 1.0$ | $41 \pm 41$ |
| Gambusia holbrooki | Eastern mosquito fish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $37.5 \pm 24.2$ | $0.0 \pm 0.0$ | $3492.3 \pm 1043.1$ |
| Leiostomus xanthurus | Spot | $857.1 \pm 857.1$ | $0.0 \pm 0.0$ | $415.7 \pm 251.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menidia spp. | Assorted silverside | $246.4 \pm 141.8$ | $6.8 \pm 6.3$ | $15.7 \pm 12.2$ | $0.7 \pm 0.7$ | $62.5 \pm 53.2$ | $0.0 \pm 0.0$ |
| Harengula jaguana | Scaled sardine | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Poecilia latipinna | Sailfin molly | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.6 \pm 5.6$ | $0.0 \pm 0.0$ | $2866.7 \pm 1832.1$ |
| Lagodon rhomboides | Pinfish | $77.7 \pm 18.3$ | $3.0 \pm 1.6$ | $153.6 \pm 64.9$ | $0.0 \pm 0.0$ | $17.7 \pm 13.9$ | $0.0 \pm 0.0$ |
| Eucinostomus spp. | Assorted mojarra | $1.8 \pm 1.1$ | $6.4 \pm 4.3$ | $18.2 \pm 9.2$ | $1.4 \pm 1.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cyprinodon variegatus | Sheepshead minnow | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus grandis | Gulf killifish | $22.9 \pm 22.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil cephalus | Striped mullet | $366.1 \pm 365.6$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $4.2 \pm 2.9$ | $0.0 \pm 0.0$ |
| Anchoa cubana | Cuban anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Floridichthys carpio | Goldspotted killifish | $18.2 \pm 16.8$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.6 \pm 2.6$ |
| Eucinostomus harengulus | Tidewater mojarra | $0.7 \pm 0.7$ | $24.6 \pm 23.9$ | $4.8 \pm 3.4$ | $11.1 \pm 9.4$ | $9.4 \pm 5.5$ | $0.0 \pm 0.0$ |
| Microgobius gulosus | Clown goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus majalis | Striped killifish | $28.6 \pm 28.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Clupeidae spp. | Assorted herring | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus gula | Silver jenny | $0.7 \pm 0.7$ | $20.0 \pm 11.0$ | $21.6 \pm 14.1$ | $2.1 \pm 2.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Anchoa hepsetus | Striped anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Brevoortia spp. | Assorted menhaden | $120.0 \pm 120.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.9 \pm 0.7$ | $0.2 \pm 0.2$ | $3.0 \pm 3.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura notata | Redfin needlefish | $0.5 \pm 0.3$ | $0.2 \pm 0.2$ | $1.6 \pm 1.6$ | $0.0 \pm 0.0$ | $2.1 \pm 2.1$ | $0.0 \pm 0.0$ |
| Syngnathus scovelli | Gulf pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.1 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphoeroides nephelus | Southern puffer | $0.4 \pm 0.2$ | $0.4 \pm 0.4$ | $1.6 \pm 1.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lophogobius cyprinoides | Crested goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Heterandria formosa | Least killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | Sarasota Baywalk Site R 9 Mean $\pm$ SE | $\begin{gathered} \hline \hline \text { Sixth St Canal } \\ \text { Site R 10 } \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Sixth St Canal <br> Site R 11 <br> Mean $\pm$ SE | Hog Creek <br> Site R 12 <br> Mean $\pm$ SE | Hog Creek Site R 13 Mean $\pm$ SE | $\begin{gathered} \hline \hline \text { Selby Shoreline } \\ \text { Site R } 14 \\ \text { Mean } \pm \text { SE } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Oligoplites saurus | Leatherjacket | $0.9 \pm 0.9$ | $2.0 \pm 1.1$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $12.5 \pm 11.2$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.0 \pm 1.0$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Synodus foetens | Inshore lizardfish | $0.7 \pm 0.4$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.8 \pm 1.8$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timucu | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chilomycterus schoepfi | Striped burrfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma robustum | Code goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diplodus holbrooki | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1 \pm 1$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.4 \pm 0.2$ | $0.4 \pm 0.3$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | $\begin{gathered} \hline \text { Sarasota Baywalk } \\ \text { Site R 9 } \\ \text { Mean } \pm \text { SE } \end{gathered}$ | Sixth St Canal Site R 10 Mean $\pm$ SE | Sixth St Canal Site R 11 Mean $\pm$ SE | Hog Creek Site R 12 Mean $\pm$ SE | $\begin{gathered} \hline \text { Hog Creek } \\ \text { Site R 13 } \\ \text { Mean } \pm \text { SE } \\ \hline \end{gathered}$ | Selby Shoreline Site R 14 <br> Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.0 \pm 1.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $1747.1 \pm 1457.4$ | $107.3 \pm 25.7$ | $4520.5 \pm 3465.9$ | $58.3 \pm 20.6$ | $112.5 \pm 75.9$ | $6402.6 \pm 2673.5$ |
| Richness |  | $9.8 \pm 0.9$ | $6.8 \pm 0.3$ | $8.5 \pm 3.1$ | $2.3 \pm 0.6$ | $3.8 \pm 1.4$ | $2.5 \pm 0.3$ |
| Diversity |  | $0.8 \pm 0.2$ | $0.9 \pm 0.0$ | $0.7 \pm 0.2$ | $0.5 \pm 0.2$ | $0.8 \pm 0.4$ | $0.6 \pm 0.1$ |

Table B-1. (Continued).
$\left.\begin{array}{llccccc}\hline \hline & & & & & \text { Selby Shoreline } & \text { Palmer Point } \\ \text { Sird Key }\end{array}\right)$

Table B-1. (Continued).

| Species name | Common name | Selby Shoreline <br> Site R 15 <br> Mean $\pm$ SE | Palmer Point Site R 16 <br> Mean $\pm$ SE | Bird Key Site R 17 <br> Mean $\pm$ SE | Hidden Bay Site R 18 <br> Mean $\pm$ SE | Total by species Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Oligoplites saurus | Leatherjacket | $0.7 \pm 0.7$ | $0.7 \pm 0.7$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.7 \pm 0.3$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.6 \pm 0.6$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.6 \pm 0.6$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.6 \pm 0.6$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $1.3 \pm 1.3$ | $0.0 \pm 0.0$ | $0.4 \pm 0.2$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.3 \pm 0.3$ |
| Synodus foetens | Inshore lizardfish | $0.4 \pm 0.2$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.3 \pm 0.2$ |
| Cynoscion nebulosus | Spotted seatrout | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.3 \pm 0.1$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ |
| Orthopristis chrysoptera | Pigfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $2.3 \pm 2.3$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Strongylura timuси | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ |
| Chilomycterus schoepfi | Striped burrfish | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.8 \pm 3.4$ | $0.2 \pm 0.1$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Gobiosoma robustum | Code goby | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.0$ |
| Diplodus holbrooki | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.9 \pm 0.9$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.2 \pm 0.2$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-1. (Continued).

| Species name | Common name | Selby Shoreline Site R 15 Mean $\pm$ SE | Palmer Point Site R 16 <br> Mean $\pm$ SE | Bird Key Site R 17 <br> Mean $\pm$ SE | Hidden Bay Site R 18 <br> Mean $\pm$ SE | Total by species Mean $\pm$ SE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $1.5 \pm 1.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $2083.7 \pm 1268.5$ | $976.1 \pm 279.3$ | $836.1 \pm 291.5$ | $76.9 \pm 31.7$ | $2224.1 \pm 645.9$ |
| Richness |  | $6.8 \pm 1.3$ | $11.5 \pm 2.5$ | $10.8 \pm 0.9$ | $3.8 \pm 0.3$ | $7.2 \pm 0.5$ |
| Diversity |  | $0.5 \pm 0.2$ | $1.3 \pm 0.3$ | $1.4 \pm 0.2$ | $1.0 \pm 0.1$ | $0.9 \pm 0.1$ |

Table B-2. Mean density (\#/100 $\mathrm{m}^{2}$ ) and standard error ( $\pm$ SE) by taxa pooled for each sampling location for all sampling events.

| Sampling Location Number of Samples |  | Tidy Island | Leffis Key | Sister Keys | El Conquistador | Bowlees Creek | South Lido |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 12 | 16 | 12 | 8 | 12 | 8 |
|  |  | $\underline{\text { Mean } \pm \text { SE }}$ | $\underline{\text { Mean } \pm \text { SE }}$ | $\underline{\text { Mean } \pm \text { SE }}$ | Mean $\pm$ SE | $\underline{\text { Mean } \pm \text { SE }}$ | $\underline{\text { Mean } \pm \text { SE }}$ |
| Species name | Common name |  |  |  |  |  |  |
| Anchoa mitchilli | Bay anchovy | $108.1 \pm 82.2$ | $1677.3 \pm 1148.7$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $5058.1 \pm 3056.5$ | $0.0 \pm 0.0$ |
| Lucania parva | Rainwater killifish | $271.7 \pm 151.9$ | $57.2 \pm 34.6$ | $1072.5 \pm 640.4$ | $162.6 \pm 91.3$ | $18.8 \pm 9.3$ | $1391.4 \pm 591.1$ |
| Gambusia holbrooki | Eastern mosquito fish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $46.2 \pm 34.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Leiostomus xanthurus | Spot | $131 \pm 51.3$ | $220.5 \pm 120.9$ | $82.5 \pm 37.8$ | $41 \pm 41$ | $719.7 \pm 687.9$ | $37.6 \pm 27.5$ |
| Menidia spp. | Assorted silverside | $50.1 \pm 29.9$ | $199.6 \pm 95.0$ | $123.0 \pm 95.9$ | $51.8 \pm 26.8$ | $181.8 \pm 111.4$ | $29 \pm 18.3$ |
| Harengula jaguana | Scaled sardine | $1.7 \pm 1.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1686.3 \pm 1677.7$ | $0.0 \pm 0.0$ |
| Poecilia latipinna | Sailfin molly | $0.0 \pm 0.0$ | $38.6 \pm 24.6$ | $10.8 \pm 7.2$ | $63.6 \pm 34$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lagodon rhomboides | Pinfish | $57.7 \pm 14.7$ | $489.3 \pm 20.00 .8$ | $106.3 \pm 45.2$ | $26.3 \pm 15.2$ | $103.8 \pm 60.6$ | $183.3 \pm 74.1$ |
| Eucinostomus spp. | Assorted mojarra | $509.8 \pm 194.3$ | $63.9 \pm 15.3$ | $10.4 \pm 7.8$ | $366.4 \pm 303.1$ | $77.6 \pm 35.9$ | $5.1 \pm 2$ |
| Cyprinodon variegatus | Sheepshead minnow | $13.3 \pm 9.0$ | $58.6 \pm 30.3$ | $78.3 \pm 54.2$ | $54.7 \pm 50.9$ | $39.6 \pm 31.6$ | $6.3 \pm 3.2$ |
| Mugil spp. | Assorted mullet | $67.6 \pm 49.7$ | $67.3 \pm 45.7$ | $188 \pm 171.4$ | $0.0 \pm 0.0$ | $116.7 \pm 94.8$ | $0.0 \pm 0.0$ |
| Fundulus grandis | Gulf killifish | $2.0 \pm 1.1$ | $12.4 \pm 7.8$ | $174.3 \pm 120.2$ | $0.0 \pm 0.0$ | $31.6 \pm 31.4$ | $2.9 \pm 2.2$ |
| Mugil cephalus | Striped mullet | $2.6 \pm 2.6$ | $0.1 \pm 0.1$ | $1.8 \pm 1.8$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $4.1 \pm 3.4$ |
| Anchoa cubana | Cuban anchovy | $6.7 \pm 6.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $178.4 \pm 121.3$ | $0.0 \pm 0.0$ |
| Floridichthys carpio | Goldspotted killifish | $42.9 \pm 15.6$ | $5.5 \pm 5.5$ | $23.8 \pm 10.3$ | $12.5 \pm 10.0$ | $18.7 \pm 15.5$ | $12.2 \pm 5.7$ |
| Eucinostomus harengulus | Tidewater mojarra | $26.5 \pm 26.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $33.2 \pm 21.9$ | $0.0 \pm 0.0$ |
| Microgobius gulosus | Clown goby | $3.8 \pm 2.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus majalis | Striped killifish | $40.0 \pm 28.8$ | $7.5 \pm 3.8$ | $4.4 \pm 2.9$ | $0.0 \pm 0.0$ | $2.0 \pm 2.0$ | $2.1 \pm 2.1$ |
| Clupeidae spp. | Assorted herring | $9.9 \pm 9.7$ | $49.3 \pm 36.8$ | $0.0 \pm 0.0$ | $0.8 \pm 0.8$ | $1.0 \pm 1.0$ | $0.0 \pm 0.0$ |
| Eucinostomus gula | Silver jenny | $11.9 \pm 5.2$ | $2.0 \pm 1.1$ | $0.0 \pm 0.0$ | $8.7 \pm 8.6$ | $1.8 \pm 1.2$ | $0.5 \pm 0.4$ |
| Anchoa hepsetus | Striped anchovy | $0.8 \pm 0.8$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $47.9 \pm 44.9$ | $0.0 \pm 0.0$ |
| Brevoortia spp. | Assorted menhaden | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.1 \pm 0.1$ | $12.9 \pm 5.7$ | $0.7 \pm 0.5$ | $0.5 \pm 0.5$ | $0.4 \pm 0.3$ | $0.2 \pm 0.2$ |
| Strongylura notata | Redfin needlefish | $1.1 \pm 1.0$ | $0.6 \pm 0.4$ | $2.5 \pm 2.3$ | $0.2 \pm 0.2$ | $1.2 \pm 1.0$ | $0.0 \pm 0.0$ |
| Syngnathus scovelli | Gulf pipefish | $1.9 \pm 1.5$ | $4.5 \pm 2.8$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.3 \pm 0.2$ | $0.0 \pm 0.0$ |
| Sphoeroides nephelus | Southern puffer | $2.7 \pm 1.6$ | $3.7 \pm 2.2$ | $0.2 \pm 0.1$ | $0.0 \pm 0.0$ | $1.5 \pm 1.3$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $0.5 \pm 0.5$ | $2.7 \pm 2.1$ | $0.2 \pm 0.1$ | $10.3 \pm 10.3$ | $0.1 \pm 0.1$ | $0.3 \pm 0.3$ |
| Lophogobius cyprinoides | Crested goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Heterandria formosa | Least killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $16.8 \pm 14$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-2. (Continued).

| Sampling Location Number of Samples |  | Tidy Island 12 | Leffis Key $16$ | $\begin{gathered} \hline \hline \text { Sister Keys } \\ 12 \end{gathered}$ | El Conquistador | Bowlees Creek 12 | South Lido $8$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE |
| Species name | Common name |  |  |  |  |  |  |
| Oligoplites saurus | Leatherjacket | $0.8 \pm 0.6$ | $1.1 \pm 0.6$ | $0.0 \pm 0.0$ | $0.6 \pm 0.5$ | $0.6 \pm 0.4$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.1 \pm 0.1$ | $6.0 \pm 5.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $8.3 \pm 8.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $12.0 \pm 8.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.5 \pm 0.5$ | $2.6 \pm 2.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Synodus foetens | Inshore lizardfish | $0.1 \pm 0.1$ | $2.2 \pm 2.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ |
| Cynoscion nebulosus | Spotted seatrout | $0.4 \pm 0.3$ | $0.7 \pm 0.5$ | $0.0 \pm 0.0$ | $0.6 \pm 0.5$ | $0.7 \pm 0.5$ | $0.1 \pm 0.1$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $1.0 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $2.2 \pm 1.2$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $2.1 \pm 2.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $1.8 \pm 1.8$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chilomycterus schoepfi | Striped burrfish | $1.4 \pm 1.4$ | $0.6 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timucu | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $2.0 \pm 2.0$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $3.3 \pm 2.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma robustum | Code goby | $1.0 \pm 1.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $1.0 \pm 1.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diplodus holbrooki | Bluntnose stingray | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.2 \pm 1.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hippocampus zosterae | Dwarf seahorse | $1.0 \pm 1.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.4 \pm 0.2$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-2. (Continued).

| Sampling Location Number of Samples |  | Tidy Island | Leffis Key | Sister Keys | El Conquistador | Bowlees Creek | South Lido |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 12 | 16 | 12 | 8 | 12 | 8 |
|  |  | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE |
| Species name | Common name |  |  |  |  |  |  |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Sphyraena barracuda | Great barracuda | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $1373.5 \pm 286.5$ | $2994 \pm 1121.8$ | $1889.7 \pm 614.2$ | $879.5 \pm 340.4$ | $8324.9 \pm 4971.3$ | $1676.2 \pm 589$ |
| Richness |  | $10.3 \pm 1$ | $11.3 \pm 1.2$ | $6 \pm 0.9$ | $6 \pm 1.2$ | $9.0 \pm 1.2$ | $6.0 \pm 0.9$ |
| Diversity |  | $1.2 \pm 0.1$ | $1.1 \pm 0.1$ | $0.9 \pm 0.2$ | $1.0 \pm 0.2$ | $0.8 \pm 0.1$ | $0.6 \pm 0.1$ |

Table B-2. (Continued).

| Sampling Location Number of Samples |  | Phillippi Creek | Heron Lagoon 8 | North Catfish 4 | Bird Keys <br> 8 | South Creek 8 | Durante Park 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE |
| Species name | Common name |  |  |  |  |  |  |
| Anchoa mitchilli | Bay anchovy | $6.7 \pm 6.6$ | $160.1 \pm 126.4$ | $7.3 \pm 7.1$ | $31.1 \pm 29$ | $118.3 \pm 107.6$ | $0.0 \pm 0.0$ |
| Lucania parva | Rainwater killifish | $5.7 \pm 2.7$ | $2933.8 \pm 1327.7$ | $0.4 \pm 0.2$ | $428.2 \pm 183.3$ | $0.0 \pm 0.0$ | $256.7 \pm 141.3$ |
| Gambusia holbrooki | Eastern mosquito fish | $1.9 \pm 1.5$ | $1430.9 \pm 1043.6$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $221.8 \pm 206.4$ |
| Leiostomus xanthurus | Spot | $0.5 \pm 0.5$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $13.8 \pm 9.1$ | $0.0 \pm 0.0$ | $416 \pm 330.6$ |
| Menidia spp. | Assorted silverside | $676.1 \pm 439.2$ | $505.5 \pm 312.5$ | $1.3 \pm 1.3$ | $70.0 \pm 41$ | $6.4 \pm 2.9$ | $261.7 \pm 168.5$ |
| Harengula jaguana | Scaled sardine | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $222.9 \pm 222.9$ | $10.7 \pm 7.5$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Poecilia latipinna | Sailfin molly | $2.8 \pm 1.7$ | $328.1 \pm 225.7$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $104.3 \pm 93.5$ |
| Lagodon rhomboides | Pinfish | $8.1 \pm 5.6$ | $0.0 \pm 0.0$ | $0.5 \pm 0.3$ | $72.1 \pm 29.5$ | $0.1 \pm 0.1$ | $126 \pm 74.9$ |
| Eucinostomus spp. | Assorted mojarra | $4.5 \pm 2.8$ | $0.2 \pm 0.2$ | $8.9 \pm 8.9$ | $27.1 \pm 20.3$ | $2.5 \pm 2.5$ | $275.9 \pm 152$ |
| Cyprinodon variegatus | Sheepshead minnow | $0.0 \pm 0.0$ | $1.5 \pm 1.5$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $142.7 \pm 112.4$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus grandis | Gulf killifish | $0.2 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $17.3 \pm 13.7$ | $0.0 \pm 0.0$ | $56.6 \pm 55.5$ |
| Mugil cephalus | Striped mullet | $0.6 \pm 0.5$ | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $72.5 \pm 61.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Anchoa cubana | Cuban anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Floridichthys carpio | Goldspotted killifish | $1.2 \pm 1.2$ | $17.6 \pm 14.6$ | $0.0 \pm 0.0$ | $35.7 \pm 20.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus harengulus | Tidewater mojarra | $16.5 \pm 11.6$ | $2.9 \pm 2.1$ | $0.2 \pm 0.2$ | $0.6 \pm 0.4$ | $0.5 \pm 0.5$ | $8.3 \pm 8.3$ |
| Microgobius gulosus | Clown goby | $16.7 \pm 12.8$ | $63.4 \pm 43.8$ | $3.0 \pm 1.2$ | $12 \pm 11.4$ | $21.1 \pm 6.9$ | $0.0 \pm 0.0$ |
| Fundulus majalis | Striped killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Clupeidae spp. | Assorted herring | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Eucinostomus gula | Silver jenny | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.8 \pm 1.4$ | $0.0 \pm 0.0$ | $0.5 \pm 0.4$ |
| Anchoa hepsetus | Striped anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Brevoortia spp. | Assorted menhaden | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.1 \pm 0.1$ | $10.1 \pm 9.9$ |
| Strongylura notata | Redfin needlefish | $0.0 \pm 0.0$ | $1.8 \pm 1.8$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Syngnathus scovelli | Gulf pipefish | $0.2 \pm 0.2$ | $4.2 \pm 4.2$ | $0.0 \pm 0.0$ | $0.9 \pm 0.5$ | $0.0 \pm 0.0$ | $1.1 \pm 0.8$ |
| Sphoeroides nephelus | Southern puffer | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.4 \pm 0.2$ |
| Callinectes sapidus | Blue crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.7 \pm 0.7$ | $0.4 \pm 0.4$ |
| Lophogobius cyprinoides | Crested goby | $12.5 \pm 10.2$ | $1.8 \pm 1.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.9 \pm 1.9$ | $0.0 \pm 0.0$ |
| Heterandria formosa | Least killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-2. (Continued).

| Sampling Location Number of Samples |  | Phillippi Creek | Heron Lagoon | North Catfish | Bird Keys | South Creek | Durante Park |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 8 | 8 | 4 | 8 | 8 | 12 |
|  |  | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | $\underline{\text { Mean } \pm \text { SE }}$ | Mean $\pm$ SE |
| Species name | Common name |  |  |  |  |  |  |
| Oligoplites saurus | Leatherjacket | $0.3 \pm 0.2$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.6 \pm 0.6$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ |
| Mugil curema | White mullet | $0.0 \pm 0.0$ | $2.9 \pm 2.9$ | $0.7 \pm 0.7$ | $1.2 \pm 1.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $6.1 \pm 3.8$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ |
| Synodus foetens | Inshore lizardfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.9 \pm 0.7$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.2$ | $0.0 \pm 0.0$ | $2.5 \pm 1.2$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chilomycterus schoepfi | Striped burrfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timucu | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ | $0.0 \pm 0.0$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.5 \pm 1.0$ |
| Gobiosoma robustum | Code goby | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.6 \pm 0.6$ | $0.9 \pm 0.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diplodus holbrooki | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.6 \pm 0.4$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.1 \pm 0.8$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-2. (Continued).

| Sampling Location Number of Samples |  | ${ }_{8}^{\text {Phillippi }}$ Creek | Heron Lagoon | North Catfish <br> 4 | Bird Keys | South Creek <br> 8 | Durante Park 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE |
| Species name | Common name |  |  |  |  |  |  |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Spottail pinfish | $0.2 \pm 0.2$ | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.3 \pm 0.2$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $755.3 \pm 472.5$ | $5462.3 \pm 2326.4$ | $248 \pm 226$ | $797.6 \pm 232.7$ | $157.6 \pm 114.8$ | $1885.9 \pm 358.9$ |
| Richness |  | $7.3 \pm 0.5$ | $5.5 \pm 0.9$ | $6.0 \pm 1.9$ | $7.9 \pm 1.3$ | $4.8 \pm 0.8$ | $5.4 \pm 1$ |
| Diversity |  | $1.1 \pm 0.2$ | $0.7 \pm 0.1$ | $1.0 \pm 0.5$ | $0.9 \pm 0.2$ | $0.6 \pm 0.1$ | $0.7 \pm 0.1$ |

Table B-2. (Continued).

| Sampling Location Number of Samples |  | Sarasota Baywalk 12 | Sixth Street | $\begin{gathered} \hline \hline \text { Hog Creek } \\ 8 \end{gathered}$ | Selby Shoreline 8 | Palmer Point 4 | $\begin{gathered} \hline \hline \text { Hidden Bay } \\ 4 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE |
| Anchoa mitchilli | Bay anchovy | $0.3 \pm 0.2$ | $1961.2 \pm 1847.1$ | $0.0 \pm 0.0$ | $845.7 \pm 718.5$ | $121.7 \pm 51.0$ | $0.0 \pm 0.0$ |
| Lucania parva | Rainwater killifish | $1.3 \pm 0.7$ | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $21.6 \pm 20.4$ | $46.3 \pm 42.5$ | $0.0 \pm 0.0$ |
| Gambusia holbrooki | Eastern mosquito fish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $18.7 \pm 13.2$ | $1746.2 \pm 817.8$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Leiostomus xanthurus | Spot | $285.8 \pm 285.7$ | $207.9 \pm 140.5$ | $0.0 \pm 0.0$ | $0.3 \pm 0.3$ | $158.1 \pm 151.3$ | $0.6 \pm 0.6$ |
| Menidia spp. | Assorted silverside | $99.6 \pm 53.6$ | $11.3 \pm 6.6$ | $31.6 \pm 27.3$ | $80.3 \pm 64.5$ | $391.5 \pm 147.2$ | $26.3 \pm 20.5$ |
| Harengula jaguana | Scaled sardine | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.9 \pm 2.9$ | $0.0 \pm 0.0$ |
| Poecilia latipinna | Sailfin molly | $53.8 \pm 28.6$ | $0.0 \pm 0.0$ | $2.8 \pm 2.8$ | $1433.3 \pm 1006.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lagodon rhomboides | Pinfish | $25.9 \pm 12.3$ | $78.3 \pm 41.4$ | $8.9 \pm 7.2$ | $102.2 \pm 63.2$ | $9.9 \pm 5.1$ | $0.0 \pm 0.0$ |
| Eucinostomus spp. | Assorted mojarra | $2.4 \pm 1.5$ | $12.3 \pm 5.2$ | $0.7 \pm 0.7$ | $7.1 \pm 5.7$ | $50.4 \pm 32.8$ | $30.8 \pm 18.3$ |
| Cyprinodon variegatus | Sheepshead minnow | $98.3 \pm 62.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $3.7 \pm 2.8$ | $0.0 \pm 0.0$ |
| Mugil spp. | Assorted mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus grandis | Gulf killifish | $14.9 \pm 8.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $2.6 \pm 2.6$ | $3.2 \pm 2.4$ |
| Mugil cephalus | Striped mullet | $138.5 \pm 120.7$ | $0.1 \pm 0.1$ | $2.1 \pm 1.6$ | $0.0 \pm 0.0$ | $91.2 \pm 91.2$ | $0.0 \pm 0.0$ |
| Anchoa cubana | Cuban anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Floridichthys carpio | Goldspotted killifish | $9.7 \pm 5.8$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $1.3 \pm 1.3$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus harengulus | Tidewater mojarra | $17.4 \pm 15.7$ | $14.7 \pm 11.8$ | $10.2 \pm 5$ | $0.0 \pm 0.0$ | $59.9 \pm 34.2$ | $7.7 \pm 6.1$ |
| Microgobius gulosus | Clown goby | $4.6 \pm 4.4$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $12.1 \pm 11.6$ | $0.0 \pm 0.0$ |
| Fundulus majalis | Striped killifish | $27.9 \pm 12.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Clupeidae spp. | Assorted herring | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Eucinostomus gula | Silver jenny | $15.6 \pm 15.1$ | $20.8 \pm 8.3$ | $1.0 \pm 1.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.9 \pm 1.2$ |
| Anchoa hepsetus | Striped anchovy | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Brevoortia spp. | Assorted menhaden | $40.1 \pm 40.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Farfantepenaeus duorarum | Pink shrimp | $0.3 \pm 0.2$ | $1.6 \pm 1.5$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Strongylura notata | Redfin needlefish | $0.2 \pm 0.1$ | $0.9 \pm 0.8$ | $1.0 \pm 1.0$ | $0.3 \pm 0.3$ | $18.0 \pm 11.3$ | $0.0 \pm 0.0$ |
| Syngnathus scovelli | Gulf pipefish | $0.0 \pm 0.0$ | $0.5 \pm 0.4$ | $0.0 \pm 0.0$ | $0.3 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphoeroides nephelus | Southern puffer | $0.1 \pm 0.1$ | $1.0 \pm 0.7$ | $0.0 \pm 0.0$ | $3.7 \pm 3.6$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Callinectes sapidus | Blue crab | $0.1 \pm 0.1$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.6 \pm 0.6$ |
| Lophogobius cyprinoides | Crested goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Heterandria formosa | Least killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-2. (Continued).

| Sampling Location Number of Samples |  | Sarasota Baywalk | Sixth Street | Hog Creek | Selby Shoreline | Palmer Point | Hidden Bay |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 12 | 8 | 8 | 8 | 4 | 4 |
|  |  | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE |
| Species name | Common name |  |  |  |  |  |  |
| Oligoplites saurus | Leatherjacket | $0.3 \pm 0.3$ | $1.1 \pm 0.6$ | $6.2 \pm 5.7$ | $0.4 \pm 0.4$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ |
| Scomberomorus maculatus | Spanish mackerel | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Adinia xenica | Diamond killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lucania goodei | Bluefin killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sciaenops ocellatus | Red drum | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Mugil curema | White mullet | $1.6 \pm 1.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Opsanus beta | Gulf toadfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Synodus foetens | Inshore lizardfish | $0.2 \pm 0.2$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.2 \pm 0.1$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Cynoscion nebulosus | Spotted seatrout | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma spp. | Assorted goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Orthopristis chrysoptera | Pigfish | $0.0 \pm 0.0$ | $0.9 \pm 0.9$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Urophycis floridana | Southern hake | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bairdiella chrysoura | Silver perch | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chilomycterus schoepfi | Striped burrfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura timuси | Timucu | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Micropterus salmoides | Largemouth bass | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Centropomus undecimalis | Snook | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $5.8 \pm 3.4$ |
| Megalops atlanticus | Tarpon | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Gobiosoma robustum | Code goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.7 \pm 0.7$ | $0.0 \pm 0.0$ |
| Syngnathus floridae | Dusky pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Strongylura marina | Atlantic needlefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Diplodus holbrooki | Bluntnose stingray | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus spp. | Assorted killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hippocampus zosterae | Dwarf seahorse | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Achirus lineatus | Lined sole | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Lutjanus griseus | Gray snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $0.1 \pm 0.1$ | $0.4 \pm 0.4$ | $0.0 \pm 0.0$ |
| Gobiosoma bosc | Naked goby | $0.2 \pm 0.2$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Bathygobius soporator | Frillfin goby | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Pogonias cromis | Black drum | $0.7 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Archosargus probatocephalus | Sheepshead | $0.2 \pm 0.1$ | $0.3 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Mycteroperca microlepis | Gag | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |

Table B-2. (Continued).

| Sampling Location Number of Samples |  | Sarasota Baywalk 12 | Sixth Street 8 | $\begin{gathered} \hline \hline \text { Hog Creek } \\ 8 \end{gathered}$ | Selby Shoreline 8 | Palmer Point 4 | $\begin{gathered} \text { Hidden Bay } \\ 4 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE | Mean $\pm$ SE |
| Species name | Common name |  |  |  |  |  |  |
| Mugil gyrans | Fantail mullet | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.5 \pm 1.0$ | $0.0 \pm 0.0$ |
| Strongylura spp. | Assorted needlefish | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $1.1 \pm 1.1$ | $0.0 \pm 0.0$ |
| Diapterus plumieri | Spottail pinfish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Calamus arctifrons | Grass porgy | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.5 \pm 0.5$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lepomis spp. | Bluegill | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Syngnathus louisianae | Chain pipefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Trinectes maculatus | Hogchoker | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Fundulus confluentus | Marsh killifish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Limulus polyphemus | Horseshoe crab | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Chaetodipterus faber | Atlantic spadefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Aluterus schoepfi | Orange filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Hyporhamphus spp. | Halfbeak (juv) | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Monacanthus hispidus | Planehead filefish | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Sphyraena barracuda | Great barracuda | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Dasyatis say | Striped mojarra | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Haemulon parrai | Sailors choice | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Lutjanus synagris | Lane snapper | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Menticirrhus saxatilis | Northern kingfish | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Paralichthys albigutta | Gulf flounder | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Selene vomer | Lookdown | $0.0 \pm 0.0$ | $0.1 \pm 0.1$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ | $0.0 \pm 0.0$ |
| Total Mean Density |  | $840.7 \pm 488.9$ | $2313.9 \pm 180.08 .2$ | $85.4 \pm 37.8$ | $4243.2 \pm 1594.6$ | $976.1 \pm 279.3$ | $76.9 \pm 31.7$ |
| Richness |  | $8.2 \pm 1.1$ | $7.6 \pm 1.5$ | $3.0 \pm 0.8$ | $4.6 \pm 1$ | $11.5 \pm 2.5$ | $3.8 \pm 0.3$ |
| Diversity |  | $1.1 \pm 0.2$ | $0.8 \pm 0.1$ | $0.6 \pm 0.2$ | $0.5 \pm 0.1$ | $1.3 \pm 0.3$ | $1.0 \pm 0.1$ |


[^0]:    ${ }^{1} \mathrm{CB}$ - Concrete block pile; CB, RB (3) - three reef balls on a concrete block pile; PVC RC - PVC "rib cage," PVC Web - PVC web design; RB - reef ball(s)

[^1]:    SBNEP Juvenile Fisheries Habitat Assessment

[^2]:    ${ }^{1}$ Selected taxa represent commercially and/or recreationally important species.

[^3]:    STD Error: Standard Error of Means

